



Beyond Cleansing: Ecosystem Services Related to Phytoremediation

Werther Guidi Nissim ^{1,4,*}, Stefano Castiglione ^{2,4}, Francesco Guarino ^{2,4}, Maria Chiara Pastore ^{3,4} and Massimo Labra ^{1,4}

- ¹ Department of Biotechnology and Biosciences, University of Milano Bicocca, Piazza della Scienza 2, 20126 Milan, Italy
- Department of Chemistry and Biology "A. Zambelli", University of Salerno, Via G. Paolo II no 132, 84084 Fisciano, Italy
- ³ Politecnico di Milano, Department of Architecture and Urban Studies, Via Bonardi 3, 20133 Milano, Italy
- ⁴ National Biodiversity Future Center (NBFC), 90133 Palermo, Italy
- * Correspondence: werther.guidinissim@unimib.it; Tel.: +39-02-6448-3465

Abstract: Phytotechnologies used for cleaning up urban and suburban polluted soils (i.e., brownfields) have shown some weakness in the excessive extent of the timeframe required for them to be effectively operating. This bottleneck is due to technical constraints, mainly related to both the nature of the pollutant itself (e.g., low bio-availability, high recalcitrance, etc.) and the plant (e.g., low pollution tolerance, low pollutant uptake rates, etc.). Despite the great efforts made in the last few decades to overcome these limitations, the technology is in many cases barely competitive compared with conventional remediation techniques. Here, we propose a new outlook on phytoremediation, where the main goal of decontaminating should be re-evaluated, considering additional ecosystem services (ESs) related to the establishment of a new vegetation cover on the site. The aim of this review is to raise awareness and stress the knowledge gap on the importance of ES associated with this technique, which can make phytoremediation a valuable tool to boost an actual green transition process in planning urban green spaces, thereby offering improved resilience to global climate change and a higher quality of life in cities. This review highlights that the reclamation of urban brownfields through phytoremediation may provide several regulating (i.e., urban hydrology, heat mitigation, noise reduction, biodiversity, and CO₂ sequestration), provisional (i.e., bioenergy and added-value chemicals), and cultural (i.e., aesthetic, social cohesion, and health) ESs. Although future research should specifically be addressed to better support these findings, acknowledging ES is crucial for an exhaustive evaluation of phytoremediation as a sustainable and resilient technology.

Keywords: phytotechnologies; phytoremediation; ecosystem services; nature-based solution; green transition

1. Introduction

The use of green living plants to address pollution is a growing option for environmental management of polluted sites, for it encompasses both environmental (solar-driven technology) and economical (cheaper than most conventional technologies) features, which are both required to meet sustainability and resilience goals for modern societies [1]. This phytotechnology shows high potential for many sites, especially in emerging countries, where it could represent a rather inexpensive reclamation technique compared to conventional, expensive, engineered technologies [2]. However, the technology shows some weaknesses that still hinder the application on a large scale. Although constraints are specific to each pollutant, low plant tolerance to environmental toxicants, long life span to be effective, and pollutant availability for the plant roots are the most common for different phytoremediation approaches [3]. Despite advances in the understanding of the mechanisms regulating the relationship between plants and pollutants and the subsequent new



Citation: Guidi Nissim, W.; Castiglione, S.; Guarino, F.; Pastore, M.C.; Labra, M. Beyond Cleansing: Ecosystem Services Related to Phytoremediation. *Plants* **2023**, *12*, 1031. https://doi.org/10.3390/ plants12051031

Academic Editor: Michela Schiavon

Received: 20 January 2023 Revised: 20 February 2023 Accepted: 22 February 2023 Published: 24 February 2023



Copyright: © 2023 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). strategies to enhance the whole reclamation process, the improvements in the functionality of the technology at a large, field scale are slow [4]. However, the use of a green approach based on living plants should imply that phytoremediation will also provide additional environmental benefits, which must be assessed in the future. The current review aims to highlight the opportunities that the phytoremediation of urban brownfields offers in terms of ecosystem services (ESs). The goal of this review is to provide a critical outlook on the potential production of ESs during brownfield reclamation when a phyto-technological approach is used.

2. Lights and Shadows of Phytomanagement of Brownfields

2.1. The Dual Identity of Brownfields: Challenges and Opportunities

Over the last several decades, cities in many parts of the world have been subjected to a dramatic rise in population. Currently, about 55% of the world's population lives in urban areas, and this proportion is expected to increase to 68% by 2050 [5]. This continuous increase in urbanization has induced many economic activities (i.e., industries) to shift from urban to suburban areas to leave space for new settlements. The displacement of industries from the city centre to peri urban areas often leaves the inner core typically with innumerable underutilized or vacant industrial sites. This has resulted in numerous sites that remain derelict or underused due to land-use restrictions based on concerns related to contamination by hazardous substances [6]. Although there is still a very active debate about how to define these sites [7], these areas are commonly referred to as brownfields [8].

Although brownfields are ubiquitous, their precise extent is not easy to quantify on a global scale, mainly due to disagreement about their definitions [9]. A survey carried out in 2001 in several EU countries highlighted that the extent of these lands varied greatly among countries (e.g., 128,000 ha in Germany; 39,000 in the United Kingdom; 20,000 in France; 13,000 in Italy; and 10,000 in the Netherland) [10]. In the USA, more than 500,000 brownfield sites still need to be redeveloped. Most brownfields are concentrated in an urban context. For instance, in the USA, 5–10% of the urban land is classified as brownfield, and the percentage in cities of the Northeast and Midwest (Rustbelt) states is even higher [11]. Brownfield contamination is generally located in both the soil and the groundwater and is due to either organic or inorganic toxicants, or in most cases, both [12]. Among inorganic contaminants, trace elements (i.e., Al, B, Ca, Cd, Co, Cr, Cu, Fe, K, Mg, Mn, Na, P, Pb, Si, Ti, V, and Zn) are very common, along with asbestos [13], whereas the main organic compounds are represented by polycyclic aromatic hydrocarbons (PAHs), polychlorinated biphenyls (PCBs), and petroleum hydrocarbons (PHCs) [14]. At present, the elevated level of pollution in urban brownfields can be attributed to several cooccurring factors, including (i) transportation, (ii) commercial and industrial emissions, and (iii) domestic activities [15].

In many countries, the redevelopment of urban brownfields is considered socially, economically, environmentally, and culturally important for city planning and a valuable tool to counteract urban sprawl [16]. However, brownfields oftentimes pose health risks for inhabitants. Some studies have reported brownfield exposure to be correlated with regional inequalities in mortality and morbidity in different regions within the UK [17]. Specifically, living close to brownfield sites may result in significantly lower naive T-cell production, suggesting accelerated immune system aging for people living near these sites [18]. Current strategies for brownfield reclamation include saturated zone and vadose zone technologies [19]. While the former approach is recommended when contamination affects both soil and groundwater, vadose technologies are more adapted for polluted groundwater. Pump-and-treat, reactive walls, air sparging, dual phase extraction, flushing, bioremediation, electrokinetic, and immobilization are the most common saturated zone technologies available on the market. Soil excavation, followed by either landfill disposal or treatment (i.e., soil washing, solvent extraction, electrokinetic, thermal desorption, incineration, vitrification, or bioremediation), is the most popular ex situ vadose approach [20]. Unfortunately, this very efficient approach shows many weaknesses related to its high operational costs and low environmental sustainability. From this perspective, on-site

remediation techniques seem a more attractive solution. In situ vadose techniques include several highly engineered approaches, some of which (e.g., soil vapor extraction, soil flushing, electrokinetic, soil heating, vitrification, and solidification) often rely on non-renewable energy sources [21] with high environmental footprints [22]. For instance, it has been shown that 2.7 million tons of CO₂ were produced using a "dig and haul" approach to remediate a single brownfield in New Jersey (USA), which is equal to 2% of the annual CO₂ emissions for the entire state [23]. By contrast, phytoremediation approaches show high potential for their small environmental footprints, low operational costs, and high social acceptance [24].

2.2. From Bench to Field: Success and Challenges of Brownfield Phytoremediation

Since the beginning, phytoremediation received positive feedback, provided by the first studies showing its high potential for specific pollutants to be either taken up and accumulated or degraded by specific plants [25] (Table 1). However, like many other environmental technologies, the scaling up of phytoremediation from lab and mesocosm to actual field conditions has often resulted in different outputs, ranging from complete success to almost complete failure, thereby attenuating the initial enthusiasm. This variability in the performance response at field scale is due to co-occurring factors, including the type, status (e.g., bioavailability for trace elements) [26], and concentrations of toxicant(s) [27]; soil chemical and physical properties (pH, conductivity, texture, porosity, nutrient levels, and presence of soil microorganisms) [28]; plant species; rate of plant growth; and climatic conditions [29].

Mechanisms of Phytoremediation	Description	Contaminant Type Addressed	Plant Species	Reference
Phytoextraction	Plants uptake pollutants via their roots and accumulataltion in aerial biomass whose harvest allows progressive removal from the soil	Inorganic pollutants (As, Cd, Cr, Cu, Ni, Pb, Se, Zn)	Hyperaccumulators (Noccaea caerulescens, Brassica juncea, Alyssum spp., Arabidopsis helleri, Pteris vittata, Sedum plumbizincicola, Arabidopsis thaliana) Fast-growing trees (Populus spp., Salix spp. Eucalyptus spp.)	[30–36]
Phytostabilization	Plants produce specific metabolites which reduce the solubility and mobility of contaminants within the rhizosphere	Inorganic pollutants (Al, Co, Cu, Cr, Fe, Mn, Mo, Pb)	Acanthus ilicifolius, Agrostis capillaris, Arundo donax, Atriplex halimus, Brassica juncea, Populus deltoides, Jatropha curcas, Lolium perenne, Miscanthus sinensis x giganteus, Pteridium aquilinum, Ricinus communis, Salix purpurea	[37–41]
Phytodegradation	Plants, frequently assisted by microorganisms, take up and transform contaminants into less harmful compounds	Organic pollutants (petroleum hydrocarbons, polycyclic aromatic hydrocarbons, pesticides)	Tagetes patula, Aster amellus, Portulaca grandiflora, Aster amellus, Iris dichotoma, Gaillardia aristata, Echinacea purpurea, Festuca arundinacea, Medicago sativa, Cytisus striatus, Nerium oleander, Ricinus communis, Populus spp., Salix spp.	[42–48]
Phytovolatilization	Plants take up the pollutant and transpire it to the atmosphere as a gas, hence removing it from the site	Inorganic pollutants (As, Hg, Se) Organic pollutants (trichloroethylene, tetrachloroethylene, MTBE)	Pteris vittata, Arundo donax, Dittrichia viscosa, Oryza sativa, Zea mays, Brassica juncea	[49–53]
Rhizodegradation	Soil contaminants are broken down by external plant processes mediated by microbial activity	Organic pollutants	Vigna unguiculata, Helianthus annuus, Zea mays, Sorghum sudanense	[54–59]

Table 1. The most common phytoremediation approaches.

One of the most successful full-scale phytoremediation approaches in the field is a particular type of degradation technique known as phytometabolism. In this case, the environmental toxicants are also plant nutrients (e.g., inorganic elements, such as N and P), which, for this reason, are directly metabolized and incorporated into the plant's biomass. This is the case of vegetation filter systems for the treatment of municipal wastewater. A recent review reported that in global terms, vegetation filters, which are mainly (70%) constituted by tree species belonging to the Salicaceae family, show average removal rates of

about 78% for N and 80% for P [60], making this technology a suitable green tool especially suited for scattered populations or isolated buildings lacking connection to sewer systems. On the opposite side are the techniques based on the extraction of some inorganic pollutants that, in some cases, might take a considerable amount of time to be removed offsite.

The removal from soil of trace elements (both metals and metalloids) by plants is one clear example of phytotechnology that shows both high potential and interest and huge operational constraints. This approach has been extensively used at full field scale in different regions. The early field applications reported promising results for *Brassica juncea* combined with soil amendments on a lead-polluted brownfield [61] and for *Buchloe dactyloides* for naphthalene phytoremediation [62]. However, in this period, scientists still claimed that there was a substantial need for more demonstration projects to prove the efficiency of green technologies in the field from a long-term perspective [63]. These circumstances have led to the development of actual rehabilitation projects at field scale. Most initiatives have assessed the potential of several approaches for the cleansing of metal-contaminated brownfields, including phytoextraction by different species [64–66] and phytostabilization [67,68]. For example, the GREENLAND project funded by the European Commission (FP7) established a large-scale, field demonstration network, where new approaches and financial aspects related to phytoremediation were investigated [69].

Another field-scale project was developed in Rozelle (Australia), where *Brassica juncea* provided positive results, taking up significant amounts of lead from contaminated soil [70]. Promising results were also obtained in the "Opération Tournesol", a Belgian-led rehabilitation project in Brussels, where *Noccaea caerulescens* showed positive results for cadmium and zinc phytoextraction [71]. Some woody species (i.e., willow and poplar) have also shown a notable capacity to remediate soil contaminated by trace elements on a former brownfield site in Detroit (USA), but additional time is required for validation [72].

