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Intercropping Practices in Mediterranean Mandarin Orchards from an Environmental and Economic Perspective

Bernardo Martin-Gorriz ¹ , José A. Zabala ² , Virginia Sánchez-Navarro ³, Belén Gallego-Elvira ^{1,*} ,
Víctor Martínez-García ², Francisco Alcon ²  and José Francisco Maestre-Valero ¹ 

¹ Irrigation Design and Management Research Group, Department of Agricultural Engineering, Technical University of Cartagena, Paseo Alfonso XIII 48, 30203 Cartagena, Spain; b.martin@upct.es (B.M.-G.); josef.maestre@upct.es (J.F.M.-V.)

² Agricultural Economics Research Group, Department of Business Economics, Technical University of Cartagena, Paseo Alfonso XIII 48, 30203 Cartagena, Spain; joseangel.zabala@upct.es (J.A.Z.); victor.martinez@upct.es (V.M.-G.); francisco.alcon@upct.es (F.A.)

³ Sustainable Use, Management, and Reclamation of Soil and Water Research Group, Department of Agricultural Engineering, Technical University of Cartagena, Paseo Alfonso XIII 48, 30203 Cartagena, Spain; virginia.sanchez@upct.es

* Correspondence: belen.gallego@upct.es; Tel.: +34-968177741

Abstract: Crop diversification is becoming increasingly important for preserving soil and ecosystems' health and, subsequently, crop productivity and sustainability. Intercropping practices adopted in monocultural woody crops, with herbaceous crops covering the otherwise bare alleyways, foster ecological interactions and can provide both environmental and economic advantages. In this study, intercropping practices were implemented in a traditional mandarin orchard in south-eastern Spain, which was monitored for three years to assess their impact on the environmental footprint and profitability. The footprint was quantified with a cradle-to-gate life cycle assessment (LCA), while the costs and revenues assessment was based on materials, labor, and machinery used in the trial. The calculated LCA indicators evidenced that, although the cultivated surface area increases with the integration of the intercrops (fava bean, purslane, cowpea, and barley/vetch mix), this does not imply any additional detrimental effects (resource depletion, acidification, eutrophication, global warming). The economic analysis showed that while intercrops may involve additional production costs, the correct choice of intercrops, purslane, and fava bean, in this case, can reduce the market risks for farmers. Overall, this study shows that positive environmental and economic impacts are to be expected of co-integrated herbaceous crops within the same field as mandarin trees.

Keywords: crop diversification; agro-silvicultural systems; woody crops sustainability; life cycle assessment; farming market risks



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1. Introduction

In recent decades, agricultural practices in Europe have been focused on increasing yield, maximizing human labor efficiency, and reducing costs. Traditional cultivation systems have been displaced by intensive and specialized farming systems, which rely on external sources of agro-chemicals and energy [1]. Intensive monocrop systems have brought many negative impacts such as biodiversity loss, water pollution, high greenhouse-gas (GHG) emissions, soil degradation, and reductions in agroecosystem services [2]. As a consequence, in order to preserve farming sustainability, Europe must address the current environmental challenges, whilst ensuring that enough agricultural products of a good quality at a fair price are produced. In fact, climate change mitigation, the sustainable management of resources, enhancement of ecosystem services, and preservation of habitats and landscapes, along with supporting viable farm income and competitiveness, are key objectives for the upcoming common agricultural policy (CAP) 2023-27, in consonance with the European Green Deal and the Farm to Fork and Biodiversity strategies [3].

Many approaches have been proposed for improving the environmental and economic performance of agricultural systems. These include precision agriculture, organic production, mixed crop-livestock farming, conservation tillage, biological pest control, cultivar selection, and catch and cover crops [4–8]. Another alternative to intensive monoculture is crop diversification, in which different crops are combined in space and/or time, fostering ecological interactions and providing economic advantages [9–12]. Moreover, diverse agroecosystems have better endurance to the stress caused by extreme weather and pest outbreaks [13–15]. There are several crop diversification strategies, such as rotations, intercropping, and agroforestry practices, to improve or at least maintain environmental quality and agroecosystem services [9,16]. Among them, intercropping, which involves at least two different crops being grown at the same time in the same field, has been reported to provide ecological improvements, such as higher biodiversity, pest and weed control, soil conservation, and nutrient optimization. Intercropping also brings economic benefits, including improved profitability with fewer risks and lower agrochemical costs [17–19].

Within intercropping practices, agro-silvicultural systems, which combine the production of tree crops with herbaceous crops, have been reported to provide the following benefits: (i) enhancement of the fresh water supply; (ii) favoring the nutrient cycling and soil health; (iii) reduction in pests, weeds, and diseases; (iv) improvement of pollination; (v) reduction in GHG emissions and climate regulation support; (vi) carbon sequestration; (vii) promoting a safe habitat for associated biodiversity, and (viii) fostering recreational, aesthetic, and cultural heritage values [20–27]. However, despite the current land and ecosystem degradation and soil and water depletion and pollution [28,29] caused by intensive monocrops, intercropping practices have been barely adopted in the Mediterranean region. The reason for this is the misconception that intercrops may negatively affect the trees' production, as herbaceous crops in the alleyways are perceived as no more than competitors for water and nutrients [30,31]. In addition, many of the rewards of intercropping are not exchanged in the market and, hence, farmers may lack financial incentives and therefore opt to neglect voluntary green measures [12,32,33].

Fortunately, such perceptions seem to be changing nowadays, with farmers and society becoming increasingly aware that moving toward more environmentally friendly farming systems is strongly necessary. There is a rising social demand for more sustainable cropping systems in general, and for intercropping practices in particular [34], which is reflected in the high willingness to pay for both agroecological and socio-cultural benefits [12,35]. Moreover, the new design of environmental payments in the CAP 2023–27, based on performance rather than the former compliance-based approach, is expected to be more efficient and better-targeted at achieving the environmental, climate, and ecosystem welfare objectives [3,36].

In this context, the H2020 Diverfarming project is developing and testing different diversified cropping systems in six European pedoclimatic regions, under low-input practices, with the aim of increasing land productivity and crop quality, and reducing machinery, fertilizer, pesticide, energy, and water demands. In particular, intercropping in mandarin orchards has been assessed, given its socio-economic importance in the Mediterranean and the reported potential benefits of intercropping in citrus crops [25,37–39]. Mandarin demand is growing in global markets, about 60% in the last ten years, and Spain accounts for over 50% of the world's exports of fresh mandarins [40].

In acknowledgement of the fact that the adoption of environmentally friendly practices depends on both environmental and socio-economic potential benefits [41], the present study assesses the viability of the different intercropping practices trialed in mandarin orchards in south-eastern Spain from both perspectives, within the framework of Diverfarming. The mandarin orchards with herbaceous intercrops have already been reported to improve the annual soil carbon balance in the short-term [25], but potential detrimental environmental impacts have not been previously assessed. Therefore, on the one hand, this study quantifies the environmental footprint of the intercropping systems versus the traditional monocrop by conducting a Life Cycle Assessment (LCA), which

is a reference method to quantitatively evaluate the environmental impact of agri-food systems [42]. The novelty of this LCA is the fact that, to the best of our knowledge, although LCA has been used to estimate the footprint of mandarin and herbaceous crops separately, it has not been performed together nor with the same methodology. On the other hand, this study evaluates the potential profitability of intercropping systems, accounts for their sensitivity to the variation in the crop market value and irrigation water price, and assesses how intercropping could influence the market risks for farmers. Our results add weight to the growing body of evidence that the adoption of intercropping practices can potentially bring both environmental and economic benefits.

2. Materials and Methods

2.1. Experimental Farm

The trial was carried out on a commercial citrus farm of 2.3 ha located in the Region of Murcia in south-eastern Spain (37°57'31" N; 0°56'17" W, altitude 152 m a.s.l) from 2018 to 2021. The orchard contained 970 mandarin trees (*Citrus reticulata* Blanco var. Clemenvilla), which were planted in 2000, with a planting pattern of 6 × 4 m (between rows and trees, respectively) and was provided with a drip irrigation system. The farm soil is a Calcaric Regosol with a silt loam texture, bulk density of 1.20 g cm⁻³, pH of 7.56, total organic carbon of 3.12 g kg⁻¹, total nitrogen of 1.27 g kg⁻¹, phosphorus of 1.49 mg kg⁻¹, and potassium of 0.96 cmol kg⁻¹. The climate in the study area is Mediterranean semiarid with hot and dry summers, mild-temperature winters, an average rainfall of 350 mm/year (with most rainfall concentrated in autumn), and a potential evapotranspiration rate of 1300 mm year⁻¹.

2.2. Intercropping Practices and Irrigation Treatments

Three different cropping systems with specific irrigation treatments were evaluated (Figure 1):

- Control (CTL). Traditional mandarin monoculture, with bare alleys and irrigated to fully satisfy crop evapotranspiration.
- Diversification 1 (D1). Mandarin with intercrop rotation 1 and regulated deficit irrigation. In the alleyways, fava bean (*Vicia faba* L.) was grown from September to December–February, and a mix (1:3 ratio) of barley/vetch (*Hordeum vulgare* L./*Vicia sativa* L.) from January–February to June each year.
- Diversification 2 (D2). Mandarin with intercrop rotation 2 and regulated deficit irrigation. In the alleyways, fava bean (*Vicia faba* L.) was grown from September 2018 to January 2019, purslane (*Portulaca oleracea*) from May to July 2019, and cowpea (*Vigna unguiculata*) from June to September 2020.

