



Biofilm Reactor with Permeable Materials as Carriers Archives Better and More Stable Performance in Treatment of Slightly Polluted Water during Long-Term Operation

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Abstract: Biofilms in reactors usually grow on impermeable surfaces, and the mass transfer of nutrients in biofilms is mainly driven by diffusion, which is inefficient especially for thick biofilms. In this study, permeable materials (i.e., nylon meshes) were used as biocarriers in a biofilm reactor, and their performance was evaluated and compared with the commercial biocarriers (PE08 and PE10) used for treating slightly polluted water. The results indicate that the mesh-based bioreactor achieved complete nitrification faster than the commercial biocarriers, with a more stable and better effluent quality during long-term operation. At a two-hour hydraulic retention time, the average effluent ammonia (NH₄⁺-N) and nitrite (NO₂⁻-N) concentrations during the stabilized phase were 0.97 \pm 0.79 and 0.61 \pm 0.32 mg-N, respectively, which are significantly lower than those with commercial carriers. The estimated specific surface area activities for the mesh, PE08, and PE10 carriers were 1620, 769, and 1300 mg-N/(m²·d)), respectively. The biofilms formed on the nylon mesh were porous, while they were compact and nonporous on the PE carriers. Water with substrates might pass through the porous biofilms formed on the meshes, which could enhance mass transfer and result in a better and more stable treatment performance.

Keywords: biofilm; biofilm reactor; biocarrier; mass transfer; slightly polluted water

1. Introduction

Biofilm-based water treatment technologies, which mainly rely on the attached microorganisms on the surface of biocarriers to degrade pollutants in wastewater [1], have been widely used in wastewater treatment since they enhance biomass retention and also increase the diversity for bioreactions and microbial communities [2–5]. Biofilm reactors can be divided into fixed bed biofilm reactors and moving bed biofilm reactors, both of which have been popularly used due to their merits, e.g., easy to implement, high treatment efficiency, high tolerance to loading fluctuation, extended sludge age, and small footprint [6–9]. Therefore, biofilm-based processes have been widely embedded in various wastewater treatment facilities, e.g., for enhancing denitrification in municipal wastewater treatment plants, improving anerobic ammonia-oxidizing bacteria growth in the anaerobic ammonia oxidation process [10–13], decolorizing dye-containing textile wastewater [14], terephthalic acid wastewater treatment [15], anaerobic digestion processes [16], removing petroleum hydrocarbons [17], removing ammonia and organics from slaughterhouse and aquaculture wastewater [18], and conventional nitrification processes [19].

Powder or granular activated carbon, ceramic materials, nonwoven materials, and plastics are materials that are widely used for biofilm carriers [20–22]. Due to their low



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density and high hardness, plastics have been the most commonly used materials for biofilm carriers [23]. In addition to developing novel materials for biofilm carriers, optimizing biocarrier design is another important research direction in the development of biofilmbased technologies [24]. Previous studies on biocarriers have mainly focused on improving the ease of microorganism attachment and growth and increasing the specific surface area, with the aim to shorten the startup time and increase biofilm retention [25]. In addition to biofilm concentration, the mass transfer in biofilms impacts the activity and reactor performance as well [26]. Currently, biofilms usually grow on impermeable surfaces. Oxygen, dissolved organics, ammonia, and other nutrients are transferred from the bulk solution to the biofilm surface and then inwards, mainly being driven by the diffusion process as a result of concentration gradients. However, the diffusion process is usually inefficient, especially for thick biofilms. For biofilms growing on permeable materials, water can pass through them from either side, especially under turbulent conditions resulting from aeration, which can create convective flows that will enhance species (e.g., oxygen, substrates, and nutrients) transport from the wastewater bulk solution towards the biofilm interior and result in better mass transfer efficiency than that by diffusion only. It is most likely that reactors with permeable materials as carriers could achieve a better and more stable treatment performance.