Hence, despite intensive research over the past few years, trace element pollution continues to be challenging, and to date, there is still not a perfect phytotechnology for cleaning up and restoring soils within a reasonable timeframe. Different models have predicted that under ideal conditions, plants with an average dry biomass yield of $10 \text{ Mg ha}^{-1} \text{ yr}^{-1}$ should show bioaccumulation coefficients higher than 7.4 to reduce the total metal concentration by 50% in 25 years of phytoextraction [73], a condition which rarely occurs in nature, even for hyperaccumulator plants [74]. Time-consuming phytoextraction has been addressed by several strategies aimed at intensifying the efficiency of phytoremediation, thereby reducing the overall duration of the process. These include synergistic growth of plants and plant growth-promoting microorganisms [75], the use of chelating agents and soil amendments [76], co-planting different species [77], and the use of transgenic plants [26]. Despite these efforts, phytoextraction yields are expected to be rather low and the time for operating still too long to compete with conventional remediation techniques, making its application at field scale less attractive. In addition, during the recent decade, some concerns have been raised about the biomass issued from the process, which in some cases may contain significant amounts of toxicants, thereby representing a waste to be managed [78]. However, new promising strategies, where polluted biomass is used as feedstock or thermo- and biochemical compounds converted into biofuels, are now under assessment [79].

3. Ecosystem Services: The Known and Unknown Aspects Related to Phytoremediation of Brownfields

The most popular definition of ecosystem services (ESs) is a hybrid ecological–economic approach, directly linking an ecosystem's functions and processes and the benefits derived for humans [80]. The most common ESs are regulating, provisioning, supporting, and cultural services. Regulating services are the benefits provided by ecosystems that moderate natural phenomena. These are the benefits obtained from the regulation of ecosystem processes, and they include flood protection, climate regulation, water purification, air quality maintenance, and biodiversity, all of which contribute to human well-being in

cities [81]. Provisioning services are those related to the production of goods from any natural process. Cultural ESs are non-material benefits obtained from ecosystems (i.e., cultural diversity, spiritual and religious values, knowledge systems, educational values, inspiration, aesthetic values, social relations, sense of place, cultural heritage values, recreation, and ecotourism) that people may take advantage of [82]. The assessment of ESs in brownfield redevelopment has been assessed in many contexts, but most studies refer to green urban brownfields. These spaces are generated when an unsealed brownfield undergoes natural processes of ecological succession, thereby leading to a particular type of urban vegetation [83]. These brownfields are thought to have the potential to provide a wide range of ESs and that the differences in their extent depends on the specific structure and composition of the vegetation cover [84]. Similar conclusions have been drawn for soft brownfield re-use approaches, where the new established green ecosystems can provide multiple ESs to improve the urban environment, citizen health, and quality of life [85]. In some cases, ESs generated by informal unmanaged green spaces are even higher than those generated by the establishment of urban parks [86]. Despite the relatively vast body of scientific information about ES generation during brownfield recovery, less is known about the role that phytoremediation could play in this context. In fact, many differences may occur when greening a brownfield using a more engineering-oriented approach. Most phytoremediation approaches use few plant species to target specific soil contaminants, and although some attempts have been done to enhance plant diversity by co-cropping different species [87], it is unlikely that floristic diversity during phytoremediation would be higher than a vegetation cover naturally established [88], where the distinctive spatial-temporal dynamics of urban brownfields induce a relatively high species diversity [89]. Another difference is that unlike for most soft brownfield re-use approaches, where the site can be entered, phytoremediation sites are normally inaccessible and, thus, several ESs related to the use of urban green spaces may be attenuated. Despite the long-term research on phytoremediation, few specific studies have been dedicated to quantifying ESs related to this green technology [90], whereas most research has focused on afforestation/reforestation with non-specific phytoremediation approaches [91-93]. The main potential ESs related to brownfield remediation are reported in Figure 1.

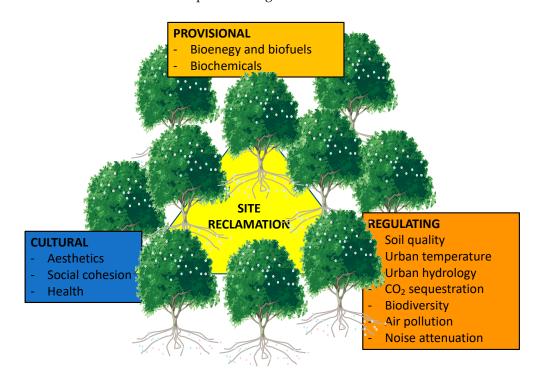


Figure 1. The main components of ecosystem services related to phytomanagement.

3.1. Regulating Ecosystem Services

The phytoremediation of urban brownfields could add important benefits in terms of regulating environmental parameters, including soil, temperature, hydrology, biodiversity, noise attenuation, and carbon sequestration.

3.1.1. Soil Quality

Using living plants for the reclamation of urban brownfield can have a profound impact on soil quality and regeneration. Urban brownfields are formed as a result of anthropogenic factors (pollution, compaction, and loss of fertility) and natural factors of soil formation [94]. Unlike most conventional remediation techniques, which are associated with high soil disturbance [95], plants involved in phytoremediation have been shown to reduce soil disturbance, thus enhancing the carbon storage in soil as organic matter [96]. Different soil properties have been shown to improve following plant establishment on polluted sites, including chemical [97], physical [98], and biological [99] characteristics. The use of arbuscular mycorrhizal fungi in phytoremediation has been found to increase stress tolerance [100] and the heavy metal bioavailability for the plants, supporting their establishment on harsh sites [101]. Enhanced soil physicochemical properties were observed when phytoremediation was assisted by soil amendments. For instance, compost has been shown to improve soil physicochemical properties, including the increment in soil organic matter and nutrient content, at a Cu-contaminated site phyto-managed with different herbaceous species [102]. As in the case of compost, adding biochar to contaminated soils has been shown to increase soil pH, water-holding capacity, and soil fertility; reduce the mobility of plant-available pollutants; and promote vegetation establishment [103]. The establishment of a new vegetation cover on a brownfield also can reduce soil erosion by wind and leaching of soil-contaminating elements to groundwater [104]. Plants can create a physical barrier that holds soil particles in place. This action can be aided by root exudates, which reduce, by precipitation, the mobility of specific contaminants in the environment [105]. The benefits of phytoremediation related to soil quality are somewhat counterbalanced by some risks. The most common threat is represented by the application of mobilizing agents, which solubilize toxic soil contaminants that, when not promptly taken up by the plants, can be spread in the environment [106]. Though these uncertainties need further research, most phytoremediation approaches represent a valuable tool to enhance the overall soil fertility, thereby strengthening its ecological resilience to further disturbances.

3.1.2. Urban Temperature

The establishment of a plant cover in an urban unvegetated area may lead to a significant mitigation in urban temperature, in particular by attenuating the urban heat island (UHI) phenomenon, i.e., severe temperature increases of several degrees (sometimes over 10 °C) compared to the surrounding areas [107]. In most cases, this action is achieved through direct shading, enhanced evapotranspiration, and thermal and optical properties specific to plants [108]. This action shows positive effects also in energy savings following climatization in city buildings [109]. Additionally, a vegetation cover in winter (especially evergreen trees) can represent a valuable windbreak protecting buildings from cold winds, therefore reducing energy consumption for heating [110]. It is well established that the cooling effect of vegetation depends on a combination of multiple factors, including structural (leaf colour and canopy structure) and functional (ecophysiological adaptations) traits of plants [111], environmental factors, and the size of the vegetated area [112]. Arboreal and herbaceous vegetation contribute differently to temperature mitigation. Trees appear to be more efficient in reducing high diurnal temperatures through the shading effect of their crown, whereas their canopy can retain heat at night by decreasing the movement of warm air and, thus, reducing emissions of long-wave radiation [111]. Another study has shown that woody vegetation (trees and shrubs) is able to reduce daily soil-surface temperatures in the summer by 5.7 °C compared to herbaceous vegetation and tends to maintain slightly higher temperatures in winter [113]. Other studies have pointed out that although both

vegetation types are able to reduce the UHI in hot weather, grass showed a lower impact on local air temperatures and on human comfort, whereas trees provided effective and substantial local cooling [114]. Despite the vast body of literature on the thermal effects of urban vegetation, no reported information, however, is currently available on whether and to what extent phytoremediation can contribute to mitigating urban temperature, and therefore, any effect on temperature is not easy to predict. Most herbaceous hyperaccumulator species commonly used in phytoremediation are likely to respond as normal herbaceous vegetation. On the other hand, fast-growing woody species in short rotation are not very predictable. These stands, which are harvested over very short intervals (1-3 years), result in a very dense shrubby structure (up to 30,000 plants ha^{-1}) and show very high evapotranspiration rates [115]. This will likely induce cooler air temperatures around the site, but the high density of the canopy structure could also decrease dissipation through soil irradiation at night. In addition, the high evapotranspiration rates of these species could be severely reduced by some sort of stress response to a high level of soil contamination in some spot of the brownfield [116]. Therefore, further research work is needed to clarify the effect of these specific plant stands on temperature regulation at the urban scale.

3.1.3. Urban Hydrology

Flooding represents one of the main hazards in modern towns and cities. The increased number of flood events per year occurring in urban areas is mainly due to two factors: a higher frequency of extreme rainfall under current global climate change events [117] and human-induced alterations in land cover [118]. Soil sealing during urban sprawl is the main anthropogenic activity that reduces water infiltration and causes stronger surface run-off and flooding [119]. Green urban area expansion represents, among others, a valid technical solution to be implemented in order to attenuate this phenomenon [120]. The reduced risk of flooding displayed by vegetation is due to its ability to intercept, retain, and infiltrate rainwater. Plants capture and evaporate large amounts of rainwater directly via their surface tissues, and intercept and transfer large amounts of water from the soil to the atmosphere via transpiration [121]. In addition, due to low surface-soil bulk density, vegetation contributes to the infiltration of stormwater, resulting in a reduction of flood frequency and severity [122]. Although several herbaceous and woody species have been selected for their capacity to develop large root systems appropriate for flooded sites, herbaceous vegetation is in general more suited to mitigating stormwater runoff via soil infiltration, whereas trees are more efficient, due to canopy interception, in reducing the amount of water reaching the ground [114]. In fact, though the degree of attenuation of stormwater runoff depends on different characteristics, including species, canopy density, plant size, bark structure, canopy storage capacity, planting density, and the presence/absence of foliage [123], the best-suited species to attenuate this risk are perennials showing high evapotranspiration rates and Leaf Area Index and elevated canopy density. It is noteworthy that most perennial woody species used for brownfield reclamation not only share an annual stormwater canopy interception rate very close to most urban forest species [124], but also have very high evapotranspiration rates, both when grown as a single tree and under high-density, short-rotation, coppiced plantations [125]. Moreover, the positive hydrological effects of fast-growing tree species are also linked to the reduced soil bulk density measured over the long term [98].

3.1.4. Carbon Dioxide Sequestration

Although plants represent one of the major environmental carbon sinks, the direct contribution of urban vegetation to carbon sequestration has been found to be smaller compared to the anthropogenic emissions of cities [126]. In fact, only young, fast-growing urban trees display a positive net carbon dioxide sequestration (CO₂), which, unfortunately, decreases as the plants mature [127]. In addition, the maintenance of urban trees (e.g., pruning, fertilization, irrigation, and removal of dead leaves) also creates CO₂ emissions [128]. However, urban vegetation, especially trees, can actively contribute to carbon

sequestration via soil organic matter accumulation [129] and can contribute to reductions in CO₂ emissions from fossil fuel combustion by decreasing the cooling and heating demand of city buildings [130]. Hence, the actual direct benefit of urban vegetation on carbon sequestration is still under active investigation [131]. From this perspective, an estimation of the contribution that phytoremediation stands can provide to a carbon budget is very difficult. While it is universally established that under environmental stress conditions, the photosynthetic activity of plants is negatively affected, thus reducing the net carbon assimilation [132], the extent of environmental stress conditions in most brownfield sites is not easy to predict. Evidence suggests that contamination in urban brownfield sites is rather low, and as such, the detrimental physiological effects are attenuated on most woody species currently used for phytoremediation [133]. In addition, tolerance to some type of contamination (e.g., heavy metals) can also be enhanced by mycorrhizal associations at the root level, which frequently occur with phytoremediation tree species [134]. Overall, it seems realistic that some phytoremediation species used for brownfield reclamation will contribute to carbon sequestration. Because these species are frequently managed in short-rotation coppice, thus kept under a juvenile status, they exhibit fast growth and high carbon sequestration rates [135]. On a heavy-metal-polluted brownfield, willow was found to sequester higher amounts of CO_2 than maize and rapeseed [68]. Poplar and willow grown on a heavy metal phytoextraction site were estimated to stock up to 26 Mg ha^{-1} CO₂ in woody biomass [136]. Other potential benefits are associated with the buildup of organic matter in the soil. Recent studies have shown urban brownfields to be capable of removing 4–59 Mg CO₂ ha⁻¹ yr⁻¹ through direct precipitation of inorganic carbon [137]. Moreover, the adding of waste compost to soil under different species during brownfield reclamation has been proven to further enhance the build-up of long-term soil organic matter [138]. Another opportunity for CO_2 mitigation is the use of biomass feedstock from phytoremediation. Once a detailed ecotoxicological risk assessment is performed, the biomass obtained from these stands, especially where woody plants are used, could contribute to decreasing CO_2 emissions by replacing fossil fuels and, thus, play an important role as environmentally renewable global energy suppliers [139].