For each intercropping practice, there were three replicates of 12 m × 24 m, which included three rows of six trees. Due to the semiarid conditions of the study area, water is very limited and drip irrigation and regulated deficit irrigation are common and well-known practices [43,44]. The irrigation needs of the CTL were estimated from the calculation of crop evapotranspiration (ET_c) according to the FAO methodology [45], with the crop coefficients provided by the Agricultural Information System of Murcia (<http://siam.imida.es>, accessed on 24 March 2022) for this area, and considering the amounts of daily rainfall recorded in the closest station (MU52, details provided in Table S1). Meteorological data of temperature, relative humidity, radiation, wind, and precipitation were also obtained from the automatic weather station MU52 (Table S1). Mandarin plots under regulated deficit irrigation (D1 and D2) were irrigated as per the CTL, except for the post-harvest period when they were irrigated at 50% CTL, in order to keep the stem water potential above −2 MPa, to preserve water while minimally impacting the production. The water saved was then used to irrigate the alleyways. Tillage up to 30 cm was performed to prepare the soil before manually sowing the intercrops, and after harvest to incorporate the crop residues into the soil. For D1, barley/vetch seeds were sown covering the entire alley surface and fava bean seeds in three rows in each alley with a planting pattern of

1 × 0.4 m. For D2, fava bean seeds were sown as in D1, purslane plants were planted in three rows in each alley with a planting pattern of 1 × 1 m, and cowpea seeds were sown in three rows in each alley with a planting pattern of 1 × 0.2 m. Mandarin, fava bean, purslane, and cowpea were harvested for food, whilst barley/vetch was for feed. Sowing and harvesting dates and the water and fertilizers applied are detailed in Tables S2 and S3 (Supplementary Materials).

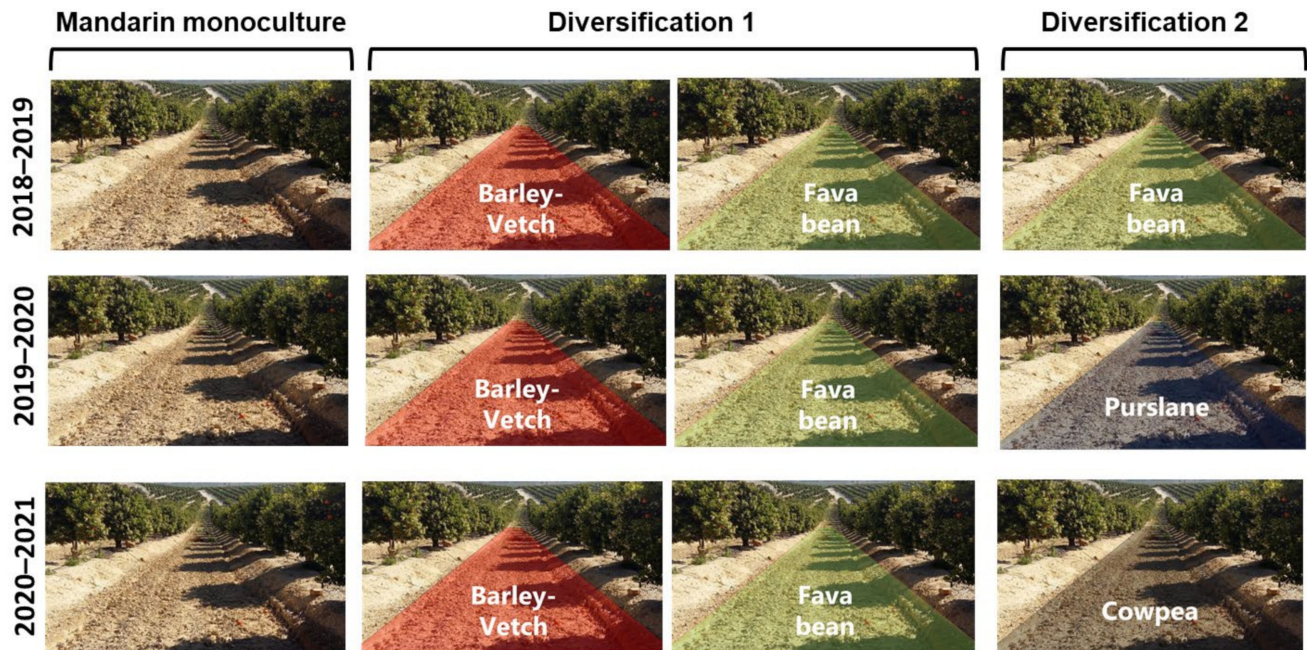


Figure 1. Schematic representation of the three intercropping practices: Mandarin monoculture (CTL), Diversification 1 (D1), and Diversification 2 (D2).

2.3. Life Cycle Assessment Methodology

The present LCA follows the guidelines and specific requirements of the International Organization for Standardization 14040/44 standard series (ISO 14040 and ISO 14044 [46,47]).

2.3.1. Goal and Scope Definition of the LCA Study

The aim of this LCA was to evaluate and compare the environmental impacts of the intercropping practices for mandarin and intercrop production. The scope of the LCA study was the ‘cradle-to-gate’ production of mandarin with and without the intercrops. All the input and output flows of materials and energy up to the farm gate were considered. The specific data regarding the three production systems studied (CTL, D1, and D2) are detailed in Tables S3 and S4.

2.3.2. Functional Units

Functional units (FUs) provide the reference to which all data in the system are normalized for the interpretation of the environmental results and the comparison of the performance of the production systems. The following FUs, commonly used for agricultural studies, were selected:

- 1 hectare of cultivated area (1 ha), which enabled the comparison of mandarin monoculture (CTL) and mandarin with intercrops (D1 and D2).
- 1 kg of harvest (mandarin, fava-bean, purslane, cowpea, or barley/vetch), to compare the performance of the same crops in different intercropping strategies.

2.3.3. System Boundaries

All the processes of the production chain of mandarin fruit and intercrops from the extraction of raw materials to the farm gate are within the system boundaries, as shown in Figure 2. The background system comprises the processes that take place before the farm stage whilst all the field operations are included in the foreground system (Table S3).

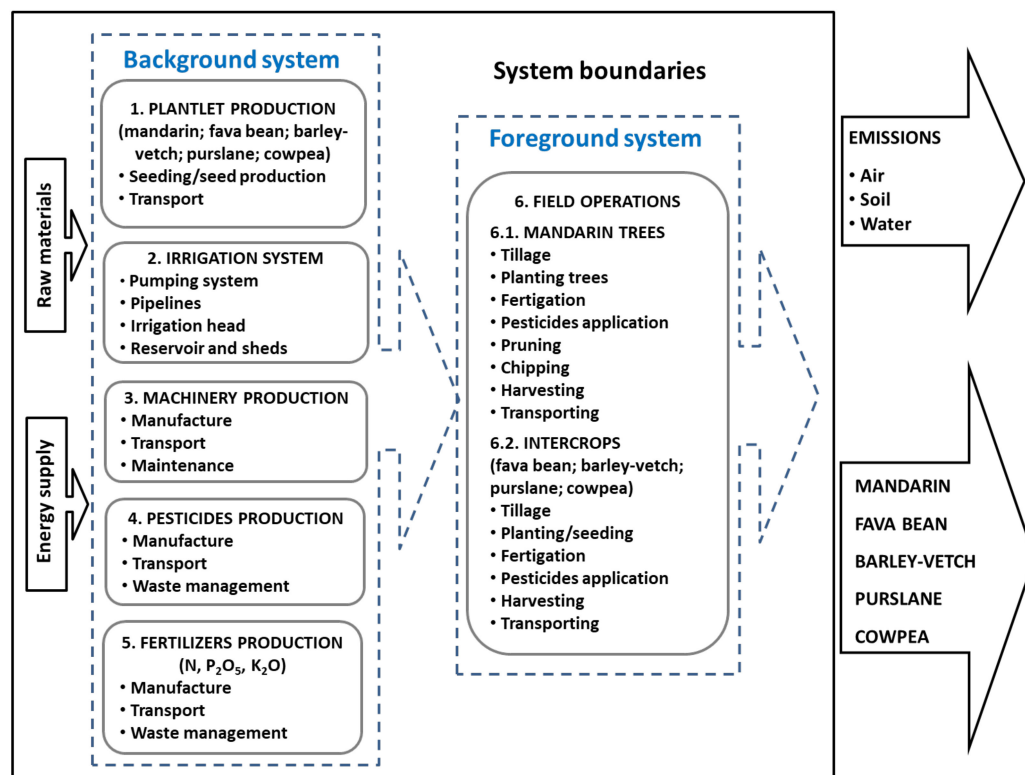


Figure 2. System boundaries for 'cradle-to-gate' production of mandarin and intercrops.