In this study, permeable materials, i.e., nylon meshes, were used as the biocarriers to build a biofilm reactor. Its performance during the startup period and long-term operation were evaluated and compared with other commercial biocarriers in the treatment of slightly polluted wastewater.

2. Materials and Methods

2.1. Biofilm Carrier Preparation

The monofilament (approximately 80 μ m in diameter) woven nylon meshes with a pore size of approximately 100 μ m (Figure 1) were used as the biocarriers for the growth of biofilms. The screens were tightly wrapped on a 10 cm \times 15 cm \times 1 cm stainless steel frame and both sides were open. Two other commercial polyethylene (PE) biocarriers, PE10 and PE08, were used for comparison in this study. PE10 had dimensions of 25 mm (diameter) \times 4 mm (height). Each PE10 carrier had 64 voids, which had a cross-sectional area ranging from 0.031 to 0.068 cm². PE08 had dimensions of 5 mm (diameter) \times 10 mm (height), and each PE08 carrier only had 8 voids with a crosssectional area of approximately 0.011 cm².

As the biofilms grow, the uneven, solid surface of the PE carriers will be covered. Therefore, the actual effective surface area of the PE carriers is usually smaller than the advertised values. In this study, the surface area of the PE carriers was calculated by assuming their solid surface is flat. Thus, the total surface area (including the wall surface for all of the voids) for one PE08 or PE10 carrier was estimated to be about 639 mm² or 2160 mm². A container with a volume of 1 L can hold 1368 PE08 carriers or 228 PE10 carriers. Therefore, the specific surface area for PE08 and PE10 carriers was estimated to be 870 m²/m³ and 520 m²/m³, respectively. The pores of a nylon mesh are very tiny, and can be easily and quickly covered by biofilms. Therefore, the total surface area for the nylon meshes was estimated based on their size in this study. One stainless steel frame (10 cm \times 15 cm \times 1 cm) for the nylon meshes had a volume of 150 cm³ and a surface area of 600 cm² (10 cm \times 15 cm \times 4 cm). Thus, the estimated specific surface area of the mesh carriers was approximately 400 m²/m³.

2.2. Biofilm Reactor Setup and Long-Term Test

Three biofilm reactors were set up, as shown in Figure 1, to evaluate the performance of the three biocarriers in the treatment of slightly polluted water, which contained a high ammonia concentration like aquaculture wastewater and urban black and odorous water. The biofilm reactors ($20 \text{ cm} \times 10 \text{ cm} \times 25 \text{ cm}$, made of polymethyl methacrylate) had an effective working volume of approximately 4 L and the volume fraction for biocarriers was

approximately 50%. In the mesh reactor, six mesh frames were installed and the distance between two frames was approximately 2.5 cm. The water level in the reactor was 20 cm to ensure all of the carriers submerged. The three biofilm reactors were used to treat the simulated slightly polluted water, which contained COD and ammonium concentrations of 30 mg/L and 30 mg-N/L, respectively. The COD and ammonium in the influent were provided by glucose and ammonium bicarbonate, respectively. In addition, trace elements were added into the influent with the formula described previously [27]. The reactors were operated under room temperature, at approximately 23 °C. The hydraulic retention time (HRT) was controlled to be about 2 h at an inflow rate of approximately 33.3 mL/min. Air was supplied by a fine bubble diffuser at the bottom of each reactor at a rate of about 1.0 L/min. Initially, most of the PE08 and PE10 carriers in the reactors were fluidized, while most of them became settled when biofilms were formed. The dissolved oxygen (DO) concentration in the reactors was above 4 mg/L and the pH was maintained at 7–8 using sodium carbonate. Analytical methods for determining the DO concentration and pH have been described previously [28]. During the tests, the detached biofilms flowed out of the system along with the effluent. The effluent's COD, ammonium (NH_4^+-N) , nitrite $(NO_2^{-}-N)$, and nitrate $(NO_3^{-}-N)$ concentrations were regularly measured using Hach test kits with a spectrophotometer (DR3900; Hach, Loveland, CO, USA) [29].