3.1.5. Biodiversity

The urban landscape is ecologically characterized by habitat fragmentation and is often associated with lower biological richness than natural ecosystems [140]. When plant species richness is high, it is often due to the occurrence of exotic plants that have been introduced accidentally or deliberately for ornamental purposes. Urban brownfields, which are primarily shaped by disturbance, show typical traits of early-stage secondary succession, in which communities are characterized by non-specialist opportunistic species. These communities are represented by annual plants that predominate the early developmental stages of a brownfield, followed by perennials that usually gain dominance in the succeeding stages [89]. Sometimes, resource limitation, which often occurs in urban brownfields (e.g., water, nutrient, and poor soil quality), can prolong the pioneering stages of the succession process [141]. However, the highly diversified ecological conditions can create a unique and impressive diversity of plant [142] and animal species [143], although the stress associated with high pollution levels recorded in some sites tends to favour the establishment of a few species that exhibit resistance traits [144]. A recent study on a brownfield in Spain abandoned for more than 20 years and polluted with arsenic and lead revealed the presence of a diverse tolerant flora, which included six plant species (i.e., Lotus hispidus, Medicago lupulina, Plantago lanceolata, Dysphania botrys, Trifolium repens, and Lotus *corniculatus*) [145]. Moreover, animal diversity is often affected by the renaturation of urban brownfields. A recent long-term study has shown that the unique ecological communities that can develop on abandoned brownfield allow for a high biological landscape diversity in terms of birds and insects [146]. From this perspective, phytoremediation represents an interesting tool to maintain high levels of biodiversity during reclamation. Unlike most traditional remediation technologies, which are associated with high environmental footprints, including a loss in biodiversity, phytoremediation has less of an impact [147]. While the establishment of a phytoremediation cover on brownfields requires appropriate land preparation that may have negative effects on some soil animal species, thereby leading to an initial decrease of biodiversity [148], over the long-term, this loss of biodiversity is generally only temporary. For example, stands of Salicaceae frequently used for phytoremediation have been shown to display higher plant species richness than agricultural land [149], and sometimes even higher than grasslands and marginal grassland strips [150]. In this case, the use of dense tree or shrub stands on brownfields can also prevent the establishment of invasive plants [151]. Moreover, these types of vegetation cover show overall increases in the abundance of birds and mammals [152], butterflies [153], arthropods [154], and earthworms [155] than agricultural land and residual habitat (i.e., urban areas). Furthermore, choosing different genotypes with varying growth habits can be helpful in creating different habitats, thereby attracting a larger diversity of animals [156]. Although some agronomic operations, such as chemical weed control, which suppresses understory vegetation, may temporarily reduce animal diversity by simplifying the heterogeneity of the habitat [157], most phytoremediation techniques play an important role in enhancing microbial diversity in the soil and increasing the relative abundance of plant-growth-promoting bacteria [158]. Interestingly, some authors have reported that the fungal community composition was directly related to the willow phylogeny following a phytoremediation study using various willow species on the site of a former petrochemical plant [159]. Enrichment in bacterial community structure and diversity was also observed where phytoremediation was supported by the incorporation of inorganic/organic amendments [160].

3.1.6. Air Pollution

Many cities worldwide are currently experiencing severe air pollution as the most serious hazard for human health. Most of the urban air pollutants, which includes particulate matter (PM_{10} , $PM_{2.5}$, and $PM_{<1}$), NO_x , SO_x , carbon monoxide, and ozone, originate from car traffic and transportation [161]. Air quality in urban areas is strongly affected by the presence of vegetation and its structure, and it is now widely acknowledged that vegetation has positive effects on the air quality of urban areas, thereby improving the liveability levels [162]. In the urban context especially, vegetated areas intercept, modify, and reduce the fluxes of air pollutants through a filtering action, both via the deposition of solid pollutants on leaf surfaces and the uptake of gaseous pollutants by stomata. While surface deposition is the quantitatively predominant mechanism for the attenuation of solid air pollutants, gaseous pollutants, such as O_3 , SO_x , and NO_x , are most likely removed via leaf stomatal uptake [163]. The ability of urban vegetation to intercept air pollutants depends on many factors, including physical urban traits (e.g., shape and size of streets), and traits related to vegetation, such as leaf longevity and phenology [164], leaf size and shape [165], and foliage density and porosity [166]. Moreover, leaf functional traits (including leaf surface free energy, single leaf area, surface roughness, specific leaf area, epicuticular wax content, and width-to-length ratio) are among the most important in determining the actual air pollution interception by vegetation in urban areas [167]. Although it is likely that the establishment of vegetation on a brownfield site could show some effect on intercepting airborne particles by acting as physical barriers, evidence is still lacking. Some woody species (e.g., willow, poplar, eucalyptus, etc.) used for environmental reclamation purpose are characterized by a shrubby structure and show similar traits (e.g., high canopy densities, leaf area index, etc.) than many vegetated barriers commonly used for air pollution attenuation along roads and highways [168]. For instance, evergreen shrub species, such as Osmanthus spp., Nerium spp., Eucalyptus spp., and Mimosa spp., are likely to show higher efficiency in airborne particle filtration than willows and poplars due to their ability to intercept and retain air pollutants all year round [169]. By contrast, plants can sometimes contribute to air pollution. They can enhance the air PM concentration through the pollen released at specific times of the year [170]. The negative effects of pollen on human health will be discussed in a subsequent section (see Section 3.3.3). Plant leaves may release numerous biogenic volatile organic

compounds (BVOCs) that react with atmospheric NO_x and contribute to the formation of O_3 . In addition, BVOCs can contribute to $PM_{2.5}$ formation, thereby reducing the overall air quality in towns [171]. Unfortunately, some common species used for brownfield reclamation (e.g., poplar) belong to the high-BVOC emitters and can negatively affect air quality if planted in very large numbers [172]. The negative impact on air quality could be also exacerbated by the fact that significant amounts of BVOCs can be emitted in response to environmental stresses [173], including soil pollution. Therefore, phytoremediation approaches that make use of these plants on heavily contaminated soils could represent a potential source of atmospheric pollutants. However, most urban brownfield areas are usually not so contaminated as to trigger a severe stress response in plants [174], thereby reducing BVOC emissions and their negative impact on the air quality. Future investigation is required into this specific topic.

3.1.7. Noise Attenuation

Urban greenspaces show noise-absorption properties and can reduce road traffic noise to a higher degree in comparison to most artificial barriers [175]. These properties are determined by the coexistence of physical [176] and psychological [177] factors. The presence of both soil and plants represents the main physical factor affecting noise attenuation. Vegetation consists of a multilayer structure, containing high amounts of both living and decayed materials (leaves, needles, branches, and decayed trunks). It is well established that the soil under vegetation covers has a pronounced influence on reducing low-frequency sound propagation [178]. Further, vegetation shows the potential of noise reduction, which depends on the species and structure of the stand. Noise attenuation by woody vegetation along streets has been positively correlated with the height and depth of the stand [179], and it also depends on plant density [180]. To date, noise attenuation in phytoremediation stands along streets and highways has never been assessed. Only a few published studies have reported noise attenuation for the most common woody species used for phytoremediation. Black poplar (Populus nigra L.) fences have been tested for noise attenuation along a highway in Erzurum, Turkey, showing positive results [181]. Less promising performances have been provided by Eucalyptus camaldulensis (Dehnh, 1832) fences near Alexandria (Egypt) [182]. Despite the lack of information for specific phytoremediation crops, it is not difficult to predict higher noise attenuation performances for approaches that make use of high-density shrubby tree species compared to herbaceous plants.

3.2. Provisioning Ecosystem Services

Polluted brownfields under reclamation using green approaches could potentially provide different provisioning services related to the plant species that are used. While using plants grown on a polluted site for food/fodder production is probably unrealistic [183] for safety reasons, much more realistic is their use for energy production within the emerging circular economy framework [184].

3.2.1. Biomass for Bioenergy

One of the most promising provisional services related to the phytoremediation of brownfields is the possibility to use the lignocellulosic biomass issued from the site for energy purposes. The use of brownfields for biomass production, instead of areas where food crops are produced, could address the issue of the food versus fuel debate [185]. The potential of fast-growing woody plant species as sources of biomass, with high yield potential, low conversion and production costs, and an energy-efficient and sustainable value chain, has been well established [186]. Many species used in brownfield phytoremediation, such as poplars and willows, can produce high amounts of biomass, even on polluted sites [187]. The lignocellulosic biomass issued from these sites can potentially be used either as woodchips to provide energy for heat and electricity production or converted into a biofuel (bioethanol) using a variety of methods that differ largely in the way cellulose is hydrolysed [64]. Although purification equipment is usually effective in reducing the

environmental risks, some concerns have been raised about the danger of spreading some toxicants (i.e., heavy metals) into the environment via fly ashes during the combustion process [188]. Previous studies have demonstrated that the concentration of some contaminants (heavy metals) in phytoremediation-borne biomass could be significantly higher than those of a reference biomass [78]. Consequently, it is crucial to better understand the fate of potential pollutants in the currently used fractionation processes and the possible dispersion of hazardous contaminants in the environment during the treatment.

3.2.2. Bioindustry

Unlike the production of thermal energy, the conversion of contaminated biomass into added-value compounds and materials provides a new idea for the green treatment of contaminated biomass and is beneficial to the improvement of phytoremediation technology with fewer environmental and health risks [189]. Many species used in PE naturally produce several compounds suitable for industrial uses. For instance, willows (genus Salix), with 330–500 species worldwide, are a valuable source of biologically active compounds, such as flavonoids, phenolic and non-phenolic glycosides, organic acids, sterols, terpenes, and lignans, all with high economic potential [190]. Since plants growing on TEcontaminated hotspots are subjected to multiple environmental stresses, and phytochemical production in plants is enhanced by them, PE hotspots may become active, added-value phytochemical factories that enhance the overall environmental and economical values of PE. Phytochemicals are a growing revenue-generating industry. Plants can produce over 8000 phenolic compounds to perform a wide range of functions, including abiotic and biotic stress tolerance [191], many of which have commercial uses. For example, condensed tannins can be used as green alternatives to synthetic compounds used in adhesive production [192], as well as environmental-friendly bioflocculants and biocoagulants [193]. Lignans are thought to be effective in mammals in vivo as antioxidants, having potential for cancer chemoprevention, as well as anti-inflammatory activity [194]. They are extractable from biomass without negatively impacting other end uses. Many of the plants used for PE have the potential to produce a large array of phytochemicals. Even some herbaceous species, such as lemongrass, can provide heavy-metal-free, value-added chemicals (e.g., essential oils) after being processed by steam distillation [195]. The prospect of integrating value-added renewable chemicals as a supplementary component of the crop's value has been shown to be feasible under non-stressed conditions [196]. However, evidence is still missing on whether and to what extent plants grown in a heavy-metal-polluted environment would enhance their phenolic compound yield, thus adding economic value to a site under PE.

3.3. Cultural Ecosystem Services

3.3.1. Aesthetics

The aesthetic function can be easily associated with any approaches aimed at increasing the green vegetated area in the urban context. In this sense, all urban vegetation used to create pleasing visual compositions and to provide perfume and auditory effects shows an aesthetic value [197]. Urban vegetation, especially trees and shrubs, is used extensively throughout ornamental horticulture and is particularly appreciated because of its inherent beauty based on the structure, form, foliage pattern, and changing nature of the fruit, flowers, and leaves [198]. Stem height, canopy size, leaf colour, branching height, and canopy density have been found to be the most important traits in predicting the public's aesthetic preference for trees in urban contexts [199]. Trees and shrubs have been shown to enhance most people's aesthetic experience in the short and long term compared to flowers and, more generally, herbaceous vegetation [200]. As such, green technologies, including phytoremediation, unlike most conventional engineered techniques, have the potential to be aesthetically appealing for citizens, thereby increasing the overall value of the approach. Moreover, some plant species currently used for phytoremediation show some interesting traits of particularly high aesthetic value. For instance, sunflower, which is currently used as a phytoextraction annual plant [201], is also popular as an ornamental plant species for its highly aesthetic features [202]. Likewise, some species of willow used in phytotechnologies display a remarkable range of bark colour, from dark brown and purple to light yellow, providing a visually pleasing scene in winter, especially when planted in clusters [203]. In other common phytoremediation species, such as eucalyptus, the colour and shape of the bark vary greatly, not only among species, but with age [204]. Among these species are those with the capacity to be managed as short-rotation coppice, and as such, can be established at high densities along roads or on the edges of reclamation sites, creating a living visual barrier to screen unsightly objects and enhance the overall aspect of the area.

3.3.2. Social Cohesion

Likewise, any approach aimed at brownfield reclamation phytoremediation can contribute to enhancing community cohesion. However, the use of green technologies could be even more successful if properly managed upon implementation. Evidence exists that the stakeholder engagement during planning and management is critical when proposing phytoremediation because this green approach is suited for sites where multiple end-uses are often envisaged [205]. Social cohesion could be attained through different strategies. Citizen science, where stakeholders from the non-scientific community are invited to participate in a research project in both scientific thinking, management, and data collection, is a very useful tool [206]. The citizen science approach has been used to address environmental pollution concerns through the collection of data for environmental management. This approach has been successfully used to improve the public's understanding of air pollution and eventually reducing their personal exposure to contamination [207]. Brownfield reclamation through phytoremediation offers a huge potential for citizen scientist programs. First, phytoremediation requires a long timeframe to operate, frequent sampling campaigns to collect data, and a level of financial support that is, in many cases, limited [208]. The involvement of the non-scientist public in all aspects of a phytoremediation project (planning, implementation, maintenance, and evaluation) could result in increased public awareness about contaminated sites and green sustainable solutions to address environmental hazards and emphasize the important social role of learning about the remediation/reclamation of soil contamination. Stakeholders, if properly trained, can also be engaged in the management of some operations, such as stand establishment (e.g., tree planting) and/or maintenance (e.g., watering and weeding), which stimulates cooperative working, mutual learning, and experience-sharing, thereby increasing the overall social acceptability of a reclamation project [209]. Despite these positive features, some aspects of the green remediation of urban brownfields are still unclear. Some research has shown that the clean-up and revalorization of urban brownfields may increase the risk of gentrification, whereby lower-income communities are displaced elsewhere due to increasing the overall cost-of-living [210]. This phenomenon should be offset, and in this regard, new governance modes and larger-scale participation might be a step in the right direction to overcome this challenge. However, the political and power aspect that is inherent within inequality issues needs to be simultaneously addressed, as demanded by some researchers [211].