The agricultural subsystems in the background (1–5) and foreground (6) are the following:

- (1) Plantlet production. Farmers in the study area do not graft mandarin trees nor do they produce seeds of the intercrops considered. Therefore, the energy and materials used at the nurseries and for the transport to the fields were accounted for, as in [48–50]. In the mandarin nursery, rootstocks and vegetative buds are grown and grafting takes place. For this activity, agricultural machinery, mineral and organic fertilizers, pesticides, and ground cultivation are required. In the case of the purslane plantlets, they are planted manually in seedling growing trays and placed in greenhouses for about 30 days, prior to being moved to the field. In the case of fava-beans, cowpea, and barley/vetch, seeds from local storehouses were used.
- (2) Irrigation subsystem. The following components were considered for the mandarin production: (i) the manufacture and transport of the irrigation infrastructure including head, filters, fertilizer tanks, electro-valves, irrigation programmer, PVC, and PE pipes; (ii) the construction of the shelters for the farm machinery and the fertigation head and the on-farm irrigation reservoir; (iii) the water and energy required for irrigation; and (iv) the extraction, production, and transport of electricity. The same components (i–iv) were considered for the intercrops along with the additional PE pipes and electricity required for the irrigation of the alleyways.
- (3) Production of machinery. This included the manufacture, transport, maintenance, repair, and waste management of the machinery used for field operations of each crop. All the crops required a tractor, tillage implements, and transport trailers. The

mandarin crop additionally required an air-blast sprayer and a boom sprayer for pest control and a chipper for pruning.

- (4) Production of pesticides. This included the transport of primary and secondary materials to the production plants, the synthesis of the chemical components, and the waste management and disposal.
- (5) Production of synthetic fertilizers. This covered the manufacturing of nitrogen (N), phosphate (P_2O_5), and potash (K_2O), as well as the packaging and transport of materials to the production plants.
- (6) Field operations. These included tillage, planting, fertigation, harvesting, and transporting for all crops. For the case of mandarin, pruning and chipping were also accounted for. Pesticides were only applied in fava bean and mandarin.

2.3.4. Life Cycle Impact Assessment (LCIA) Methodology and Types of Impacts

The impact categories assessed were five mid-point impact categories defined according to the CML 2001 (April 2013 version) impact assessment method [51]: abiotic depletion elements (ADe, kg Sb_{eq}); abiotic depletion fossil fuels (ADf, MJ); acidification (AC, kg SO₂_{eq}); eutrophication (EU, kg PO₄_{eq}); and global warming (GW, kg CO₂_{eq}). Additionally, the use of water was evaluated, considering its relevance in the water-stressed study region.

The software used for the LCA was SimaPro version 9.1.0.8 (Pre Sustainability, 2020). The emission factors for the operations were taken from the ecoinvent database (v.3.7.1), except for the on-farm emissions of fertilizers, which were taken from the following models, as they were considered more realistic and representative: (i) N-emissions to air: NH₃ emissions = 3% total N_{applied} [52], N₂O emissions = 1.25% total N_{applied} [52], and NO_x emissions = 10% total N₂O emissions [53]; and (ii) NO₃[−] emissions to water: 5% total N_{applied} [54].

2.3.5. Data Quality

The primary data used in this LCA, collected in the experimental plots, were the type and amount of fertilizers, agrochemicals, fuel, electricity, water, irrigation materials, and infrastructure and machinery used for crop production. In addition, secondary data from the ecoinvent database were used to calculate the materials and sources of energy used. The quality of the input data was evaluated following the ILCD requirements [55]. All primary data used for the study were classified as ‘high-quality’ (Table S5) and the secondary data from ecoinvent as ‘basic-quality’ (Table S6).

2.3.6. Assumptions

For the transport of raw materials (pesticides, fertilizers, plantlets, etc.) from the local storehouse to the farm, typical standard cargo transport (EURO 6, ecoinvent database) was assumed to be used. The average transport distances were estimated to be 50 km for the raw materials, as well as for the disposal of waste and the transport of pesticide and herbicide containers to the recycling center.

2.3.7. Life Cycle Inventory Analysis (LCI)

The data collection covered three cultivation seasons between 2018 and 2021. All farming activities concerning the three systems studied (CTL, D1, and D2) were inventoried for each experimental plot in order to build the database. The values used for the LCA from each crop and year are shown in Table S3 and the detailed inventory for the infrastructure can be found in Table S4.

2.3.8. Statistical Analysis

For the comparison of environmental impacts between CTL, D1, and D2, a one-way analysis of variance (ANOVA) was carried out with the statistical package Statgraphics Plus V5.1. The results were used to test the hypothesis of equal means of the environmental impacts for the different systems at a significance level of $p < 0.05$.

2.3.9. Sensitivity Analysis

A sensitivity analysis was performed to evaluate the variation in the impact indicators with different fertilizer doses in the mandarin crop, which was identified as one of the most influential variables for this system. This is particularly relevant due to the identified tendency toward excessive crop fertilization [56] and as the European Union has set a goal to reduce fertilizer use by 20% and fertilizer diffuse pollution by 50% by 2030 [57]. We considered two different scenarios for this analysis with a 10 and 20% reduction in fertilizer doses.

2.4. Economic Assessment Methodology

2.4.1. Costs and Revenues

Costs and revenues, following the definition established by Fernández et al. [58], were calculated for each system (CTL, D1, and D2) on the basis of raw material, labor, and machinery used during the experimental period.

Revenues were calculated as the market value of the yield, considering the official crop prices published by the Government of the Region of Murcia [59] and the Spanish Ministry of Agriculture, Fisheries and Food [60]. The 10-year average prices for the period 2011–2020, deflated with the national averaged consumer price index, were taken to avoid the effect of price volatility and inflation. The average price for mandarin was EUR 0.33/kg, and for the intercrops, they were EUR 0.13, 1.50, 1.00, and 1.93/kg for vetch/barley, fava bean, purslane, and cowpea, respectively.

The total farming costs included the following costs:

- Fixed costs (FC), invariant with respect to the yield, included yearly costs incurred due to assets depreciation, the start-up costs associated with the initial nonproductive period of the orchard (2 years), and the insurance and taxes paid by the farm. The annual assets depreciation was obtained from the allocation of the initial investment costs over the assets' lifespan. The investment costs were estimated to be EUR 120,704 for a 10 ha farm with a water reservoir with capacity for 25,000 m³, of which EUR 4185/ha corresponded to soil preparation and plantation and EUR 4290/ha to the drip irrigation system [61]. Extra drip irrigation costs in the case of diversifications D1 and D2 were EUR 63/ha. The lifespan of the orchard was assumed to be 25 years, whilst a specific lifespan was applied to each individual asset.
- Variable costs (VC) were the costs for each growing season, which included fertilizers, water, labor, fuel, and energy, according to the farm operational activities for each system and required inputs (Table S3).
- Opportunity costs (OC) were defined as the foregone benefits due to the implementation of the systems over other feasible economic alternatives. They mainly included renting the land, priced according to the officially published regional data [59], and the interest from the fixed and variable capital, which was assumed to be 3%.

Finally, the profit was calculated per hectare of cultivated area as the difference between revenues and FC, VC, and OC.

2.4.2. Statistical Analysis

As in the case of the LCA, an ANOVA was carried out to check for statistically significant differences in revenues, costs, and profit among the different systems. When significant differences among cropping practices were found, a Tukey's post hoc test was used to determine which practices were different at a significance level of $p < 0.05$.

2.4.3. Sensitivity Analysis

In the study area, important fluctuations have been reported in the financial profitability of citrus orchards driven by changes in the market price of the fruit and the price of irrigation water [62,63]. Therefore, a sensitivity analysis was carried out to assess how the variation in the price of mandarin and irrigation water could impact profitability, and

whether or not the crop diversifications enabled the market risks for farmers to be reduced. Apart from exploring D1 and D2, an alternative diversification was proposed (Dp) with purslane and fava bean as intercrops, as they were observed to be the most profitable intercrops during the trial.

3. Results

3.1. Environmental Assessment

3.1.1. Impact Indicators

The values of the environmental impact indicators per hectare of cultivated area for the mandarin monocrop CTL, and the diversifications D1 and D2 are shown in Figure 3 (and Table S7). The disaggregated values for each crop in D1 and D2 are also given in Figure 3 (bars with diagonal line pattern). The values of the environmental impact indicators per kg of product harvested are presented in Table S8. The most remarkable findings were:

- No significant differences between the CTL and diversifications D1 and D2 were observed in any of the five mid-point impact categories analyzed (Table S7). Only water use was observed to gradually decrease from the CTL to D1 to D2 (3820, 3351, and 3007 m³/ha, respectively), although the differences were not statistically significant, due to the high variability in the data collected in D2. Note that, in D2, the higher variability is related to fact that there was a different intercrop each year, unlike in the CTL and D1.
- Looking at the environmental footprint of the mandarin crop alone, it may seem from the data per ha shown in Figure 3 that the footprint was higher in the CTL. However, this was just an artifact caused by the fact that the production of mandarin was lower when the intercrops were introduced (Figure S1). In fact, the footprint per kg of harvested fruit in the CTL (GW: 0.35 kg CO₂ eq/kg) was very similar to both D1 (GW: 0.38 kg CO₂ eq/kg) and D2 (GW: 0.38 kg CO₂ eq/kg) (Table S8).
- Barley-vetch and fava bean in D1 showed similar levels of impacts in all categories and a significantly lower impact than the mandarin crop with deficit irrigation.
- In contrast with D1, the environmental impacts of intercrops in D2 showed a progressive increase in environmental impacts in the following order: fava bean, purslane, and cowpea, due to the corresponding higher fertilizer needs.