Figure 1. Biofilm reactor setup with different carriers: (a) PE08, (b) PE10, and (c) nylon mesh.

2.3. Biofilm Morphology Observation by SEM

After their stable operation, the biofilms on the biocarriers were observed using scanning electron microscopy (SEM). The biofilm samples were dehydrated first and then coated with an aurum–platinum alloy by following the procedure provided in a previous study [30]. Finally, an S-4700 scanning electron microscope (Hitachi, Japan) was used to observe the samples.

2.4. Estimating the Surface Loading Rates and Specific Surface Area Activity The surface area loading rates were calculated using Equation (1):

$$SALR = \frac{ANLR_{avg} \times V}{TSA}$$
(1)

where $ANLR_{avg}$ is the average ammonium loading rate (ANLR) (kg-N/(m³·d)), and TSA is the total surface area, which is 0.30, 0.87, and 0.52 m² for the mesh, PE08, and PE10 biocarriers at a filling rate of 50%.

The specific surface area activity (*SSAA*, mg-N/($m^2 \cdot d$)), which assumes the surface area available for microbe growth, can be estimated using Equation (2):

$$SSAA = \frac{Q(S_0 - S)}{SSA \times R \times V}$$
(2)

where *Q* is the inflow rate (m³/d); *S*₀ is the substrate concentration in the influent (mg-N/L); S is the substrate concentration in the effluent (mg-N/L); *SSA* is the specific surface area (m²/m³), which is 400, 870, and 520 m²/m³ for the mesh, PE08, and PE10 biocarriers; *V* is the reactor volume (m³); and *R* is the filling rate of the biocarriers in the reactors, which is 50% in the study.

3. Results and Discussion

3.1. Reactor Performance during Start-Up Period

Figure 2 depicts the effluent NH_4^+ -N, NO_2^- -N, and NO_3^- -N concentrations of the three reactors equipped with different biocarriers during the 455 days of operation. In the start-up phase, the effluent NH_4^+ -N concentration in the PE08 biofilm reactor decreased to below 1 mg-N/L after 26 days of operation, while the effluent NO_2^- -N concentration fluctuated between 20 and 25 mg-N/L from the 23rd to the 120th day, and then decreased to approximately 1 mg-N/L after the 130th day. In the PE10 biofilm reactor, the effluent NH_4^+ -N and NO_2^- -N concentrations decreased to less than 1 mg-N/L after 40 and 105 days of operation, respectively. In the mesh-based biofilm reactor, however, the effluent NH_4^+ -N and NO_2^- -N concentrations were less than 1 mg-N/L starting from the 69th and 78th day, respectively. In addition, during the start-up period, the peak value of NO_2^- for the PE08 and PE10 biofilm reactor was as high as 20 mg-N/L or even more, while it was below 20 mg-N/L in the mesh-based biofilm reactor. During nitrification, ammonia is oxidized into nitrite by ammonia oxidizers, and then into nitrate by nitrite oxidizers [31]. These results, as shown in Figure 2, again indicate that the ammonia oxidizers grew faster than nitrite oxidizers in the biofilm reactors without the retention of suspended solids.

It is well known that a long start-up time is required due to the slow growth of nitrifiers, especially for nitrite oxidizers [32]. In this study, it took 113, 119, and 78 days for the biofilm reactors filled with PE08, PE10, and mesh biocarriers to achieve complete nitrification, suggesting that mesh biocarriers could shorten the start-up time even at a short HRT of 2 h without the retention of suspended solids. In another study, it took 84 days to initiate complete nitrification using a novel biocarrier [33]. Previous studies have shown that although PE is the most commonly used carrier for biofilm reactors [26], its high hydrophobicity and low surface energy limit the initial microbial cell attachment, which may be the reason for the longer start-up time [34].