3.3.3. Effects on Health

There is increasing evidence that green spaces in urban areas produce measurable benefits on psychological and physical health, including daily stress attenuation [212], increased self-discipline [213], and decreased anxiety, stress, and depression [214], and a generalised improvement of health conditions [215]. Although these functions are mostly associated with the full physical fruition of city green spaces (for example, walking, resting, and running activities within the green space), there is also increasing consensus that specific benefits related to human health can be provided by urban greening indirectly, without having to physically visit these spaces. For instance, during the recent COVID-19 pandemic period, the beneficial effect of a green view on people's mental health has been proven to be

stronger than that of the direct use of greenspace [216]. Other studies have shown that the green view through windows is associated with faster recovery from illness [217]. In this regard, green approaches for brownfield reclamation could provide these type of ecosystem services. In this sense, since the higher psychological benefits are reached when plant species richness is high, a mixture of different species/clones at the reclamation site would be strongly advisable. However, plants used for the reclamation of polluted sites should also be evaluated for any potential drawback they may show. One of the most common problems with plants is represented by their allergenic hazard through their pollen. Some woody species used for phytoremediation show some allergenic potential. For example, birch and alder, which are sometimes used in phytoremediation, are considered very highly allergenic, and their use in urban areas should be avoided [218]. Eucalyptus is another plant that may pose some concerns in terms of pollen allergenicity [219]. Some Salicaceae, such as *Populus* and *Salix*, which produce high quantities of highly allergenic pollen, are also considered risky [220]. However, these two genera are dioecious plants, and the choice of female plant material can thus attenuate the risk of pollen emission and the related health hazard. Moreover, the common management practice of short-rotation coppice keeps the stems of the plants to a juvenile phase, where flowering and pollen production is very rare. Another perceived hazard associated to phytoremediation is the release of small quantities of toxicants into the environment, which could represent a risk to human health. Though in some cases a release of some pollutants has been reported, the quality of the surroundings was not significantly affected [51]. In this regard, the management of the stand is of paramount importance to reduce such risks. For example, when operating with the goal of the phytoextraction of heavy metals from the soil, all aerial parts of the plant should be harvested before leaf shedding. This helps also in removing pollutants from the system entirely and avoiding the return of the pollutants to the soil via litterfall.

4. Conclusions

The phytomanagement of polluted brownfields may represent an avenue to meet the current sustainable development and planning goals for modern urban areas. Nevertheless, its practical application on a full-scale is still challenging due to a number of technical constraints that must be better understood and eventually overcome. In the meanwhile, a new perspective for looking at this approach is proposed that aims at valuing all potential side ecological services associated with this phytotechnology. Furthermore, by taking the type and intensity of pollution within sites into consideration, a system of prioritization of the different ESs possible could be created to determine the specific phytotechnology applied. For instance, areas of low pollution presenting lower risk for human health could prioritize social and cultural ESs, such as providing open green spaces, opportunities for community engagement, and environmental education in an urban setting. In these cases, the specific needs of the local community should be accessed to see how the site could be best incorporated into the social landscape of the city. In contrast, areas of intense pollution should instead prioritize provisional and regulating ESs, limiting direct community involvement. However, the priorities for intensely polluted sites could change over time as soil conditions improve, and as several studies have shown, there are mental and physical health benefits to be derived from simple visual exposure to vegetated areas. These sites in particular should be accessed and planned with a long-term vision in mind, leaving room for shifting priorities from provisional to social and cultural. Although some phytomanagement techniques using woody species are likely to provide similar services to those of urban forests, such as the specificity of most brownfields (i.e., pollution, inaccessibility, and harsh environmental condition for plants), a better understanding of the extent of these services under these particular environmental conditions is fundamental and represents one of the research topics to be investigated by multidisciplinary teams in the near future. Given the long-term investment and timeframe of these projects, the actual related benefits (and sometimes hazards) associated with phytomanagement should be carefully considered. While ESs emanating from phytotechnologies provide both tangible

and intangible products, the risks and costs associated with the management of polluted biomass and increased allergens should be evaluated, and solutions and alternatives found, aided by improved technologies for waste management and greater knowledge of plant responses to abiotic stresses. Future studies focused on evaluating the efficacy and efficiency of phytotechnologies should aim to incorporate an analysis of the associated ecosystems systems, and where possible, provide opportunities for long-term monitoring. This comprehensive approach evaluating the environmental, social, and economic costs and benefits of phytoremediation and continually building on a foundation of information from field-based trials could allow this green technology to be used on a larger scale in our cities for their sustainable, brownfield regeneration processes.

Author Contributions: Conceptualization, W.G.N. and M.L.; methodology, W.G.N., M.L., S.C., F.G. and M.C.P.; writing—original draft preparation, W.G.N. and M.L.; writing—review and editing, W.G.N., S.C., F.G. and M.C.P.; visualization, W.G.N.; supervision, M.L. and S.C.; project administration, M.C.P. and M.L. funding acquisition, M.C.P. and M.L. All authors have read and agreed to the published version of the manuscript.

Funding: This research was funded by the "National Biodiversity Future Center—NBFC"—National Recovery and Resilience Plan (NRRP), Mission 4 Component 2 Investment 1.4—Call for tender No. 3138 of 16 December 2021, rectified by Decree n.3175 of 18 December 2021 of Italian Ministry of University and Research funded by the European Union—Next Generation EU, Project code CN_00000033, and by Fondazione Alia Falck, grant number 2022-NOECO-0198/PER.

Data Availability Statement: No data are available for this research.

Acknowledgments: The authors are grateful to Emily Palm for the revision of the English language and critical reading of the manuscript.

Conflicts of Interest: The authors declare no conflict of interest.

References

- Shmaefsky, B.R. Principles of Phytoremediation. In *Phytoremediation. Concepts and Strategies in Plant Sciences*; Shmaefsky, B., Ed.; Springer: Cham, Switzerland, 2020; pp. 1–26. [CrossRef]
- Prabakaran, K.; Li, J.; Anandkumar, A.; Leng, Z.; Zou, C.B.; Du, D. Managing environmental contamination through phytoremediation by invasive plants: A review. *Ecol. Eng.* 2019, 138, 28–37. [CrossRef]
- 3. Mudgal, V.; Raninga, M.; Patel, D.; Ankoliya, D.; Mudgal, A. A review on Phytoremediation: Sustainable method for removal of heavy metals. *Mater. Today Proc.* 2022, *in press.* [CrossRef]
- 4. Gavrilescu, M. Enhancing phytoremediation of soils polluted with heavy metals. *Curr. Opin. Biotechnol.* **2022**, *74*, 21–31. [CrossRef] [PubMed]
- 5. United Nations. *World Urbanization Prospects 2018*; Department of Economic and Social Affairs, Population Division: New York, NY, USA, 2019.
- 6. De Sousa, C. Brownfield Redevelopment versus Greenfield Development: A Private Sector Perspective on the Costs and Risks Associated with Brownfield Redevelopment in the Greater Toronto Area. J. Environ. Plan. Manag. 2000, 43, 831–853. [CrossRef]
- Tang, Y.-T.; Nathanail, C.P. Sticks and Stones: The Impact of the Definitions of Brownfield in Policies on Socio-Economic Sustainability. *Sustainability* 2012, *4*, 840–862. [CrossRef]
- 8. De Sousa, C.A. Brownfields Redevelopment and the Quest for Sustainability; Emerald Group Publishing: Bingley, UK, 2008; Volume 3.
- 9. Swickard, T.J. Regulatory Incentives to Promote Private Sector Brownfield Remediation and Reuse. *Soil Sediment Contam. Int. J.* **2008**, *17*, 121–136. [CrossRef]
- 10. Grimski, D.; Ferber, U. Urban brownfields in Europe. Land Contam. Reclam. 2001, 9, 143–148.
- 11. Simons, R.A. *Turning Brownfields into Greenbacks: Developing and Financing Environmentally Contaminated Urban Real Estate;* Urban Land Institute: Washington, DC, USA, 1998.
- Thornton, I.; Farago, M.E.; Thums, C.R.; Parrish, R.R.; McGill, R.A.; Breward, N.; Fortey, N.; Simpson, P.; Young, S.; Tye, A. Urban geochemistry: Research strategies to assist risk assessment and remediation of brownfield sites in urban areas. *Environ. Geochem. Health* 2008, 30, 565–576. [CrossRef]
- 13. Hellawell, E.E.; Hughes, S.J. Asbestos contamination on brownfield development sites in the UK. *Environ. Res.* **2020**, *198*, 110480. [CrossRef]
- 14. Yang, Y.; Li, C.; Chen, Z.; Dong, Y.; Zhang, N.; Wei, Y.; Xi, H.; Wang, W. Characterization and Assessment of Organic Pollution at a Fumaric Acid Chemical Brownfield Site in Northwestern China. *Sustainability* **2022**, *14*, 12476. [CrossRef]
- 15. Modabberi, S.; Tashakor, M.; Soltani, N.S.; Hursthouse, A.S. Potentially toxic elements in urban soils: Source apportionment and contamination assessment. *Environ. Monit. Assess.* **2018**, *190*, 715. [CrossRef] [PubMed]