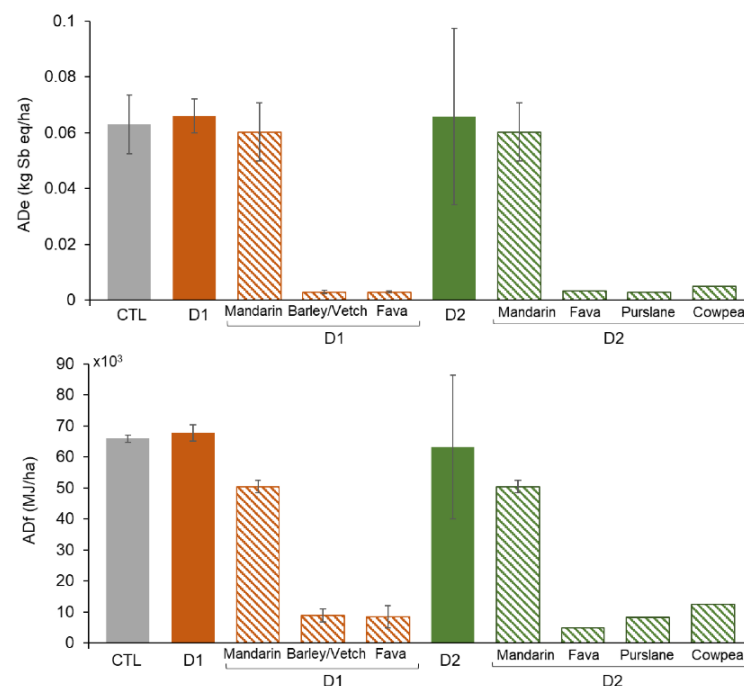


Figure 3. Cont.

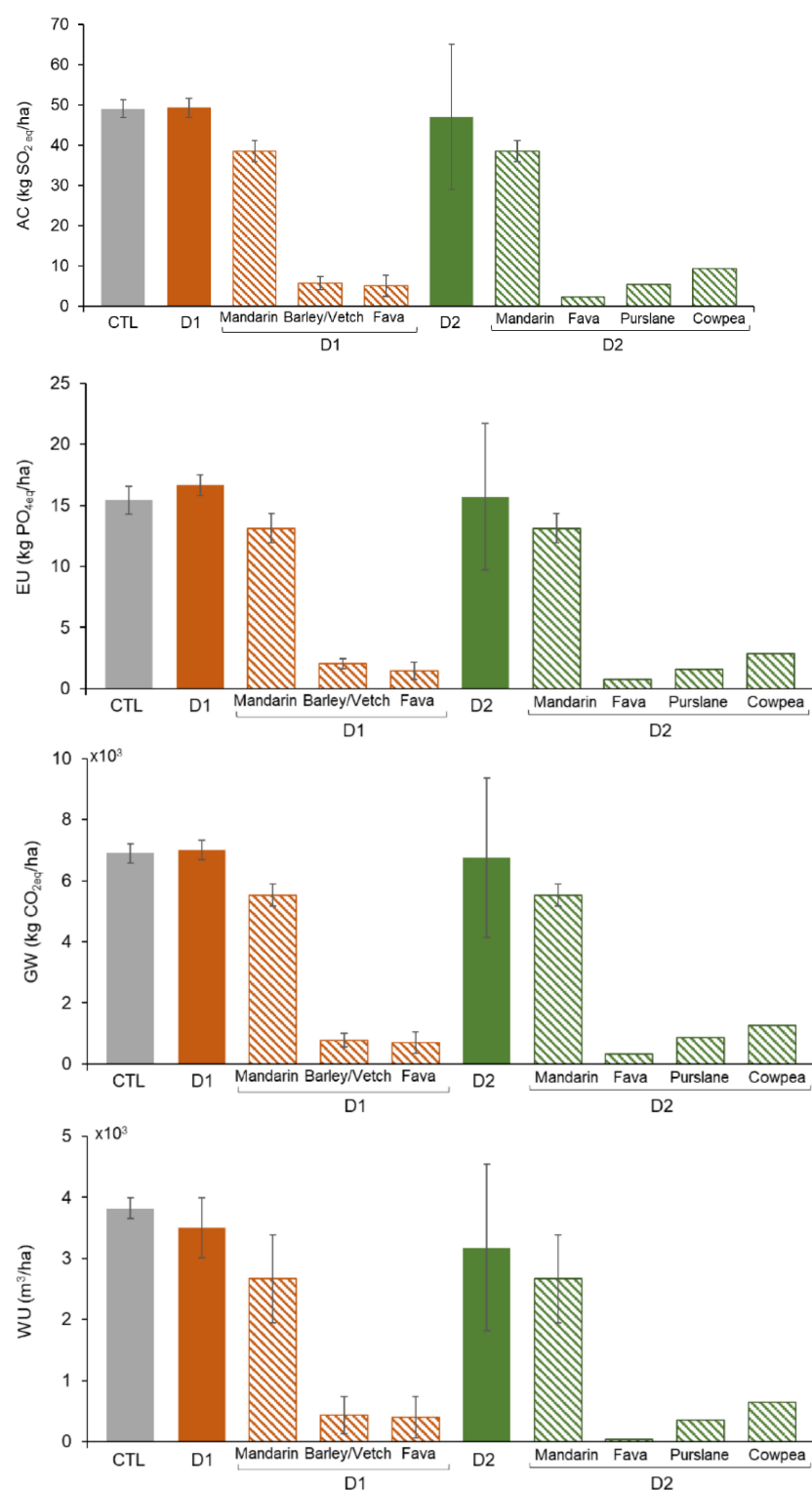


Figure 3. Mean value and standard deviation of the environmental impact indicators per unit of area. The bars with solid patterns present the total values per hectare of cultivated area for mandarin monoculture (CTL), diversification 1 (D1), and diversification 2 (D2). The bars with diagonal line patterns correspond to the disaggregated values for each crop in D1 and D2. Impact categories: Abiotic depletion elements (ADE); Abiotic depletion fossil fuels (ADf); Acidification (AC); Eutrophication (EU); Global warming (GW); and Water use (WU).

The values of the environmental impacts separated by subsystems were also analyzed (Figure S2). For all categories and for all systems studied (CTL, D1, D2), the field work, the production of fertilizers, and the irrigation subsystem dominated (more than 80% together) the footprint in the ADf, AC, EU, and GW categories. A higher weight of the irrigation subsystem in ADf, AC, and GW was observed, which was mainly associated with the fact that the electricity mix in Spain still relies largely on the use of fossil fuels. The production of fertilizers was the main contributor (34%) to the EU impact in CTL and D2, while in D1, it was the field work (31%) because of the higher diesel fuel consumption used in tillage practices for the two annual crops (barley-vetch and fava bean).

3.1.2. Sensitivity Analysis

The potential impact of reducing fertilizer doses by 10 and 20% was studied in this analysis. A 10% (20%) reduction in fertilizers decreased the environmental impact of all the studied systems by 2 to 4% (4 to 9%) depending on the impact category (Table S9). Overall, fertilizers contribute to all the impact categories, as they consume resources and emit pollutants to the air, water, and soils during their manufacture, transport, and application.

3.2. Economic Assessment

3.2.1. Costs and Revenues

Figure 4 summarizes the costs, revenues, and profit per hectare of cultivated area for CTL, D1, and D2. The values for each of the crops in D1 and D2 are also shown in Figure 4. The disaggregation of fixed, variable, and opportunity costs along with the values of gross and net margins are provided in Table S10. The most relevant facts from the economic assessment were:

- The total costs of diversifications D1 and D2 were very similar and higher than that of the CTL. This was mainly caused by the higher labor needs for harvesting and planting/sowing of intercrops, which caused a significant increase (four-fold) in labor costs with regard to the monocrop. The rest of the variable costs were similar in all three systems (Table S10). The presence of intercrops barely affected the cost of the use of machinery and raw materials, and the implementation of deficit irrigation in D1 and D2 served to compensate the irrigation over-costs of the intercrops.
- The average revenues of the diversifications were lower than those of the monocrop, but considering the high variability of the harvests in the experimental period, the differences were not statistically significant. It is worth noting that in D2, purslane provided higher revenues than the mandarin itself and had lower variability. Fava bean also provided reasonable revenues and moderate variability in the diversifications. Consequently, these intercrops could partially compensate for the reduction in productivity of the main crop. Furthermore, they may help to reduce revenue variability, thereby reducing the economic risk for farmers.
- Overall, albeit not statistically significant, profits were lower in the studied diversifications D1 and D2 than in the mandarin monocrop. However, some intercrops (purslane and fava bean) were observed to be profitable and presented a potential opportunity to increase revenues and to control the interannual volatility of revenue, as presented in the next section.

3.2.2. Sensitivity Analysis

The existence of different co-integrated crops within the same field reduces the economic dependence on only one crop and may enable the market risk for farmers to be reduced. Figure 5 shows the results of the sensitivity analysis, in which we explored whether the diversifications helped to reduce the market risk when there were fluctuations in the price of the main crop (mandarin) and of irrigation water. As stated in the methods, the proposed diversification Dp, with purslane and fava bean as intercrops, was included in the analysis, given the observed promising profitability of these intercrops.