3.2. Reactor Performance under Stabilized Conditions

During the subsequent operation after start-up, the effluent nitrite in the PE08 biofilm reactor significantly fluctuated between 4 and 10 mg-N, while the effluent ammonia was more stable, ranging from 0.01 to 3.0 mg-N. In the PE10 biofilm reactor, the effluent ammonia fluctuated significantly between 0.1 and 10 mg-N/L, while the effluent nitrite was more stable, being generally less than 2 mg-N/L. However, both the effluent ammonia and nitrite were very stable in the mesh-based biofilm reactors. As shown in Figure 3, the average effluent ammonia concentration for the mesh biofilm reactor was 0.97 ± 0.79 mg/L, which was significantly (p < 0.05) lower than those in the PE08 (1.84 ± 2.18 mg/L) and PE10 (2.11 ± 3.07 mg/L) biofilm reactors. In addition, the average effluent nitrite concentration of the mesh biofilm reactor was significantly (p < 0.01) lower than those in the PE10 (2.37 ± 2.09 mg/L) biofilm



reactors. These results indicate that a better and more stable effluent quality was obtained in the biofilm reactor with mesh as the biocarrier.

Figure 2. Effluent NH_4^+ -N, NO_2^- -N, and NO_3^- -N concentrations during long-term operation in the biofilm reactors with carriers of (**a**) PE08, (**b**) PE10, and (**c**) mesh.



Figure 3. The box plot of effluent NH₄⁺-N, NO₂⁻-N, and NO₃⁻-N concentrations in the reactors with different biocarriers under stabilized conditions (ns = no significant differences; * indicates p < 0.05; ** indicates p < 0.01; *** indicates p < 0.001).

3.3. SALR and SSAA for Different Biocarriers during Long-Term Tests

The surface area loading rates (*SALR*) and the specific surface area activity (*SSAA*) for the three biocarriers were calculated, with results shown in Figure 4. Due to the PE08 and PE10 biocarriers having a higher specific surface area, the estimated *SALR* for PE08 (828 mg-N/(m²·d)) and PE10 (1385 mg-N/(m²·d)) were lower than that (2400 mg-N/(m²·d)) in the mesh biofilm reactors. However, the mesh biofilm reactor had a better treatment performance, and its *SSAA* (1620 mg-N/(m²·d)) was significantly greater than those in the PE08 (769 mg-N/(m²·d)) and PE10 (1300 mg-N/(m²·d)) biocarriers. These results suggest that, although the mesh biocarriers had a lower specific surface area, they performed significantly better than the commercial biocarriers of PE08 and PE10.



Figure 4. Surface area loading rates (*SALR*) and specific surface area activity (*SSAA*) for different biocarriers during long-term tests.

3.4. Biofilm Morphology Formed on Nylon Mesh and PE Carriers

As shown in Figure 5, the biofilms formed on the surface of the PE biocarrier are very compact without obvious pores. The biofilm cracks shown in Figure 5f might be caused by excessive dehydration prior to SEM analysis. After zooming in further on the image, it can be observed that bacterial cells were embedded in compact gel-like EPS. EPS refers to macromolecular polymers secreted by microorganisms, and mainly includes extracellular proteins, polysaccharides, and nucleic acids, which can accelerate the formation of biofilms. As shown in Figure 5b,c, however, porous and thick biofilm layers were formed on mesh-based carriers. A large number of microorganisms were observed in the biofilms. These results indicate that the biofilms formed on the PE carriers and permeable materials (i.e., nylon meshes) had significantly different structures. However, why the biofilms that formed on the permeable and impermeable materials had significantly different structures demands further studies.

Liquid diffusion and nutrient penetration through the biofilms are important factors affecting the overall treatment performance. Previous studies have suggested that thin and porous biofilms perform much better in terms of substrate transport even under normal operation conditions [35]. During a long-term operation, the initial advantages of fast biofilm growth and high microbial activity might be lost due to partial or complete blockage [36], particularly under a high organic load [9]. Desmond et al. (2018) found that the biofilms with compact and homogeneous structures had a higher hydraulic resistance, while those with a heterogeneous and porous structure had a lower hydraulic resistance [37]. Obviously, the biofilms with lower hydraulic resistance will have much better substrate transfer efficiency inside the biofilms. As shown in Figure 5b,c, the porous structure for the



biofilms formed on the new biocarriers will allow water, oxygen, and substrates to pass through the biofilms easier, and then created a better environment for inner biofilm growth.