- 16. European Union. Gentle Remediation of Trace Element-Contaminated Land; European Commission: Brussels, Belgium, 2014.
- Bambra, C.; Cairns, J.M.; Kasim, A.; Smith, J.; Robertson, S.; Copeland, A.; Johnson, K. This divided land: An examination of regional inequalities in exposure to brownfield land and the association with morbidity and mortality in England. *Health Place* 2015, 34, 257–269. [CrossRef] [PubMed]
- Lodge, E.K.; Engel, L.S.; Ferrando-Martínez, S.; Wildman, D.; Uddin, M.; Galea, S.; Aiello, A.E. The association between residential proximity to brownfield sites and high-traffic areas and measures of immunity. *J. Expo. Sci. Environ. Epidemiol.* 2020, 30, 824–834. [CrossRef] [PubMed]
- 19. Reddy, K.R.; Adams, J.A.; Richardson, C. Potential Technologies for Remediation of Brownfields. *Pract. Period. Hazard. Toxic Radioact. Waste Manag.* **1999**, *3*, 61–68. [CrossRef]
- 20. Khan, F.I.; Husain, T.; Hejazi, R. An overview and analysis of site remediation technologies. *J. Environ. Manag.* 2004, 71, 95–122. [CrossRef] [PubMed]
- Wang, Y.; O'Connor, D.; Shen, Z.; Lo, I.M.; Tsang, D.C.; Pehkonen, S.; Pu, S.; Hou, D. Green synthesis of nanoparticles for the remediation of contaminated waters and soils: Constituents, synthesizing methods, and influencing factors. *J. Clean. Prod.* 2019, 226, 540–549. [CrossRef]
- Hou, D.; Song, Y.; Zhang, J.; Hou, M.; O'Connor, D.; Harclerode, M. Climate change mitigation potential of contaminated land redevelopment: A city-level assessment method. *J. Clean. Prod.* 2018, 171, 1396–1406. [CrossRef]
- Garon, K.P. Sustainability Analysis for Improving Remedial Action Decisions; Association of State and Territorial Solid Waste Management Offices: Scottsdale, AZ, USA, 2008.
- 24. O'Connor, D.; Zheng, X.; Hou, D.; Shen, Z.; Li, G.; Miao, G.; O'Connel, S.; Guo, M. Phytoremediation: Climate change resilience and sustainability assessment at a coastal brownfield redevelopment. *Environ. Int.* **2019**, *130*, 104945. [CrossRef]
- 25. McGrath, S.P.; Sidoli, C.M.D.; Baker, A.J.M.; Reeves, R.D. *The Potential for the Use of Metal-Accumulating Plants for the In-Situ Decontamination of Metal-Polluted Soils*; Kluwer: Dordrecht, The Netherlands, 1993.
- 26. Suman, J.; Uhlik, O.; Viktorova, J.; Macek, T. Phytoextraction of heavy metals: A promising tool for clean-up of polluted environment? *Front. Plant Sci.* 2018, *9*, 1476. [CrossRef]
- 27. Ahmad, I.; Malik, S.A.; Saeed, S.; Rehman, A.; Munir, T.M. Phytoremediating a Wastewater-Irrigated Soil Contaminated with Toxic Metals: Comparing the Efficacies of Different Crops. *Soil Syst.* **2022**, *6*, 77. [CrossRef]
- Koptsik, G.N. Problems and prospects concerning the phytoremediation of heavy metal polluted soils: A review. *Eurasian Soil Sci.* 2014, 47, 923–939. [CrossRef]
- 29. Nouri, J.; Khorasani, N.; Lorestani, B.; Karami, M.; Hassani, A.H.; Yousefi, N. Accumulation of heavy metals in soil and uptake by plant species with phytoremediation potential. *Environ. Earth Sci.* **2009**, *59*, 315–323. [CrossRef]
- 30. Garbisu, C.; Alkorta, I. Phytoextraction: A cost-effective plant-based technology for the removal of metals from the environment. *Bioresour. Technol.* **2001**, 77, 229–236. [CrossRef]
- Rosselli, W.; Keller, C.; Boschi, K. Phytoextraction capacity of trees growing on a metal contaminated soil. *Plant Soil* 2003, 256, 265–272. [CrossRef]
- Yan, L.; Van Le, Q.; Sonne, C.; Yang, Y.; Yang, H.; Gu, H.; Ma, N.L.; Lam, S.S.; Peng, W. Phytoremediation of radionuclides in soil, sediments and water. J. Hazard. Mater. 2021, 407, 124771. [CrossRef] [PubMed]
- Prasad, J.; Tiwari, S.; Singh, B.K.; Dubey, N.K. Phytoextraction of heavy metals: Challenges and opportunities. In *Phytoremediation* Technology for the Removal of Heavy Metals and Other Contaminants from Soil and Water; Elsevier Inc.: Amsterdam, The Netherlands, 2022; pp. 173–187.
- 34. Gleba, D.; Borisjuk, N.V.; Borisjuk, L.G.; Kneer, R.; Poulev, A.; Skarzhinskaya, M.; Dushenkov, S.; Logendra, S.; Gleba, Y.Y.; Raskin, I. Use of plant roots for phytoremediation and molecular farming. *Proc. Natl. Acad. Sci. USA* **1999**, *96*, 5973–5977. [CrossRef]
- 35. Jiang, Y.; Lei, M.; Duan, L.; Longhurst, P. Integrating phytoremediation with biomass valorisation and critical element recovery: A UK contaminated land perspective. *Biomass Bioenergy* **2015**, *83*, 328–339. [CrossRef]
- Pajević, S.; Borišev, M.; Nikolić, N.; Arsenov, D.D.; Orlović, S.; Župunski, M. Phytoextraction of Heavy Metals by Fast-Growing Trees: A Review. In *Phytoremediation: Management of Environmental Contaminants*; Ansari, A.A., Gill, S., Gill, R., Lanza, G., Newman, L., Eds.; Springer International Publishing: Cham, Switzerland, 2016; Volume 3, pp. 29–64.
- 37. Bolan, N.S.; Park, J.H.; Robinson, B.; Naidu, R.; Huh, K.Y. Chapter four–Phytostabilization: A Green Approach to Contaminant Containment. In *Advances in Agronomy*; Sparks, D.L., Ed.; Academic Press: Cambridge, MA, USA, 2011; Volume 112, pp. 145–204.
- Singh, N.P.; Santal, A.R. Phytoremediation of Heavy Metals: The Use of Green Approaches to Clean the Environment. In *Phytoremediation: Management of Environmental Contaminants*; Ansari, A.A., Gill, S., Gill, R., Lanza, G., Newman, L., Eds.; Springer International Publishing: Cham, Switzerland, 2015; Volume 2, pp. 115–129.
- Conesa, H.M.; García, G.; Faz, Á.; Arnaldos, R. Dynamics of metal tolerant plant communities' development in mine tailings from the Cartagena-La Unión Mining District (SE Spain) and their interest for further revegetation purposes. *Chemosphere* 2007, 68, 1180–1185. [CrossRef]
- Hupy, J.P. Influence of vegetation cover and crust type on wind-blown sediment in a semi-arid climate. J. Arid. Environ. 2004, 58, 167–179. [CrossRef]
- Wenzel, W.; Bunkowski, M.; Puschenreiter, M.; Horak, O. Rhizosphere characteristics of indigenously growing nickel hyperaccumulator and excluder plants on serpentine soil. *Environ. Pollut.* 2003, 123, 131–138. [CrossRef]
- 42. Newman, L.A.; Reynolds, C.M. Phytodegradation of organic compounds. Curr. Opin. Biotechnol. 2004, 15, 225–230. [CrossRef]

- 43. Germaine, K.J.; Byrne, J.; Liu, X.; Keohane, J.; Culhane, J.; Lally, R.D.; Kiwanuka, S.; Ryan, D.; Dowling, D.N. Ecopiling: A combined phytoremediation and passive biopiling system for remediating hydrocarbon impacted soils at field scale. *Front. Plant Sci.* **2015**, *5*, 756. [CrossRef] [PubMed]
- Ouvrard, S.; Barnier, C.; Bauda, P.; Beguiristain, T.; Biache, C.; Bonnard, M.; Caupert, C.; Cebron, A.; Cortet, J.; Cotelle, S. In situ assessment of phytotechnologies for multicontaminated soil management. *Int. J. Phytoremediation* 2011, 13 (Suppl. 1), 245–263. [CrossRef] [PubMed]
- 45. Gerhardt, K.E.; Gerwing, P.D.; Greenberg, B.M. Opinion: Taking phytoremediation from proven technology to accepted practice. *Plant Sci.* 2017, 256, 170–185. [CrossRef] [PubMed]
- Huang, H.; Yu, N.; Wang, L.; Gupta, D.; He, Z.; Wang, K.; Zhu, Z.; Yan, X.; Li, T.; Yang, X.-E. The phytoremediation potential of bioenergy crop Ricinus communis for DDTs and cadmium co-contaminated soil. *Bioresour. Technol.* 2011, 102, 11034–11038. [CrossRef]
- 47. Schwitzguébel, J.-P. Phytoremediation of soils contaminated by organic compounds: Hype, hope and facts. *J. Soils Sediments* **2017**, 17, 1492–1502. [CrossRef]
- 48. Lee, J.H. An overview of phytoremediation as a potentially promising technology for environmental pollution control. *Biotechnol. Bioprocess Eng.* **2013**, *18*, 431–439. [CrossRef]
- Ma, X.; Burken, J.G. TCE diffusion to the atmosphere in phytoremediation applications. *Environ. Sci. Technol.* 2003, 37, 2534–2539. [CrossRef]
- 50. Limmer, M.; Burken, J. Phytovolatilization of Organic Contaminants. Environ. Sci. Technol. 2016, 50, 6632–6643. [CrossRef]
- 51. Narayanan, M.; Davis, L.C.; Erickson, L.E. Fate of Volatile Chlorinated Organic Compounds in a Laboratory Chamber with Alfalfa Plants. *Environ. Sci. Technol.* **1995**, *29*, 2437–2444. [CrossRef]
- Sakakibara, M.; Watanabe, A.; Inoue, M.; Sano, S.; Kaise, T. Phytoextraction and phytovolatilization of arsenic from Ascontaminated soils by Pteris vittata. In Proceedings of the Annual International Conference on Soils, Sediments, Water and Energy, Amherst, MA, USA, 16–19 October 2010; Volume 12, p. 26.
- Zhu, K.; Chen, H.; Nan, Z. Phytoremediation of loess soil contaminated by organic compounds. In *Application of Phytotechnologies* for Cleanup of Industrial, Agricultural, and Wastewater Contamination; Springer: Berlin/Heidelberg, Germany, 2010; pp. 159–176.
- 54. Sivaram, A.K.; Subashchandrabose, S.R.; Logeshwaran, P.; Lockington, R.; Naidu, R.; Megharaj, M. Rhizodegradation of PAHs differentially altered by C3 and C4 plants. *Sci. Rep.* 2020, *10*, 16109. [CrossRef] [PubMed]
- Aisien, F.A.; Aisien, E.T.; Oboh, I.O. Phytoremediation of Petroleum-Polluted Soils. In *Phytoremediation: Management of Environmental Contaminants*; Ansari, A.A., Gill, S., Gill, R., Lanza, G., Newman, L., Eds.; Springer International Publishing: Cham, Switzerland, 2015; Volume 1, pp. 243–252.
- Singh, O.V.; Jain, R.K. Phytoremediation of toxic aromatic pollutants from soil. *Appl. Microbiol. Biotechnol.* 2003, 63, 128–135. [CrossRef] [PubMed]
- 57. Dietz, A.C.; Schnoor, J.L. Advances in phytoremediation. Environ. Health Perspect. 2001, 109 (suppl. 1), 163–168. [PubMed]
- 58. Liang, Y.; Meggo, R.; Hu, D.; Schnoor, J.L.; Mattes, T.E. Enhanced Polychlorinated Biphenyl Removal in a Switchgrass Rhizosphere by Bioaugmentation with Burkholderia xenovorans LB400. *Ecol. Eng.* **2014**, *71*, 215–222. [CrossRef]
- Meggo, R.E.; Schnoor, J.L. Cleaning Polychlorinated Biphenyl (PCB) Contaminated Garden Soil by Phytoremediation. *Environ.* Sci. 2013, 1, 33–52. [CrossRef]
- Pradana, R.; Hernández-Martín, J.A.; Martínez-Hernández, V.; Meffe, R.; de Santiago-Martín, A.; Pérez Barbón, A.; de Bustamante, I. Attenuation mechanisms and key parameters to enhance treatment performance in vegetation filters: A review. *J. Environ. Manag.* 2021, 300, 113752. [CrossRef]
- 61. Blaylock, M.J.; Elless, M.P.; Huang, J.W.; Dushenkov, S.M. Phytoremediation of lead-contaminated soil at a New Jersey Brownfield site. *Remediat. J.* **1999**, *9*, 93–101. [CrossRef]
- Qiu, X.; Leland, T.W.; Shah, S.I.; Sorensen, D.L.; Kendall, E.W. Field Study: Grass Remediation for Clay Soil Contaminated with Polycyclic Aromatic Hydrocarbons. In *Phytoremediation of Soil and Water Contaminants*; American Chemical Society ACS Publications: Washington, DC, USA, 1997; Volume 664, pp. 186–199.
- 63. Van Der Lelie, D.; Schwitzguébel, J.; Glass, D.; Vangronsveld, I.; Baker, A. Peer Reviewed: Assessing phytoremediation's progress in the United States and Europe. *Environ. Sci. Technol.* **2001**, *35*, 446A–452A. [CrossRef]
- Witters, N.; Mendelsohn, R.; Van Passel, S.; Van Slycken, S.; Weyens, N.; Schreurs, E. Phytoremediation, a sustainable remediation technology? II: Economic assessment of CO2 abatement through the use of phytoremediation crops for renewable energy production. *Biomass Bioenergy* 2012, *39*, 470477. [CrossRef]
- Ruttens, A.; Boulet, J.; Weyens, N.; Smeets, K.; Adriaensen, K.; Meers, E.; Van Slycken, S.; Tack, F.; Meiresonne, L.; Thewys, T. Short rotation coppice culture of willows and poplars as energy crops on metal contaminated agricultural soils. *Int. J. Phytoremediation* 2011, 13 (suppl. 1), 194–207. [CrossRef]
- Hattab-Hambli, N.; Motelica-Heino, M.; Mench, M. Aided phytoextraction of Cu, Pb, Zn, and As in copper-contaminated soils with tobacco and sunflower in crop rotation: Mobility and phytoavailability assessment. *Chemosphere* 2016, 145, 543–550. [CrossRef] [PubMed]
- Cundy, A.; Bardos, R.; Church, A.; Puschenreiter, M.; Friesl-Hanl, W.; Müller, I.; Neu, S.; Mench, M.; Witters, N.; Vangronsveld, J. Developing principles of sustainability and stakeholder engagement for "gentle" remediation approaches: The European context. *J. Environ. Manag.* 2013, 129, 283–291. [CrossRef] [PubMed]