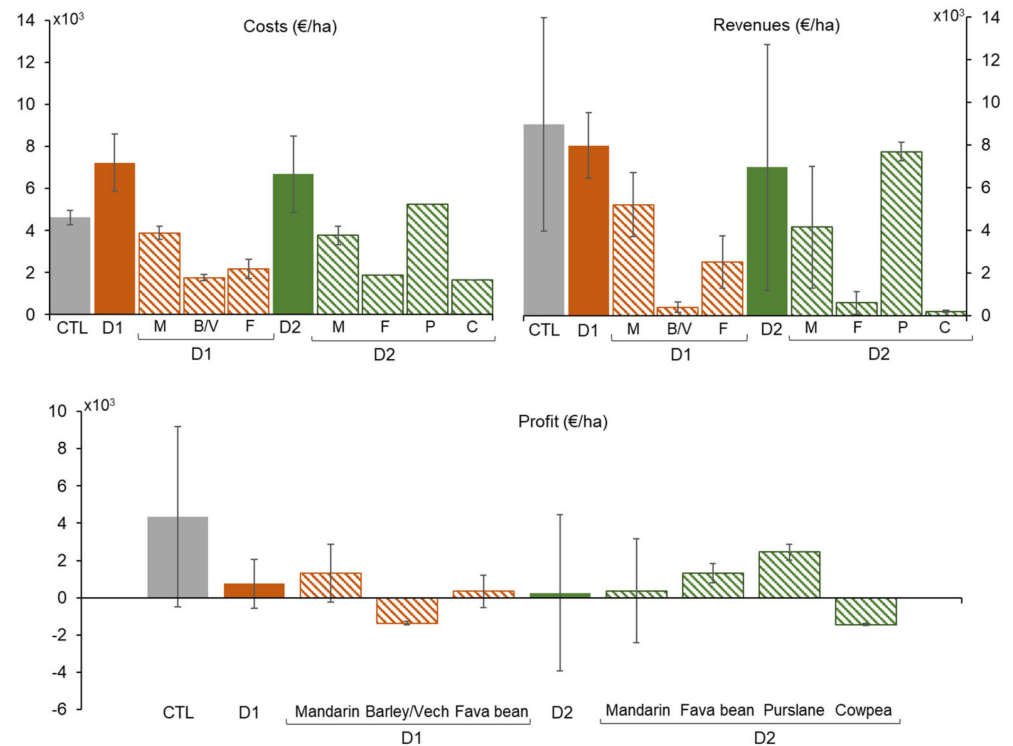


Figure 4. Mean value and standard deviation of the annual costs, revenues, and profit per unit of area. The bars with solid patterns show the total values per hectare of cultivated area for mandarin (CTL), diversification D1 (D1), and diversification D2 (D2). The bars with diagonal line patterns correspond to the disaggregated values for each crop in D1 and D2.

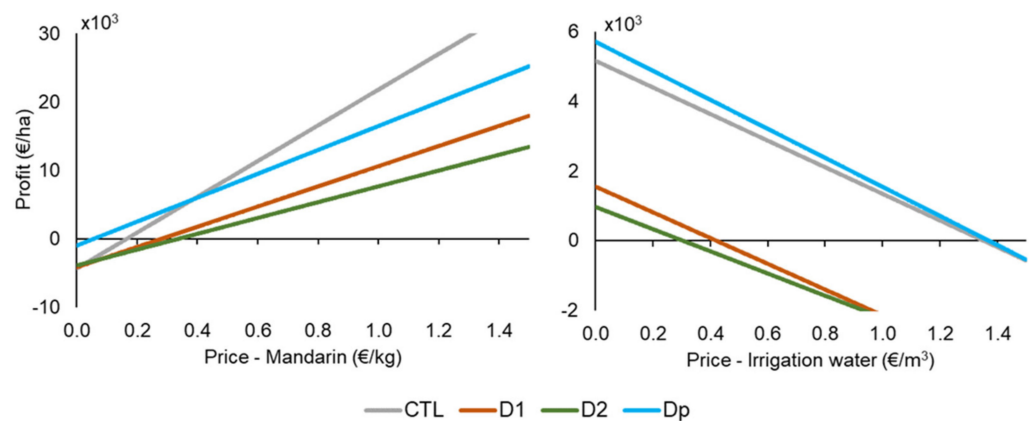


Figure 5. Sensitivity analysis of the annual profit (EUR/ha) to the variation in the price of mandarin and irrigation water, in mandarin monoculture (CTL), diversification D1 (D1), diversification D2 (D2), and diversification Dp (purslane and fava bean as intercrops).

The analysis of profit when the market price of mandarin varied from EUR 0 to 1.5/kg (Figure 5) revealed that D1 and D2 were unable to reduce the market risk. Conversely, Dp lowered the risk and outperformed the CTL, reducing the break-even point from EUR 0.17/kg in CTL to EUR 0.06/kg in Dp. However, it is important to note that the CTL remained more profitable for mandarin market values over EUR 0.4/kg.

The results presented in Figure 5 also show that diversifications D1 and D2 were substantially more sensitive to changes in the irrigation water price than the CTL. However, Dp showed a very similar pattern to the CTL and the same break-even point as the CTL at EUR 1.36/m³. Considering that the latter figure is about five times higher than the current

water price paid by farmers in the area, Dp can help to tolerate substantial fluctuations in the water price.

4. Discussion

In the present study, we assessed if the studied intercropping practices, which can potentially bring a wide array of benefits to the soils and ecosystems [25,64,65], may also imply detrimental effects on the environment (namely: resource depletion, acidification, eutrophication, global warming). Our results showed that there were no significant differences between the CTL and diversifications D1 and D2 in any environmental impact indicator per land area unit or mass unit. Therefore, the intercropping practices were not found to cause additional pollution or other detrimental effects on the environment. It is important to highlight that the present LCA only looked at indicators of potential harm to the environment. The benefits of intercropping in woody crops with annual crops, such as an increase in soil organic carbon and nitrogen content [18] or reduced soil erosion and runoff [25], were not accounted for in the LCA. Therefore, considering all the reported potential benefits in previous studies, the diversified systems studied here are expected to be more environmentally friendly than traditional orchards.

The environmental impact indicators of mandarin crop and intercrops of this study were compared to previous similar studies. When carrying out comparisons, it is important to bear in mind that many factors affect the LCA results, including, but not limited to, the type and quantities of inputs (water and energy for irrigation, fertilizer doses, or diesel consumption for field operations). In conventional orange production, similar values of GW have been reported by De Luca et al. [66] in Italy and Aguilera et al. [67] in Spain (6800 and 6520 kg CO₂ eq/ha, respectively), whilst [68] estimated values 36% higher (9830 kg CO₂ eq/ha). The latter could be attributed to their more intensive use of pesticides and manure. It should be noted that, as the trials of this study were conducted within the Diverfarming project, lower-input practices were used, such as deficit irrigation and controlled fertilization. In fact, the results of the present study compared to a previous footprint assessment of citrus in the same area [69] show that a reduction in inputs can effectively lead to a significant decrease not only in GW (4.46×10^{-1} vs. 3.5×10^{-1} kg CO₂ eq/kg FU) but in all the impact indicators: ADe (3.39×10^{-6} vs. 3.26×10^{-6} kg Sb_{eq}/kg FU), ADf (5.33 vs. 3.34 MJ/kg FU), AC (2.61×10^{-3} vs. 2.46×10^{-3} kg SO₂ eq/kg FU), and EU (1.13×10^{-3} vs. 7.69×10^{-4} kg PO₄³⁻ eq/kg FU). Lower footprints have also been reported in other studies of orange production, i.e., [70], who estimated AP to be half that of our study, but for very different reasons. These lower values could be explained by the reference period used (50 years) and the fact that both productive and unproductive periods were included. Regarding the footprint of the intercrops, when comparing the environmental impact of legumes (fava bean and cowpea) versus the literature, our GW values were in the range of the green beans category (0.24–1.55 kg CO₂ eq/kg FU) of previous comparable studies [71].

In this study, we also proposed to mitigate the environmental footprint by reducing the use of synthetic fertilizers as previously suggested by other studies [23]. Our results are well aligned with prior studies reporting that the use of fertilizers was critical from an environmental perspective, with a reduction of 30% in fertilizers enabling decreases of up to 15% in the EU, 10% in the GW, and 12% in the AC [69,72]. In addition, the replacement of synthetic fertilizers by organic fertilizers in citrus, apart from reducing the footprint, is key to maintaining healthy levels of soil organic carbon [18].

From the economic perspective, the results show that diversifications may imply additional production costs, which, in our case, were mainly related to the higher labor demand, compared with monoculture. The over-cost derived from intercrops may not be recovered by the farmers in the short term, but it is likely to be outbalanced in the long term [19], due to the many unaccounted potential benefits, mainly derived from the expected improvement in the environmental and soil quality conditions [73,74]. In addition, the present study has shown how the correct choice of intercrops (Dp) can

reduce the market risk. In this line, Martins et al. [75] determined that the use of cowpea as an alley crop provides a good environmental and economic performance for citrus orchards in Northeastern Brazil. However, their results contrast with ours in terms of the potential economic benefits. This difference may arise from the combination of relatively low farming labor costs together with a greater demand and prices for cowpea in Brazil. In consonance with [75], Yan et al. [76] also showed that the intercropping of ferns and mandarin trees provides positive economic returns compared with a mandarin monocrop in Hunan Province (China). Conversely, Srivastava et al. [77] revealed that only legume-based intercropped mandarin orchards provided greater yields, in comparison with other types of intercropping practices. Therefore, the optimal intercropping practices are to be determined, depending on regional crop market values and conditions.