Figure 5. Scanning electron microscopy analysis of different biocarriers: (a) nylon mesh (having pore size of approximately 100 μ m) with biofilm scraped; (b,c) biofilms formed on the nylon mesh biocarriers; (d–f) biofilms formed on the side surface of commercial polyethylene (PE) biocarriers.

3.5. Proposed Mass Transfer in the Biofilms Formed on Permeable Materials

In biofilm reactors, biofilms usually grow on impermeable surfaces, i.e., dead-end biofilms (Figure 6a). Oxygen, dissolved organics, ammonia, and other nutrients are transferred from the bulk solution to the biofilm surface and then inwards to support biofilm growth [38]. When biofilms become thick, usually the inner biofilms will become inactive since oxygen and other electron acceptors are used up during the diffusion process from the surface to the interior [36–41]. As a result, the adhesions for the inner biofilms or those between the biofilms and solid surface become weak, leading to biofilm detachment. In biofilm reactors, biofilm detachments will impact treatment performance until fresh and active biofilms are formed again [42]. Therefore, the mass transfer in the biofilms is very important, as it not only impacts the treatment performance but also the activity of inner biofilms and biofilm detachment.



Figure 6. (a) The schematic for the mass transfer in the biofilms formed on the impermeable surface; (b) the proposed schematic for the mass transfer in the biofilms formed on the permeable surface (i.e., nylon meshes).

For the biofilms developed on the impermeable surface (i.e., dead-end biofilms), the mass transfer of nutrients in the biofilms is mainly driven by the diffusion process as the result of concentration gradients [43]. The differences in the concentration gradient, biofilm thickness, and inner structures will impact the diffusion efficiency [44]. Previous studies have shown that in aerobic systems, sufficient turbulence is created by aeration to control the turnover of microorganisms in biofilms, and then appropriate biofilm thickness is maintained by shear force [9]. As a result, the activity of biofilms and their effective mass transfer performance are maintained [9]. Our research showed that when permeable materials, e.g., meshes, are used as biocarriers, thick biofilms will be formed on both sides of the mesh and also inside the mesh pores (Figure 5). Since the mesh are thin and porous, water may pass through the biofilms from either side, especially under turbulent conditions resulting from aeration, creating convective flows that will bring species (e.g., oxygen, substrates, and nutrients) from bulk solutions to the biofilm interior and then result in greater mass transfer efficiency than that by diffusion only. These water-permeable biofilms that are formed on the meshes are shown in Figure 6b. The mesh pores can additionally provide more spaces for biofilm inhabitation, further enhancing the treatment efficiency. Moreover, after the biofilms form and grow inside the pores of PE08 and PE10, the contact area between biofilms and wastewater per unit carrier can be higher in the nylon meshes than in PE08 and PE10, which may be another reason for the better and more stable treatment performance of the mesh biofilm reactor.

This study demonstrated that the biofilm reactor with permeable materials (i.e., meshes) as carriers achieved a better and more stable effluent quality. However, more studies should be conducted to characterize the mass transfer performance and activities in biofilms formed on permeable surfaces (Figure 6b) and validate the proposed mechanisms.

4. Conclusions

The performance of a biofilm reactor with a nylon mesh as the biocarrier was evaluated and compared with other commercial biocarriers in the treatment of slightly polluted water. The mesh-based bioreactor achieved a faster start-up time and more stable and better effluent quality than the commercial biocarriers. The estimated specific surface area activities for the mesh, PE08, and PE10 carriers were 1620, 769, and 1300 mg-N/(m²·d)), respectively. On the permeable meshes, convective and porous biofilms formed and water with substrates passed through the mesh biofilms, which enhanced the mass transfer, and in turn, achieved a better and more stable treatment performance.

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