- Kidd, P.; Mench, M.; Álvarez-López, V.; Bert, V.; Dimitriou, I.; Friesl-Hanl, W.; Herzig, R.; Olga Janssen, J.; Kolbas, A.; Müller, I. Agronomic practices for improving gentle remediation of trace element-contaminated soils. *Int. J. Phytoremediation* 2015, 17, 1005–1037. [CrossRef] [PubMed]
- 69. European Union. Communication from the commission regions for economic change. In *Staff Working Paper SEC*; European Commission: Brussels, Belgium, 2006; Volume 1432, pp. 1–7.
- 70. Ware, S.; Johnstone, C.; Sparks, K.; Allan, P.; Bryant, M.; Murray, A. *Power Plants. Phytoremediation Gardens*; Stage One Report; NSW Government, Landcom: Sydney, Australia, 2018.
- 71. Francioso, A.; Kampelmann, S. La Phytoremédiation au Service du Développement Durable–Rapport Final de l'Opération Tournesol; Urban Ecology: Ixelles, Belgium, 2015.
- 72. Zalesny, R.; Eanes, S.; Foen, F. *Phytoremediation of Soils Using Fast-Growing Trees in Vacant Lots and Landfills*; American Public Gardens Association: Wilmington, DE, USA, 2020.
- Robinson, B.H.; Anderson, C.W.N.; Dickinson, N.M. Phytoextraction: Where's the action? J. Geochem. Explor. 2015, 151, 34–40. [CrossRef]
- 74. van der Ent, A.; Baker, A.J.M.; Reeves, R.D.; Pollard, A.J.; Schat, H. Hyperaccumulators of metal and metalloid trace elements: Facts and fiction. *Plant Soil* **2013**, *362*, 319–334. [CrossRef]
- 75. Khalid, M.; Saeed, U.R.; Hassani, D.; Hayat, K.; Zhou, P.; Hui, N. Advances in fungal-assisted phytoremediation of heavy metals: A review. *Pedosphere* **2021**, *31*, 475–495. [CrossRef]
- Gul, I.; Manzoor, M.; Hashim, N.; Shah, G.M.; Waani, S.P.T.; Shahid, M.; Antoniadis, V.; Rinklebe, J.; Arshad, M. Challenges in microbially and chelate-assisted phytoextraction of cadmium and lead–A review. *Environ. Pollut.* 2021, 287, 117667. [CrossRef]
- Massenet, A.; Bonet, A.; Laur, J.; Labrecque, M. Co-planting Brassica napus and Salix nigra as a phytomanagement alternative for copper contaminated soil. *Chemosphere* 2021, 279, 130517. [CrossRef]
- 78. Migeon, A.; Richaud, P.; Guinet, F.; Chalot, M.; Blaudez, D. Metal Accumulation by Woody Species on Contaminated Sites in the North of France. *Water Air Soil Pollut.* **2009**, 204, 89. [CrossRef]
- Edgar, V.-N.; Fabián, F.-L.; Mario, P.-C.J.; Ileana, V.-R. Coupling Plant Biomass Derived from Phytoremediation of Potential Toxic-Metal-Polluted Soils to Bioenergy Production and High-Value by-Products—A Review. *Appl. Sci.* 2021, 11, 2982. [CrossRef]
- Danley, B.; Widmark, C. Evaluating conceptual definitions of ecosystem services and their implications. *Ecol. Econ.* 2016, 126, 132–138. [CrossRef]
- 81. Barbier, E.B. Economics of the Regulating Services. In *Encyclopedia of Biodiversity*, 2nd ed.; Levin, S.A., Ed.; Academic Press: Cambridge, MA, USA, 2013; pp. 45–54.
- 82. Millennium Ecosystem Assessment. *Ecosystems and Human Well-Being: Wetlands and Water Synthesis;* World Resources Institute: Washington, DC, USA, 2005.
- 83. Kowarik, I. Cities and wilderness. A new perspective. Int. J. Wilderness 2013, 19, 32–36.
- 84. Mathey, J.; Rößler, S.; Banse, J.; Lehmann, I.; Bräuer, A. Brownfields as an element of green infrastructure for implementing ecosystem services into urban areas. *J. Urban Plan. Dev.* **2015**, *141*, A4015001. [CrossRef]
- 85. Wolch, J.R.; Byrne, J.; Newell, J.P. Urban Green Space, Public Health, and Environmental Justice: The Challenge of Making Cities 'Just Green Enough'. *Landsc. Urban Plan.* **2014**, *125*, 234–244. [CrossRef]
- Sikorski, P.; Gawryszewska, B.; Sikorska, D.; Chormański, J.; Schwerk, A.; Jojczyk, A.; Ciężkowski, W.; Archiciński, P.; Łepkowski, M.; Dymitryszyn, I.; et al. The value of doing nothing—How informal green spaces can provide comparable ecosystem services to cultivated urban parks. *Ecosyst. Serv.* 2021, 50, 101339. [CrossRef]
- 87. Zeng, P.; Guo, Z.; Xiao, X.; Peng, C.; Liao, B.; Zhou, H.; Gu, J. Facilitation of Morus alba L. intercropped with Sedum alfredii H. and Arundo donax L. on soil contaminated with potentially toxic metals. *Chemosphere* **2022**, *290*, 133107. [CrossRef]
- Chapman, E.E.V.; Moore, C.; Campbell, L.M. Native plants for revegetation of mercury- and arsenic-contaminated historical mining waste—Can a low-dose selenium additive improve seedling growth and decrease contaminant bioaccumulation? *Water Air Soil Pollut.* 2019, 230, 225. [CrossRef]
- 89. Schadek, U.; Strauss, B.; Biedermann, R.; Kleyer, M. Plant species richness, vegetation structure and soil resources of urban brownfield sites linked to successional age. *Urban Ecosyst.* **2009**, *12*, 115–126. [CrossRef]
- Blanco-Velázquez, F.J.; Pino-Mejías, R.; Anaya-Romero, M. Evaluating the provision of ecosystem services to support phytoremediation measures for countering soil contamination. A case-study of the Guadiamar Green Corridor (SW Spain). Land Degrad. Dev. 2020, 31, 2914–2924. [CrossRef]
- 91. Robinson, S.L.; Lundholm, J.T. Ecosystem services provided by urban spontaneous vegetation. *Urban Ecosyst.* **2012**, *15*, 545–557. [CrossRef]
- Pueffel, C.; Haase, D.; Priess, J.A. Mapping ecosystem services on brownfields in Leipzig, Germany. *Ecosyst. Serv.* 2018, 30, 73–85. [CrossRef]
- Masiero, M.; Biasin, A.; Amato, G.; Malaggi, F.; Pettenella, D.; Nastasio, P.; Anelli, S. Urban Forests and Green Areas as Nature-Based Solutions for Brownfield Redevelopment: A Case Study from Brescia Municipal Area (Italy). *Forests* 2022, 13, 444. [CrossRef]
- Zhiyanski, M.; Sokolovska, M.; Glushkova, M.; Vilhar, U.; Lozanova, L. Soil Quality. In *The Urban Forest: Cultivating Green* Infrastructure for People and the Environment; Pearlmutter, D., Calfapietra, C., Samson, R., O'Brien, L., Ostoić, S.K., Sanesi, G., del Amo, R.A., Eds.; Springer International: Cham, Switzerland, 2017; pp. 49–58.

- Gil-Díaz, M.; González, A.; Alonso, J.; Lobo, M.C. Evaluation of the stability of a nanoremediation strategy using barley plants. J. Environ. Manag. 2016, 165, 150–158. [CrossRef] [PubMed]
- Teixeira, L.A.J.; Berton, R.S.; Coscione, A.R.; Saes, L.A. Biosolids application on banana production: Soil chemical properties and plant nutrition. *Appl. Environ. Soil Sci.* 2011, 2011, 1–8. [CrossRef]
- Foulon, J.; Zappelini, C.; Durand, A.; Valot, B.; Blaudez, D.; Chalot, M. Impact of poplar-based phytomanagement on soil properties and microbial communities in a metal-contaminated site. *FEMS Microbiol. Ecol.* 2016, 92, fiw163. [CrossRef] [PubMed]
- Kahle, P.; Janssen, M. Impact of short-rotation coppice with poplar and willow on soil physical properties. J. Plant Nutr. Soil Sci. 2020, 183, 119–128. [CrossRef]
- Fiorentino, N.; Ventorino, V.; Rocco, C.; Cenvinzo, V.; Agrelli, D.; Gioia, L.; Di Mola, I.; Adamo, P.; Pepe, O.; Fagnano, M. Giant reed growth and effects on soil biological fertility in assisted phytoremediation of an industrial polluted soil. *Sci. Total Environ.* 2017, 575, 1375–1383. [CrossRef]
- 100. Cicatelli, A.; Lingua, G.; Todeschini, V.; Biondi, S.; Torrigiani, P.; Castiglione, S. Arbuscular mycorrhizal fungi restore normal growth in a white poplar clone grown on heavy metal-contaminated soil, and this is associated with upregulation of foliar metallothionein and polyamine biosynthetic gene expression. *Ann. Bot.* 2010, *106*, 791–802. [CrossRef]
- Fillion, M.; Brisson, J.; Guidi, W.; Labrecque, M. Increasing phosphorus removal in willow and poplar vegetation filters using arbuscular mycorrhizal fungi. *Ecol. Eng.* 2011, 37, 199–205. [CrossRef]
- 102. Burges, A.; Oustriere, N.; Galende, M.; Marchand, L.; Bes, C.M.; Paidjan, E.; Puschenreiter, M.; Becerril, J.M.; Mench, M. Phytomanagement with grassy species, compost and dolomitic limestone rehabilitates a meadow at a wood preservation site. *Ecol. Eng.* 2021, 160, 106132. [CrossRef]
- 103. Kelly, C.N.; Peltz, C.D.; Stanton, M.; Rutherford, D.W.; Rostad, C.E. Biochar application to hardrock mine tailings: Soil quality, microbial activity, and toxic element sorption. *Appl. Geochem.* **2014**, *43*, 35–48. [CrossRef]
- 104. Tordoff, G.M.; Baker, A.J.M.; Willis, A.J. Current approaches to the revegetation and reclamation of metalliferous mine wastes. *Chemosphere* **2000**, *41*, 219–228. [CrossRef] [PubMed]
- Brunner, I.; Luster, J.; Günthardt-Goerg, M.S.; Frey, B. Heavy metal accumulation and phytostabilisation potential of tree fine roots in a contaminated soil. *Environ. Pollut.* 2008, 152, 559–568. [CrossRef] [PubMed]
- 106. Kumar, M.; Bolan, N.; Jasemizad, T.; Padhye, L.P.; Sridharan, S.; Singh, L.; Bolan, S.; O'Connor, J.; Zhao, H.; Shaheen, S.M.; et al. Mobilization of contaminants: Potential for soil remediation and unintended consequences. *Sci. Total. Environ.* 2022, *839*, 156373. [CrossRef] [PubMed]
- Santamouris, M.; Cartalis, C.; Synnefa, A.; Kolokotsa, D. On the impact of urban heat island and global warming on the power demand and electricity consumption of buildings—A review. *Energy Build.* 2015, 98, 119–124. [CrossRef]
- 108. Oliveira, S.; Andrade, H.; Vaz, T. The cooling effect of green spaces as a contribution to the mitigation of urban heat: A case study in Lisbon. *Build. Environ.* **2011**, *46*, 2186–2194. [CrossRef]
- Hsieh, C.-M.; Li, J.-J.; Liman, Z.; Schwegler, B. Effects of tree shading and transpiration on building cooling energy use. *Energy Build.* 2018, 159, 382–397. [CrossRef]
- 110. Raeissi, S.; Taheri, M. Energy saving by proper tree plantation. Build. Environ. 1999, 34, 565–570. [CrossRef]
- 111. Bowler, D.E.; Buyung-Ali, L.; Knight, T.; Pullin, A. Urban greening to cool towns and cities: A systematic review of the empirical evidence. *Landsc. Urban Plan.* 2010, 97, 147–155. [CrossRef]
- 112. Algretawee, H. The effect of graduated urban park size on park cooling island and distance relative to land surface temperature (LST). *Urban Clim.* **2022**, *45*, 101255. [CrossRef]
- 113. Edmondson, J.L.; Stott, I.; Davies, Z.G.; Gaston, K.J.; Leake, J.R. Soil surface temperatures reveal moderation of the urban heat island effect by trees and shrubs. *Sci. Rep.* **2016**, *6*, 33708. [CrossRef]
- 114. Armson, D.; Stringer, P.; Ennos, A.R. The effect of street trees and amenity grass on urban surface water runoff in Manchester, UK. *Urban For. Urban Green.* **2013**, *12*, 282–286. [CrossRef]
- 115. Fischer, M.; Zenone, T.; Trnka, M.; Orság, M.; Montagnani, L.; Ward, E.; Tripathi, A.; Hlavinka, P.; Seufert, G.; Zalud, Z.; et al. Water requirements of short rotation poplar coppice: Experimental and modelling analyses across Europe. *Agric. For. Meteorol.* 2017, 250–251, 343–360. [CrossRef]
- Chandra, R.; Ho-duck, K. Mixed heavy metal stress on photosynthesis, transpiration rate, and chlorophyll content in poplar hybrids. *For. Sci. Technol.* 2016, 12, 55–61. [CrossRef]
- 117. Ahmed, F.; Moors, E.; Khan, M.S.A.; Warner, J.; Van Scheltinga, C.T. Tipping points in adaptation to urban flooding under climate change and urban growth: The case of the Dhaka megacity. *Land Use Policy* **2018**, *79*, 496–506. [CrossRef]
- 118. Sofia, G.; Roder, G.; Dalla Fontana, G.; Tarolli, P. Flood dynamics in urbanised landscapes: 100 years of climate and humans' interaction. *Sci. Rep.* **2017**, *7*, 40527. [CrossRef]
- Tobias, S.; Conen, F.; Duss, A.; Wenzel, L.M.; Buser, C.; Alewell, C. Soil sealing and unsealing: State of the art and examples. *Land Degrad. Dev.* 2018, 29, 2015–2024. [CrossRef]
- 120. Zimmermann, E.; Bracalenti, L.; Piacentini, R.; Inostroza, L. Urban Flood Risk Reduction by Increasing Green Areas For Adaptation To Climate Change. *Procedia Eng.* **2016**, *161*, 2241–2246. [CrossRef]
- 121. Xiao, Q.; McPherson, E.G.; Ustin, S.L.; Grismer, M.E. A new approach to modeling tree rainfall interception. *J. Geophys. Res. Atmos.* **2000**, *105*, 29173–29188. [CrossRef]
- 122. Bolund, P.; Hunhammar, S. Ecosystem services in urban areas. Ecol. Econ. 1999, 29, 293–301. [CrossRef]