Crop diversification practices also bring benefits to the society, associated with ecosystem services provided by agroecosystems [27]. Farms with high diversity, apart from providing an opportunity for farmers to diversify their incomes, have an added aesthetic value that can attract tourism to rural areas [78]. There is in fact a reported strong social demand for more sustainable and diverse cropping systems. Alcon et al. [12] and Latvala et al. [35] identified strong preferences for intercropping systems in the Mediterranean region and Finland, respectively. From their surveys, it was clear that consumers are willing to pay for nonmarket values such as biodiversity, conservation of soils and ecosystems, and mitigation of pollution. They estimated the value of the social demand at EUR 900–1400 and 245/ha year, respectively, which, in some cases, can even be comparable to the market value of agricultural production.

5. Concluding Remarks and Future Research Directions

This study compared the environmental footprint and economic performance of traditional Mediterranean mandarin monoculture versus mandarin crop with intercrops in the alleyways and controlled deficit irrigation. The calculated LCA indicators of potential harm to the environment showed that the integration of the intercrops (fava bean, purslane, cowpea, and barley/vetch mix) does not imply any additional pollution compared to the monocrop. The economic assessment evidenced that intercropping may involve additional production costs, mostly related to higher labor demand compared with monoculture. However, the correct choice of intercropping practices can bring economic advantages. In fact, our results showed that mandarin crop with purslane and fava bean as intercrops could be profitable and reduce the farmer's risk to crop price volatility. Consequently, considering all the potential environmental and economic benefits of the intercropping practices reported in the literature, the studied diversified systems are recommended to move toward more sustainable but still profitable agricultural systems.

Further and longer-term research is still needed to identify the optimal intercropping practices for woody crops, which depend on regional cropping conditions and possibilities, as well as the ever-changing crop markets. On the one hand, future research can help to curb labor costs by adapting farm machinery to the specific intercropping tasks, together with a coordinated design of the alley crop plantations. On the other hand, the adoption of intercropping practices by farmers needs to be encouraged by agricultural stakeholders and alongside the food value chain. The valorization of greener agricultural products by the consumer and the design of new farm business models based on intercropping practices supported by funding schemes (e.g., direct subsidies to farmers who implement intercropping) are key aspects to boost the adoption of these practices.

Supplementary Materials: The following supporting information can be downloaded at: <https://www.mdpi.com/article/10.3390/agriculture12050574/s1>, Figure S1: Mandarin fruit yield and standard error; Figure S2: Subsystem contributions to selected impact categories for mandarin monoculture (CTL), diversification 1 (D1), and diversification 2 (D2); Table S1: Monthly meteorological data during the study period from MU52 weather station (37°58'39"; 0°59'1"; 125 m a.s.l.) of the Agricultural Information System of Murcia (<http://siam.imida.es>, accessed on 24 March 2022). Table S2: Sowing and harvesting dates in mandarin monoculture (CTL) and diversification treatments

(D1 and D2) during three crop cycles; Table S3: Primary input/output data for mandarin monoculture (CTL) and mandarin with intercrops D1 and D2 during three crop cycles; Table S4: Primary input data used for the irrigation subsystem for mandarin monoculture (CTL) and mandarin with intercrops D1 and D2; Table S5: Quality evaluation of the primary input data. Self-evaluation against the ILCD data quality indicators of EC-JRC; Table S6: Quality evaluation of the input data from ecoinvent database. Self-evaluation against the ILCD data quality indicators of EC-JRC; Table S7: Mean values of impact categories per unit of area (ha) in mandarin monoculture (CTL), diversification 1 (D1), and diversification 2 (D2); Table S8: Mean value and standard deviation for the potential environmental impacts per kg of product harvested in mandarin monoculture (CTL), diversification 1 (D1), and diversification 2 (D2) by crop; Table S9: Contributions to reduction in fertilizer doses versus the reference situation by impact category in mandarin monoculture (CTL), diversification 1 (D1), and diversification 2 (D2); Table S10: Average 2018–20 values for costs, revenues, and gross margins for the mandarin monoculture (CTL), diversification 1 (D1), and diversification 2 (D2).

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References

1. Bommarco, R.; Kleijn, D.; Potts, S.G. Ecological intensification: Harnessing ecosystem services for food security. *Trends Ecol. Evol.* **2013**, *28*, 230–238. [\[CrossRef\]](#)
2. Wezel, A.; Goris, M.; Bruil, J.; Felix, G.F.; Peeters, A.; Barberi, P.; Bellon, S.; Migliorini, P. Challenges and Action Points to Amplify Agroecology in Europe. *Sustainability* **2018**, *10*, 1598. [\[CrossRef\]](#)
3. European Council. Council Adopts Fairer, Greener and More Performance-Based Farming Policy for 2023–2027. 2021. Available online: <https://www.consilium.europa.eu/en/press/press-releases/2021/12/02/council-adopts-fairer-greener-and-more-performance-based-farming-policy-for-2023-2027/> (accessed on 24 March 2022).
4. Gebbers, R.; Adamchuk, V.I. Precision Agriculture and Food Security. *Science* **2010**, *327*, 828–831. [\[CrossRef\]](#)
5. Eyhorn, F.; Muller, A.; Reganold, J.P.; Frison, E.; Herren, H.R.; Luttikholt, L.; Mueller, A.; Sanders, J.; Scialabba, N.E.-H.; Seufert, V.; et al. Sustainability in global agriculture driven by organic farming. *Nat. Sustain.* **2019**, *2*, 253–255. [\[CrossRef\]](#)
6. Schut, A.G.T.; Coolidge, E.C.; Moraine, M.; Van de Ven, G.W.J.; Jones, D.L.; Chadwick, D.R. Reintegration of crop-livestock systems in Europe: An overview. *Front. Agric. Sci. Eng.* **2021**, *8*, 111–129. [\[CrossRef\]](#)
7. Liu, H.; Meng, J.; Bo, W.; Cheng, D.; Li, Y.; Guo, L.; Li, C.; Zheng, Y.; Liu, M.; Ning, T.; et al. Biodiversity management of organic farming enhances agricultural sustainability. *Sci. Rep.* **2016**, *6*, 23816. [\[CrossRef\]](#) [\[PubMed\]](#)
8. Sanden, T.; Spiegel, H.; Stueger, H.P.; Schlatter, N.; Haslmayr, H.P.; Zavattaro, L.; Grignani, C.; Bechini, L.; D’Hose, T.; Molendijk, L.; et al. European long-term field experiments: Knowledge gained about alternative management practices. *Soil Use Manag.* **2018**, *34*, 167–176. [\[CrossRef\]](#)
9. Francaviglia, R.; Alvaro-Fuentes, J.; Di Bene, C.; Gai, L.; Regina, K.; Turtola, E. Diversified Arable Cropping Systems and Management Schemes in Selected European Regions Have Positive Effects on Soil Organic Carbon Content. *Agriculture* **2019**, *9*, 261. [\[CrossRef\]](#)