- 123. Schooling, J.T.; Carlyle-Moses, D.E. The influence of rainfall depth class and deciduous tree traits on stemflow production in an urban park. *Urban Ecosyst.* **2015**, *18*, 1261–1284. [CrossRef]
- 124. Kermavnar, J.; Vilhar, U. Canopy precipitation interception in urban forests in relation to stand structure. *Urban Ecosyst.* 2017, 20, 1373–1387. [CrossRef]
- 125. Guidi, W.; Piccioni, E.; Bonari, E. Evapotranspiration and crop coefficient of poplar and willow short-rotation coppice used as vegetation filter. *Bioresour. Technol.* 2008, *99*, 4832–4840. [CrossRef] [PubMed]
- 126. Velasco-Jiménez, M.J.; Alcázar, P.; Cariñanos, P.; Galán, C. Allergenicity of the urban green areas in the city of Córdoba (Spain). *Urban For. Urban Green.* **2020**, *49*, 126600. [CrossRef]
- 127. Nowak, D.J.; Greenfield, E.J.; Hoehn, R.E.; Lapoint, E. Carbon storage and sequestration by trees in urban and community areas of the United States. *Environ. Pollut.* 2013, 178, 229–236. [CrossRef]
- Riikonen, A.; Pumpanen, J.; Mäki, M.; Nikinmaa, E. High carbon losses from established growing sites delay the carbon sequestration benefits of street tree plantings–A case study in Helsinki, Finland. Urban For. Urban Green. 2017, 26, 85–94. [CrossRef]
- 129. Downey, A.E.; Groffman, P.M.; Mejía, G.A.; Cook, E.M.; Sritrairat, S.; Karty, R.; Palmer, M.I.; McPhearson, T. Soil carbon sequestration in urban afforestation sites in New York City. *Urban For. Urban Green.* **2021**, *65*, 127342. [CrossRef]
- 130. Ko, Y. Trees and vegetation for residential energy conservation: A critical review for evidence-based urban greening in North America. *Urban For. Urban Green.* **2018**, *34*, 318–335. [CrossRef]
- 131. Fares, S.; Paoletti, E.; Calfapietra, C.; Mikkelsen, T.N.; Samson, R.; Le Thiec, D. Carbon Sequestration by Urban Trees. In *The Urban Forest. Future City*; Springer: Cham, Switzerland, 2017; Volume 7.
- Rennenberg, H.; Loreto, F.; Polle, A.; Brilli, F.; Fares, S.; Beniwal, R.S.; Gessler, A. Physiological Responses of Forest Trees to Heat and Drought. *Plant Biol.* 2006, *8*, 556–571. [CrossRef]
- 133. Radwanski, D.; Gallagher, F.; Vanderklein, D.W.; Schäfer, K.V.R. Photosynthesis and aboveground carbon allocation of two co-occurring poplar species in an urban brownfield. *Environ. Pollut.* **2017**, *223*, 497–506. [CrossRef]
- Lingua, G.; Franchin, C.; Todeschini, V.; Castiglione, S.; Biondi, S.; Burlando, B.; Parravicini, V.; Torrigiani, P.; Berta, G. Arbuscular mycorrhizal fungi differentially affect the response to high zinc concentrations of two registered poplar clones. *Environ. Pollut.* 2008, 153, 137–147. [CrossRef] [PubMed]
- 135. Quinkenstein, A.; Jochheim, H. Assessing the carbon sequestration potential of poplar and black locust short rotation coppices on mine reclamation sites in Eastern Germany–Model development and application. J. Environ. Manag. 2016, 168, 53–66. [CrossRef] [PubMed]
- 136. Riccioli, F.; Guidi Nissim, W.; Masi, M.; Palm, E.; Mancuso, S.; Azzarello, E. Modeling the Ecosystem Services Related to Phytoextraction: Carbon Sequestration Potential Using Willow and Poplar. *Appl. Sci.* **2020**, *10*, 8011. [CrossRef]
- 137. Jorat, M.E.; Goddard, M.A.; Manning, P.; Lau, H.K.; Ngeow, S.; Sohi, S.P.; Manning, D.A. Passive CO2 removal in urban soils: Evidence from brownfield sites. *Sci. Total Environ.* **2020**, *703*, 135573. [CrossRef]
- Lord, R.; Sakrabani, R. Ten-year legacy of organic carbon in non-agricultural (brownfield) soils restored using green waste compost exceeds 4 per mille per annum: Benefits and trade-offs of a circular economy approach. *Sci. Total Environ.* 2019, 686, 1057–1068. [CrossRef] [PubMed]
- Jha, A.B.; Misra, A.N.; Sharma, P. Phytoremediation of Heavy Metal-Contaminated Soil Using Bioenergy Crops. In *Phytoremedia*tion Potential of Bioenergy Plants; Bauddh, K., Singh, B., Korstad, J., Eds.; Springer: Singapore, 2017; pp. 63–96.
- McDonnell, M.J.; Pickett, S.T. Ecosystem structure and function along urban-rural gradients: An unexploited opportunity for ecology. *Ecology* 1990, 71, 1232–1237. [CrossRef]
- 141. Muratet, A.; Machon, N.; Jiguet, F.; Moret, J.; Porcher, E. The role of urban structures in the distribution of wasteland flora in the greater Paris area, France. *Ecosystems* **2007**, *10*, 661–671. [CrossRef]
- 142. Desjardins, D.; Nissim, W.G.; Pitre, F.E.; Naud, A.; Labrecque, M. Distribution patterns of spontaneous vegetation and pollution at a former decantation basin in southern Québec, Canada. *Ecol. Eng.* **2014**, *64*, 385–390. [CrossRef]
- Kattwinkel, M.; Biedermann, R.; Kleyer, M. Temporary conservation for urban biodiversity. *Biol. Conserv.* 2011, 144, 2335–2343.
 [CrossRef]
- 144. Gallagher, F.J.; Pechmann, I.; Holzapfel, C.; Grabosky, J. Altered vegetative assemblage trajectories within an urban brownfield. *Environ. Pollut.* **2011**, *159*, 1159–1166. [CrossRef]
- 145. Matanzas, N.; Afif, E.; Díaz, T.E.; Gallego, J.R. Phytoremediation Potential of Native Herbaceous Plant Species Growing on a Paradigmatic Brownfield Site. *Water Air Soil Pollut.* **2021**, 232, 290. [CrossRef]
- Macgregor, C.J.; Bunting, M.J.; Deutz, P.; Bourn, N.A.; Roy, D.B.; Mayes, W.M. Brownfield sites promote biodiversity at a landscape scale. *Sci. Total Environ.* 2022, 804, 150162. [CrossRef] [PubMed]
- 147. Wang, R.Y.; Huang, S.; Chen, R.; Wang, J. Application for Ecological Restoration of Contaminated Soil: Phytoremediation. *Int. J. Environ. Res. Public Health* **2022**, *19*, 13124.
- 148. Minor, M.A.; Volk, T.A.; Norton, R.A. Effects of site preparation techniques on communities of soil mites (Acari: Oribatida, Acari: Gamasida) under short-rotation forestry plantings in New York, USA. *Appl. Soil Ecol.* **2004**, 25, 181–192. [CrossRef]
- 149. Baum, S.; Bolte, A.; Weih, M. High value of short rotation coppice plantations for phytodiversity in rural landscapes. *Glob. Change Biol. Bioenergy* **2012**, *4*, 728–738. [CrossRef]

- 150. Fry, D.; Slater, F. The Biodiversity of Short Rotation Willow Coppice in the Welsh Landscape. 2009. Available online: http://www.willow4wales.co.uk/ (accessed on 30 November 2022).
- 151. Dommanget, F.; Evette, A.; Breton, V.; Daumergue, N.; Forestier, O.; Poupart, P.; Martin, F.-M.; Navas, M.-L. Fast-growing willows significantly reduce invasive knotweed spread. *J. Environ. Manag.* 2019, 231, 1–9. [CrossRef]
- 152. Riffell, S.; Verschuyl, J.; Miller, D.; Wigley, T.B. A meta-analysis of bird and mammal response to short-rotation woody crops. *Glob. Change Biol. Bioenergy* **2011**, *3*, 313–321. [CrossRef]
- Rowe, R.L.; Hanley, M.E.; Goulson, D.; Clarke, D.J.; Doncaster, C.P.; Taylor, G. Potential benefits of commercial willow Short Rotation Coppice (SRC) for farm-scale plant and invertebrate communities in the agri-environment. *Biomass Bioenergy* 2011, 35, 325–336. [CrossRef]
- 154. Allegro, G.; Sciaky, R. Assessing the potential role of ground beetles (Coleoptera, Carabidae) as bioindicators in poplar stands, with a newly proposed ecological index (FAI). *For. Ecol. Manag.* **2003**, 175, 275–284. [CrossRef]
- 155. Schrama, M.; Vandecasteele, B.; Carvalho, S.; Muylle, H.; van der Putten, W.H. Effects of first-and second-generation bioenergy crops on soil processes and legacy effects on a subsequent crop. *Glob. Change Biol. Bioenergy* **2016**, *8*, 136–147. [CrossRef]
- 156. Stauffer, M.; Leyval, C.; Brun, J.-J.; Leportier, P.; Berthelin, J. Effect of willow short rotation coppice on soil properties after three years of growth as compared to forest, grassland and arable land uses. *Plant Soil* **2014**, 377, 423–438. [CrossRef]
- 157. Sage, R. Short rotation coppice for energy: Towards ecological guidelines. Biomass Bioenergy 1998, 15, 39–47. [CrossRef]
- 158. Liu, C.; Lin, H.; Li, B.; Dong, Y.; Yin, T. Responses of microbial communities and metabolic activities in the rhizosphere during phytoremediation of Cd-contaminated soil. *Ecotoxicol. Environ. Saf.* **2020**, 202, 110958. [CrossRef]
- Bell, T.H.; El-Din Hassan, S.; Lauron-Moreau, A.; Al-Otaibi, F.; Hijri, M.; Yergeau, E.; St-Arnaud, M. Linkage between bacterial and fungal rhizosphere communities in hydrocarbon-contaminated soils is related to plant phylogeny. *ISME J.* 2014, *8*, 331–343. [CrossRef] [PubMed]
- Touceda-González, M.; Prieto-Fernández, Á.; Renella, G.; Giagnoni, L.; Sessitsch, A.; Brader, G.; Kumpiene, J.; Dimitriou, I.; Eriksson, J.; Friesl-Hanl, W.; et al. Microbial community structure and activity in trace element-contaminated soils phytomanaged by Gentle Remediation Options (GRO). *Environ. Pollut.* 2017, 231, 237–251. [CrossRef]
- 161. Khan, J.; Ketzel, M.; Kakosimos, K.; Sørensen, M.; Jensen, S.S. Road traffic air and noise pollution exposure assessment–A review of tools and techniques. *Sci. Total Environ.* **2018**, 634, 661–676. [CrossRef]
- 162. Janhäll, S. Review on urban vegetation and particle air pollution–Deposition and dispersion. *Atmos. Environ.* **2015**, *105*, 130–137. [CrossRef]
- 163. Nowak, D.J.; Hirabayashi, S.; Bodine, A.; Greenfield, E. Tree and forest effects on air quality and human health in the United States. *Environ. Pollut.* 2014, 193, 119–129. [CrossRef]
- Sæbø, A.; Popek, R.; Nawrot, B.; Hanslin, H.M.; Gawronska, H.; Gawronski, S. Plant species differences in particulate matter accumulation on leaf surfaces. Sci. Total Environ. 2012, 427, 347–354. [CrossRef]
- 165. Weerakkody, U.; Dover, J.W.; Mitchell, P.; Reiling, K. Evaluating the impact of individual leaf traits on atmospheric particulate matter accumulation using natural and synthetic leaves. *Urban For. Urban Green.* **2018**, *30*, 98–107. [CrossRef]
- 166. Abhijith, K.V.; Kumar, P.; Gallagher, J.; McNabola, A.; Baldauf, R.; Pilla, F.; Broderick, B.; Di Sabatino, S.; Pulvirenti, B. Air pollution abatement performances of green infrastructure in open road and built-up street canyon environments–A review. *Atmos. Environ.* 2017, 162, 71–86. [CrossRef]
- 167. Zhang, X.; Lyu, J.; Han, Y.; Sun, N.; Sun, W.; Li, J.; Liu, C.; Yin, S. Effects of the leaf functional traits of coniferous and broadleaved trees in subtropical monsoon regions on PM_{2.5} dry deposition velocities. *Environ. Pollut.* 2020, 26, 114845. [CrossRef] [PubMed]
- 168. Jia, Y.-P.; Lu, K.-F.; Zheng, T.; Li, X.-B.; Liu, X.; Peng, Z.-R.; He, H.-D. Effects of roadside green infrastructure on particle exposure: A focus on cyclists and pedestrians on pathways between urban roads and vegetative barriers. *Atmos. Pollut. Res.* 2021, 12, 1–12. [CrossRef]
- 169. Hirabayashi, S.; Nowak, D.J. Comprehensive national database of tree effects on air quality and human health in the United States. *Environ. Pollut.* **2016**, *215*, 48–57. [CrossRef]
- 170. Cariñanos, P.; Adinolfi, C.; Diaz de la Guardia, C.; De Linares, C.; Casares-Porcel, M. Characterization of allergen emission sources in urban areas. *J. Environ. Qual.* **2016**, *45*, 244–252. [CrossRef] [PubMed]
- 171. Arneth, A.; Harrison, S.P.; Zaehle, S.; Tsigaridis, K.; Menon, S.; Bartlein, P.J.; Feichter, J.; Korhola, A.; Kulmala, M.; O'Donnell, D.; et al. Terrestrial biogeochemical feedbacks in the climate system. *Nat. Geosci.* **2010**, *3*, 525–532. [CrossRef]
- 172. Donovan, R.G.; Stewart, H.E.; Owen, S.M.; MacKenzie, A.R.; Hewitt, C.N. Development and Application of an Urban Tree Air Quality Score for Photochemical Pollution Episodes Using the Birmingham, United Kingdom, Area as a Case Study. *Environ. Sci. Technol.* 2005, 39, 6730–6738. [CrossRef]
- 173. Niinemets, Ü. Mild versus severe stress and BVOCs: Thresholds, priming and consequences. *Trends Plant Sci.* **2010**, *15*, 145–153. [CrossRef]
- 174. Guidi Nissim, W.; Palm, E.; Mancuso, S.; Azzarello, E. Trace element phytoextraction from contaminated soil: A case study under Mediterranean climate. *Environ. Sci. Pollut. Res.* 2018, 25, 9114–9131. [CrossRef]
- 175. Yang, F.; Bao, Z.Y.; Zhu, Z.J. An assessment of psychological noise reduction by landscape plants. *Int. J. Environ. Res. Public Health* **2011**, *8*, 1032–1048. [CrossRef]
- 176. Van Renterghem, T.; Botteldooren, D.; Verheyen, K. Road traffic noise shielding by vegetation belts of limited depth. *J. Sound Vib.* **2012**, *331*, 2404–2425. [CrossRef]

- 177. Viollon, S.; Lavandier, C.; Drake, C. Influence of visual setting on sound ratings in an urban environment. *Appl. Acoust.* **2002**, *63*, 493–511. [CrossRef]
- 178. Martens, M.J.M.; van der Heijden, L.A.M.; Walthaus, H.H.J.; van Rens, W.J.J.M. Classification of soils based on acoustic impedance, air flow resistivity, and other physical soil parameters. *J. Acoust. Soc. Am.* **1985**, *78*, 970–980. [CrossRef]
- 179. Fang, C.-F.; Ling, D.-L. Guidance for noise reduction provided by tree belts. Landsc. Urban Plan. 2005, 71, 29–34. [CrossRef]
- 180. Ow, L.F.; Ghosh, S. Urban cities and road traffic noise: Reduction through vegetation. Appl. Acoust. 2017, 120, 15–20. [CrossRef]
- Ozer, S.; Irmak, M.A.; Yilmaz, H. Determination of roadside noise reduction effectiveness of *Pinus sylvestris* L. and *Populus nigra* L. in Erzurum, Turkey. *Env. Monit Assess* 2008, 144, 191–197. [CrossRef]
- 182. Omran, T.A.; Elshorbagy, K.A.; El-Sayed, A.B. Attenuation of noise by windbreaks. Appl. Acoust. 1982, 15, 389–395. [CrossRef]
- 183. Chowdhury, S.; Kain, J.-H.; Adelfio, M.; Volchko, Y.; Norrman, J. Greening the Browns: A Bio-Based Land Use Framework for Analysing the Potential of Urban Brownfields in an Urban Circular Economy. *Sustainability* 2022, 12, 6278. [CrossRef]
- 184. MacArthur, E. Towards the circular economy. J. Ind. Ecol. 2013, 2, 23–44.
- 185. Guldhe, A.; Singh, B.; Renuka, N.; Singh, P.; Misra, R.; Bux, F. Bioenergy: A Sustainable Approach for Cleaner Environment. In Phytoremediation Potential of Bioenergy Plants; Bauddh, K., Singh, B., Korstad, J., Eds.; Springer: Singapore, 2017.
- 186. Rönnberg-Wästljung, A.C.; Dufour, L.; Gao, J.; Hansson, P.-A.; Herrmann, A.; Jebrane, M.; Johansson, A.-C.; Kalita, S.; Molinder, R.; Nordh, N.-E.; et al. Optimized utilization of Salix—Perspectives for the genetic improvement toward sustainable biofuel value chains. GCB Bioenergy 2022, 14, 1128–1144. [CrossRef]
- Chalot, M.; Girardclos, O.; Ciadamidaro, L.; Zappelini, C.; Yung, L.; Durand, A.; Blaudez, D. Poplar rotation coppice at a trace element-contaminated phytomanagement site: A 10-year study revealing biomass production, element export and impact on extractable elements. *Sci. Total Environ.* 2020, 699, 134260. [CrossRef] [PubMed]
- 188. Nzihou, A.; Stanmore, B. The fate of heavy metals during combustion and gasification of contaminated biomass-a brief review. *J. Hazard. Mater.* **2013**, 256–257, 56–66. [CrossRef] [PubMed]
- Chai, Y.; Chen, A.; Bai, M.; Peng, L.; Shao, J.; Yuan, J.; Shang, C.; Zhang, J.; Huang, H.; Peng, C. Valorization of heavy metal contaminated biomass: Recycling and expanding to functional materials. *J. Clean. Prod.* 2022, 366, 132771. [CrossRef]
- 190. Tawfeek, N.; Mahmoud, M.F.; Hamdan, D.I.; Sobeh, M.; Farrag, N.; Wink, M.; El-Shazly, A.M. Phytochemistry, Pharmacology and Medicinal Uses of Plants of the Genus Salix: An Updated Review. *Front. Pharmacol.* **2021**, *12*, 593856. [CrossRef]
- 191. Dai, J.; Mumper, R.J. Plant phenolics: Extraction, analysis and their antioxidant and anticancer properties. *Molecules* **2010**, *15*, 7313–7352. [CrossRef]
- 192. Ping, L.; Brosse, N.; Chrusciel, L.; Navarrete, P.; Pizzi, A. Extraction of condensed tannins from grape pomace for use as wood adhesives. *Ind. Crops Prod.* 2011, *33*, 253–257. [CrossRef]
- Beltrán-Heredia, J.; Sánchez-Martín, J. Municipal wastewater treatment by modified tannin flocculant agent. *Desalination* 2009, 249, 353–358. [CrossRef]
- 194. Wan, M.L.Y.; Co, V.A.; El-Nezami, H. Dietary polyphenol impact on gut health and microbiota. *Crit. Rev. Food Sci. Nutr.* 2021, *61*, 690–711. [CrossRef]
- 195. Lal, K.; Yadav, R.K.; Kaur, R.; Bundela, D.S.; Khan, M.I.; Chaudhary, M.; Meena, R.L.; Singh, G. Productivity, essential oil yield, and heavy metal accumulation in lemon grass (Cymbopogon flexuosus) under varied wastewater–groundwater irrigation regimes. *Ind. Crops Prod.* 2013, 45, 270–278. [CrossRef]
- Brereton, N.J.; Berthod, N.; Lafleur, B.; Pedneault, K.; Pitre, F.E.; Labrecque, M. Extractable phenolic yield variation in five cultivars of mature short rotation coppice willow from four plantations in Quebec. *Ind. Crops Prod.* 2017, 97, 525–535. [CrossRef]
- 197. Feng, Y.; Tan, P.Y. Imperatives for Greening Cities: A Historical Perspective. In *Greening Cities. Advances in 21st Century Human* Settlements; Tan, P., Jim, C., Eds.; Springer: Singapore, 2017.
- 198. Smardon, R.C. Perception and aesthetics of the urban environment: Review of the role of vegetation. *Landsc. Urban Plan.* **1988**, 15, 85–106. [CrossRef]
- 199. Zhao, J.; Xu, W.; Li, R. Visual preference of trees: The effects of tree attributes and seasons. Urban For. Urban Green. 2017, 25, 19–25. [CrossRef]
- Hoyle, H.; Hitchmough, J.; Jorgensen, A. All about the 'wow factor'? The relationships between aesthetics, restorative effect and perceived biodiversity in designed urban planting. *Landsc. Urban Plan.* 2017, 164, 109–123. [CrossRef]
- Zehra, A.; Sahito, Z.A.; Tong, W.; Tang, L.; Hamid, Y.; Wang, Q.; Cao, X.; He, Z.; Khan, M.B.; Yang, X.; et al. Identification of high cadmium-accumulating oilseed sunflower (Helianthus annuus) cultivars for phytoremediation of an Oxisol and an Inceptisol. *Ecotoxicol. Environ. Saf.* 2020, 187, 109857. [CrossRef]
- Pavani, V. Sunflower, *Helianthus annus* L., Cut Flower Variety Trial. Master's Thesis, Western Kentucky University (USA), Bowling Green, KY, USA, 2005.
- 203. Newsholme, C. Willows: The Genus Salix; Timber Press, Inc.: Portland, OR, USA, 1992.
- 204. Brooker, I.; Kleinig, D. Eucalyptus: An Illustrated Guide to Identification; New Holland: Turin, Italy, 2013.
- Cundy, A.B.; Bardos, R.P.; Puschenreiter, M.; Mench, M.; Bert, V.; Friesl-Hanl, W.; Müller, I.; Li, X.N.; Weyens, N.; Witters, N.; et al. Brownfields to green fields: Realising wider benefits from practical contaminant phytomanagement strategies. *J. Environ. Manag.* 2016, 5, 67–77. [CrossRef]
- Dickinson, J.L.; Zuckerberg, B.; Bonter, D.N. Citizen Science as an Ecological Research Tool: Challenges and Benefits. *Annu. Rev. Ecol. Evol. Syst.* 2010, 41, 149–172. [CrossRef]

- 207. Mahajan, S.; Kumar, P.; Pinto, J.A.; Riccetti, A.; Schaaf, K.; Camprodon, G.; Smári, V.; Passani, A.; Forino, G. A citizen science approach for enhancing public understanding of air pollution. *Sustain. Cities Soc.* **2020**, *52*, 101800. [CrossRef]
- Marmiroli, N.; McCutcheon, S.C. Making Phytoremediation a Successful Technology. In *Phytoremediation*; Schnoor, J., Zehnder, A., McCutcheon, S., Schnoor, J., Eds.; John Wiley & Sons, Inc.: Hoboken, NJ, USA, 2003.
- Weir, E.; Doty, S. Social acceptability of phytoremediation: The role of risk and values. *Int. J. Phytoremediation* 2016, *18*, 1029–1036.
 [CrossRef]
- Curran, W.; Hamilton, T. Just green enough: Contesting environmental gentrification in Greenpoint, Brooklyn. *Local Environ.* 2012, 17, 1027–1042. [CrossRef]
- 211. Kabisch, N.; Frantzeskaki, N.; Pauleit, S.; Naumann, S.; Davis, M.; Artmann, M.; Haase, D.; Knapp, S.; Korn, H.; Stadler, J.; et al. Nature-based solutions to climate change mitigation and adaptation in urban areas: Perspectives on indicators, knowledge gaps, barriers, and opportunities for action. *Ecol. Soc.* 2016, 21, 39. [CrossRef]
- 212. Wells, N.M.; Evans, G.W. Nearby nature: A buffer of life stress among rural children. Environ. Behav. 2003, 35, 311–330. [CrossRef]
- Taylor, A.F.; Kuo, F.E.; Sullivan, W.C. Views of nature and self-discipline: Evidence from inner city children. J. Environ. Psychol. 2002, 22, 49–63. [CrossRef]
- Beyer, K.M.; Kaltenbach, A.; Szabo, A.; Bogar, S.; Nieto, F.J.; Malecki, K.M. Exposure to neighborhood green space and mental health: Evidence from the survey of the health of Wisconsin. *Int. J. Environ. Res. Public Health* 2014, 11, 3453–3472. [CrossRef] [PubMed]
- Cox, D.T.; Shanahan, D.F.; Hudson, H.L.; Fuller, R.A.; Anderson, K.; Hancock, S.; Gaston, K.J. Doses of nearby nature simultaneously associated with multiple health benefits. *Int. J. Environ. Res. Public Health* 2017, 14, 172. [CrossRef] [PubMed]
- Zhang, J.; Browning, M.; Liu, J.; Cheng, Y.; Zhao, B.; Dadvand, P. Is indoor and outdoor greenery associated with fewer depressive symptoms during COVID-19 lockdowns? A mechanistic study in Shanghai, China. *Build. Environ.* 2023, 227, 109799. [CrossRef]
- 217. Jimenez, M.P.; DeVille, N.V.; Elliott, E.G.; Schiff, J.E.; Wilt, G.E.; Hart, J.E.; James, P. Associations between Nature Exposure and Health: A Review of the Evidence. *Int. J. Environ. Res. Public Health* **2021**, *18*, 4790. [CrossRef]
- Lorenz, A.; Lüttkopf, D.; May, S.; Scheurer, S.; Vieths, S. The principle of homologous groups in regulatory affairs of allergen products–a proposal. *Int. Arch. Allergy Immunol.* 2009, 148, 1–17. [CrossRef]
- 219. Gibbs, J.E.M. Eucalyptus Pollen Allergy and Asthma in Children: A Cross-Sectional Study in South-East Queensland, Australia. *PLoS ONE* **2015**, *10*, e0126506. [CrossRef]
- Velasco, E.; Roth, M.; Norford, L.; Molina, L.T. Does urban vegetation enhance carbon sequestration? *Landsc. Urban Plan.* 2016, 148, 99–107. [CrossRef]

Disclaimer/Publisher's Note: The statements, opinions and data contained in all publications are solely those of the individual author(s) and contributor(s) and not of MDPI and/or the editor(s). MDPI and/or the editor(s) disclaim responsibility for any injury to people or property resulting from any ideas, methods, instructions or products referred to in the content.