10. Van Asten, P.J.A.; Wairegi, L.W.I.; Mukasa, D.; Uringi, N.O. Agronomic and economic benefits of coffee-banana intercropping in Uganda's smallholder farming systems. *Agric. Syst.* **2011**, *104*, 326–334. [[CrossRef](#)]
11. Wood, S.A.; Karp, D.S.; DeClerck, F.; Kremen, C.; Naeem, S.; Palm, C.A. Functional traits in agriculture: Agrobiodiversity and ecosystem services. *Trends Ecol. Evol.* **2015**, *30*, 531–539. [[CrossRef](#)]
12. Alcon, F.; Marin-Minano, C.; Zabala, J.A.; de-Miguel, M.-D.; Martinez-Paz, J.M. Valuing diversification benefits through intercropping in Mediterranean agroecosystems: A choice experiment approach. *Ecol. Econ.* **2020**, *171*, 106593. [[CrossRef](#)]
13. Malezieux, E.; Crozat, Y.; Dupraz, C.; Laurans, M.; Makowski, D.; Ozier-Lafontaine, H.; Rapidel, B.; de Tourdonnet, S.; Valantin-Morison, M. Mixing plant species in cropping systems: Concepts, tools and models. A review. *Agron. Sustain. Dev.* **2009**, *29*, 43–62. [[CrossRef](#)]
14. Lin, B.B. Resilience in Agriculture through Crop Diversification: Adaptive Management for Environmental Change. *Bioscience* **2011**, *61*, 183–193. [[CrossRef](#)]
15. Hei, Z.; Xiang, H.; Zhang, J.; Liang, K.; Zhong, J.; Li, M.; Ren, X. Intercropping of Rice and Water Mimosa (*Neptunia oleracea* Lour.): A Novel Model to Control Pests and Diseases and Improve Yield and Grain Quality while Reducing N Fertilizer Application. *Agriculture* **2022**, *12*, 13. [[CrossRef](#)]
16. Wezel, A.; Casagrande, M.; Celette, F.; Vian, J.-F.; Ferrer, A.; Peigne, J. Agroecological practices for sustainable agriculture: A review. *Agron. Sustain. Dev.* **2014**, *34*, 1–20. [[CrossRef](#)]
17. Cuartero, J.; Pascual, J.; Vivo, J.; Özbolat, O.; Sánchez-Navarro, V.; Egea-Cortines, M.; Zornoza, R.; Mena, M.M.; Garcia, E.; Ros, M. A first-year melon/cowpea intercropping system improves soil nutrients and changes the soil microbial community. *Agric. Ecosyst. Environ.* **2022**, *328*, 107856. [[CrossRef](#)]
18. Morugan-Coronado, A.; Linares, C.; Gomez-Lopez, M.D.; Faz, A.; Zornoza, R. The impact of intercropping, tillage and fertilizer type on soil and crop yield in fruit orchards under Mediterranean conditions: A meta-analysis of field studies. *Agric. Syst.* **2020**, *178*, 102736. [[CrossRef](#)]
19. Rosa-Schleich, J.; Loos, J.; Musshoff, O.; Tschamtk, T. Ecological-economic trade-offs of Diversified Farming Systems—A review. *Ecol. Econ.* **2019**, *160*, 251–263. [[CrossRef](#)]
20. Glover, J.D.; Reganold, J.P.; Cox, C.M. Plant perennials to save Africa's soils. *Nature* **2012**, *489*, 359–361. [[CrossRef](#)]
21. Tang, Y.-L.; Yu, L.-L.; Guan, A.-M.; Zhou, X.-Y.; Wang, Z.-G.; Gou, Y.-G.; Wang, J.-W. Soil mineral nitrogen and yield-scaled soil N₂O emissions lowered by reducing nitrogen application and intercropping with soybean for sweet maize production in southern China. *J. Integr. Agric.* **2017**, *16*, 2586–2596. [[CrossRef](#)]
22. Kuyah, S.; Dietz, J.; Muthuri, C.; van Noordwijk, M.; Neufeldt, H. Allometry and partitioning of above- and below-ground biomass in farmed eucalyptus species dominant in Western Kenyan agricultural landscapes. *Biomass Bioenergy* **2013**, *55*, 276–284. [[CrossRef](#)]
23. Aguilera, E.; Lassaletta, L.; Sanz-Cobena, A.; Garnier, J.; Vallejo, A. The potential of organic fertilizers and water management to reduce N₂O emissions in Mediterranean climate cropping systems: A review. *Agric. Ecosyst. Environ.* **2013**, *164*, 32–52. [[CrossRef](#)]
24. Constantin, J.; Mary, B.; Laurent, F.; Aubrion, G.; Fontaine, A.; Kerveillant, P.; Beaudoin, N. Effects of catch crops, no till and reduced nitrogen fertilization on nitrogen leaching and balance in three long-term experiments. *Agric. Ecosyst. Environ.* **2010**, *135*, 268–278. [[CrossRef](#)]
25. Martinez-Mena, M.; Boix-Fayos, C.; Carrillo-Lopez, E.; Diaz-Pereira, E.; Zornoza, R.; Sanchez-Navarro, V.; Acosta, J.A.; Martinez-Martinez, S.; Almagro, M. Short-term impact of crop diversification on soil carbon fluxes and balance in rainfed and irrigated woody cropping systems under semiarid Mediterranean conditions. *Plant Soil* **2021**, *467*, 499–514. [[CrossRef](#)]
26. Haggard, J.; Pons, D.; Saenz, L.; Vides, M. Contribution of agroforestry systems to sustaining biodiversity in fragmented forest landscapes. *Agric. Ecosyst. Environ.* **2019**, *283*, 106567. [[CrossRef](#)]
27. Kremen, C.; Miles, A. Ecosystem Services in Biologically Diversified versus Conventional Farming Systems: Benefits, Externalities, and Trade-Offs. *Ecol. Soc.* **2012**, *17*, 40. [[CrossRef](#)]
28. Ibrikci, H.; Cetin, M.; Karnez, E.; Fluegel, W.A.; Tilkici, B.; Bulbul, Y.; Ryan, J. Irrigation-induced nitrate losses assessed in a Mediterranean irrigation district. *Agric. Water Manag.* **2015**, *148*, 223–231. [[CrossRef](#)]
29. Parras-Alcantara, L.; Lozano-Garcia, B.; Keesstra, S.; Cerda, A.; Brevik, E.C. Long-term effects of soil management on ecosystem services and soil loss estimation in olive grove top soils. *Sci. Total Environ.* **2016**, *571*, 498–506. [[CrossRef](#)]
30. Ramos, M.E.; Benitez, E.; Garcia, P.A.; Robles, A.B. Cover crops under different managements vs. frequent tillage in almond orchards in semiarid conditions: Effects on soil quality. *Appl. Soil Ecol.* **2010**, *44*, 6–14. [[CrossRef](#)]
31. Cerda, A.; Rodrigo-Comino, J.; Gimenez-Morera, A.; Keesstra, S.D. Hydrological and erosional impact and farmer's perception on catch crops and weeds in citrus organic farming in Canyoles river watershed, Eastern Spain. *Agric. Ecosyst. Environ.* **2018**, *258*, 49–58. [[CrossRef](#)]
32. Was, A.; Malak-Rawlikowska, A.; Zavalloni, M.; Viaggi, D.; Kobus, P.; Sulewski, P. In search of factors determining the participation of farmers in agri-environmental schemes—Does only money matter in Poland? *Land Use Policy* **2021**, *101*, 105190. [[CrossRef](#)]
33. Dupraz, P.; Guyomard, H. Environment and Climate in the Common Agricultural Policy. *Eurochoices* **2019**, *18*, 18–25. [[CrossRef](#)]
34. Himanen, S.J.; Maekinen, H.; Rimhanen, K.; Savikko, R. Engaging Farmers in Climate Change Adaptation Planning: Assessing Intercropping as a Means to Support Farm Adaptive Capacity. *Agriculture* **2016**, *6*, 34. [[CrossRef](#)]

35. Latvala, T.; Regina, K.; Lehtonen, H. Evaluating Non-Market Values of Agroecological and Socio-Cultural Benefits of Diversified Cropping Systems. *Environ. Manag.* **2021**, *67*, 988–999. [\[CrossRef\]](#) [\[PubMed\]](#)
36. Cagliero, R.; Legnini, M.; Licciardo, F. Evaluating the New Common Agricultural Policy: Improving the Rules. *Eurochoices* **2021**, *20*, 27–33. [\[CrossRef\]](#)
37. Arantes, A.C.C.; Cotta, S.R.; da Conceicao, P.M.; Meneghin, S.P.; Martinelli, R.; Prospero, A.G.; Boaretto, R.M.; Andreote, F.D.; Mattos, D.; de Azevedo, F.A. Implication of *Urochloa* spp. Intercropping and Conservation Agriculture on Soil Microbiological Quality and Yield of Tahiti Acid Lime in Long Term Orchard Experiment. *Agriculture* **2020**, *10*, 491. [\[CrossRef\]](#)
38. Dubey, S.K.; Sharma, N.; Sharma, J.P.; Sharma, A.; Kishore, N. Assessing citrus (lemon) based intercropping in the irrigated areas of northern plains of Haryana. *Indian J. Hort.* **2016**, *73*, 441–444. [\[CrossRef\]](#)
39. Zhang, Y.; Ni, J.P.; Yang, J.; Zhang, T.; Xie, D.T. Citrus stand ages regulate the fraction alteration of soil organic carbon under a citrus/*Stropharia rugodo-annulata* intercropping system in the Three Gorges Reservoir area, China. *Environ. Sci. Pollut. Res.* **2017**, *24*, 18363–18371. [\[CrossRef\]](#)
40. FAOSTAT. Value of Agricultural Production. Available online: <https://www.fao.org/faostat/en/#data/QV> (accessed on 24 March 2022).
41. Tan, D.; Adedoyin, F.F.; Alvarado, R.; Ramzan, M.; Kayesh, M.S.; Shah, M.I. The effects of environmental degradation on agriculture: Evidence from European countries. *Gondwana Res.* **2022**, *106*, 92–104. [\[CrossRef\]](#)
42. Soussana, J.-F. Research priorities for sustainable agri-food systems and life cycle assessment. *J. Clean. Prod.* **2014**, *73*, 19–23. [\[CrossRef\]](#)
43. Conesa, M.R.; de la Rosa, J.M.; Fernandez-Trujillo, J.P.; Domingo, R.; Perez-Pastor, A. Deficit irrigation in commercial mandarin trees: Water relations, yield and quality responses at harvest and after cold storage. *Span. J. Agric. Res.* **2018**, *16*, e1201. [\[CrossRef\]](#)
44. Gonzalez-Dugo, V.; Ruz, C.; Testi, L.; Orgaz, F.; Fereres, E. The impact of deficit irrigation on transpiration and yield of mandarin and late oranges. *Irrig. Sci.* **2018**, *36*, 227–239. [\[CrossRef\]](#)
45. Allen, R.G.; Pereira, L.S.; Raes, D.; Smith, M. *Crop Evapotranspiration. Guidelines for Computing Crop Water Requirements*; FAO Irrigation and Drainage Paper No. 56; FAO: Rome, Italy, 1998.
46. ISO 14040; Environmental Management—Life Cycle Assessment—Principles and Framework. ISO: Geneva, Switzerland, 2006.
47. ISO 14044; Environmental Management—Life Cycle Assessment—Requirements and Guidelines. ISO: Geneva, Switzerland, 2006.
48. Cerutti, A.K.; Galizia, D.; Bruun, S.; Mellano, G.M.; Beccaro, G.L.; Bounous, G. Assessing environmental sustainability of different apple supply chains in Northern Italy. In *Towards Life Cycle Sustainability Management*; Springer: Dordrecht, The Netherlands, 2011; pp. 341–348.
49. Bessou, C.; Basset-Mens, C.; Tran, T.; Benoist, A. LCA applied to perennial cropping systems: A review focused on the farm stage. *Int. J. Life Cycle Assess.* **2013**, *18*, 340–361. [\[CrossRef\]](#)
50. Perrin, A.; Basset-Mens, C.; Gabrielle, B. Life cycle assessment of vegetable products: A review focusing on cropping systems diversity and the estimation of field emissions. *Int. J. Life Cycle Assess.* **2014**, *19*, 1247–1263. [\[CrossRef\]](#)
51. Guinee, J. Handbook on life cycle assessment—Operational guide to the ISO standards. *Int. J. Life Cycle Assess.* **2001**, *6*, 255. [\[CrossRef\]](#)
52. Brentrup, F.; Kuesters, J.; Lammel, J.; Kuhlmann, H. Methods to Estimate On-Field Nitrogen Emissions from Crop Production as an Input to LCA Studies in the Agricultural Sector. *Int. J. Life Cycle Assess.* **2000**, *5*, 349–357. [\[CrossRef\]](#)
53. Audsley, E.; Alber, S.; Clift, R.; Cowell, S.; Crettaz, P.; Gaillard, G. *Harmonisation of Environmental Life Cycle Assessment for Agriculture*; Final Report, Concerted Action AIR3-CT94-2028; European Commission DG VI Agriculture: Brussels, Belgium, 1997; 139p.
54. Martinez-Alcantara, B.; Quinones, A.; Forner-Giner, M.A.; Iglesias, D.J.; Primo-Millo, E.; Legaz, F. Impact of fertilizer-water management on nitrogen use efficiency and potential nitrate leaching in citrus trees. *Soil Sci. Plant Nutr.* **2012**, *58*, 659–669. [\[CrossRef\]](#)
55. European Commission-Joint Research Centre (EC-JRC). *Handbook e Recommendations for Life Cycle Impact. Assessment in the European Context*; European Commission-Joint Research Centre: Luxembourg, 2011.
56. Berbel, J.; Martinez-Dalmau, J. A Simple Agro-Economic Model for Optimal Farm Nitrogen Application under Yield Uncertainty. *Agriculture* **2021**, *11*, 1107. [\[CrossRef\]](#)
57. European Commission. *A Farm to Fork Strategy: For a Fair, Healthy and Environmentally-Friendly Food System*; COM (2020) 381 Final; European Commission: Brussels, Belgium, 2020.
58. Fernandez, J.E.; Alcon, F.; Diaz-Espejo, A.; Hernandez-Santana, V.; Cuevas, M.V. Water use indicators and economic analysis for on-farm irrigation decision: A case study of a super high density olive tree orchard. *Agric. Water Manag.* **2020**, *237*, 106074. [\[CrossRef\]](#)
59. CARM. Estadística agraria. Precios Agrarios. Comunidad Autónoma de la Región de Murcia. 2021. Available online: [http://www.carm.es/web/pagina?IDCONTENIDO=1396&IDTIPO=100&RASTRO=c1415\\$m1174](http://www.carm.es/web/pagina?IDCONTENIDO=1396&IDTIPO=100&RASTRO=c1415$m1174) (accessed on 24 March 2022).
60. MAPA. Índice y Precios Percibidos Agrarios. Ministerio de Agricultura, Pesca y Alimentación, 2021. Available online: <https://www.mapa.gob.es/es/estadistica/temas/estadisticas-agrarias/economia/precios-percibidos-pagados-salarios/precios-percibidos-por-los-agricultores-y-ganaderos/default.aspx> (accessed on 24 March 2022).

61. Alcon, F.; Martin-Ortega, J.; Pedrero, F.; Alarcon, J.J.; de Miguel, M.D. Incorporating Non-market Benefits of Reclaimed Water into Cost-Benefit Analysis: A Case Study of Irrigated Mandarin Crops in southern Spain. *Water Resour. Manag.* **2013**, *27*, 1809–1820. [\[CrossRef\]](#)
62. Caballero, P.; Carmona, B.; Alcon, F.; Fernandez-Zamudio, M.A.; de-Miguel, M.D. Trends in Varietal Composition in Spanish Citrus-Growing and Underlying Variables. In Proceedings of the XII International Citrus Congress—International Society of Citriculture, Valencia, Spain, 18–23 November 2012.
63. Alcon, F.; Angeles Fernandez-Zamudio, M.; Lopez-Becerra, E.I.; Dolores de-Miguel, M. Survival analysis of orange tree varieties in Spain. *Span. J. Agric. Res.* **2019**, *17*, e0103. [\[CrossRef\]](#)
64. Bossio, D.A.; Cook-Patton, S.C.; Ellis, P.W.; Fargione, J.; Sanderman, J.; Smith, P.; Wood, S.; Zomer, R.J.; von Unger, M.; Emmer, I.M.; et al. The role of soil carbon in natural climate solutions. *Nat. Sustain.* **2020**, *3*, 391–398. [\[CrossRef\]](#)
65. Tamburini, G.; Bommarco, R.; Wanger, T.C.; Kremen, C.; van der Heijden, M.G.A.; Liebman, M.; Hallin, S. Agricultural diversification promotes multiple ecosystem services without compromising yield. *Sci. Adv.* **2020**, *6*, eaba1715. [\[CrossRef\]](#) [\[PubMed\]](#)
66. De Luca, A.; Falcone, G.; Stillitano, T.; Strano, A.; Gulisano, G. Sustainability assessment of quality-oriented citrus growing systems in Mediterranean area. *Calitatea* **2014**, *15*, 103.
67. Aguilera, E.; Guzman, G.; Alonso, A. Greenhouse gas emissions from conventional and organic cropping systems in Spain. I. Herbaceous crops. *Agron. Sustain. Dev.* **2015**, *35*, 713–724. [\[CrossRef\]](#)
68. Ribal, J.; Ramirez-Sanz, C.; Estruch, V.; Clemente, G.; Sanjuan, N. Organic versus conventional citrus: Impact assessment and variability analysis in the Comunitat Valenciana (Spain). *Int. J. Life Cycle Assess.* **2017**, *22*, 571–586. [\[CrossRef\]](#)
69. Martin-Gorrioz, B.; Gallego-Elvira, B.; Martinez-Alvarez, V.; Maestre-Valero, J.F. Life cycle assessment of fruit and vegetable production in the Region of Murcia (south-east Spain) and evaluation of impact mitigation practices. *J. Clean. Prod.* **2020**, *265*, 121656. [\[CrossRef\]](#)
70. Pergola, M.; D'Amico, M.; Celano, G.; Palese, A.M.; Scuderi, A.; Di Vita, G.; Pappalardo, G.; Inglese, P. Sustainability evaluation of Sicily's lemon and orange production: An energy, economic and environmental analysis. *J. Environ. Manag.* **2013**, *128*, 674–682. [\[CrossRef\]](#)
71. Clune, S.; Crossin, E.; Verghese, K. Systematic review of greenhouse gas emissions for different fresh food categories. *J. Clean. Prod.* **2017**, *140*, 766–783. [\[CrossRef\]](#)
72. Torrellas, M.; Anton, A.; Lopez, J.C.; Baeza, E.J.; Parra, J.P.; Munoz, P.; Ignacio Montero, J. LCA of a tomato crop in a multi-tunnel greenhouse in Almeria. *Int. J. Life Cycle Assess.* **2012**, *17*, 863–875. [\[CrossRef\]](#)
73. Gottwald, T.R.; Hall, D.G.; Kriss, A.B.; Salinas, E.J.; Parker, P.E.; Beattie, G.A.C.; Nguyen, M.C. Orchard and nursery dynamics of the effect of interplanting citrus with guava for huanglongbing, vector, and disease management. *Crop Prot.* **2014**, *64*, 93–103. [\[CrossRef\]](#)
74. Li, H.; Zhu, N.; Wang, S.; Gao, M.; Xia, L.; Kerr, P.G.; Wu, Y. Dual benefits of long-term ecological agricultural engineering: Mitigation of nutrient losses and improvement of soil quality. *Sci. Total Environ.* **2020**, *721*, 137848. [\[CrossRef\]](#) [\[PubMed\]](#)
75. Martins, C.R.; Rodrigues, G.S.; de Barros, I. Assessment of economic and environmental performance in citrus-based intercropping systems. *Rev. Bras. Frutic.* **2021**, *43*, e463. [\[CrossRef\]](#)
76. Yan, Y.; Yang, J.; Guo, Y.; Yang, J.; Wan, X.; Zhao, C.; Guo, J.; Chen, T. Potential evaluation of different intercropping remediation modes based on remediation efficiency and economic benefits—A case study of arsenic-contaminated soil. *Int. J. Phytoremediation* **2022**, *24*, 25–33. [\[CrossRef\]](#) [\[PubMed\]](#)
77. Srivastava, A.K.; Huchche, A.D.; Ram, L.; Singh, S. Yield prediction in intercropped versus monocropped citrus orchards. *Sci. Hortic.* **2007**, *114*, 67–70. [\[CrossRef\]](#)
78. Brandth, B.; Haugen, M.S. Farm diversification into tourism—Implications for social identity? *J. Rural Stud.* **2011**, *27*, 35–44. [\[CrossRef\]](#)