



Article Impacts of Complex Terrain Features on Local Wind Field and PM_{2.5} Concentration

Yuqiang Song¹ and Min Shao^{2,*}



² School of Environment, Nanjing Normal University, Nanjing 210046, China

* Correspondence: mshao@masonlive.gmu.edu

Abstract: Complex topography has nonnegligible effects on local meteorological conditions as well as the transportation of atmospheric pollutants, which deserves more extensive study. In this study, the impacts of complex terrain features (mountains and river valleys) on local wind field and PM_{2.5} concentration in a typically developed mega city along the Yangtze River were studied numerically using the WRFCALMET-CALPUFF system. The impacts of different model grid and terrain horizontal resolutions were firstly investigated against observations. Then, the impacts of terrain features, specifically the impacts of Mt. LS and the Yangtze River, on wind field and PM_{2.5} transportation were analyzed by "removing" Mt. LS and the Yangtze River from the meteorological diagnostic model and simulating the dispersion of PM_{2.5} from three virtual point sources in the chemical model. Results showed that: (i) higher terrain elevation and model horizontal resolutions, and updated land cover types, can effectively improve the prediction of wind direction where terrain features are complex; (ii) Mt. LS mainly acts as a barrier, and ridge wind is weakened after "removing" Mt. LS; (iii) after "removing" the Yangtze River, the transport of PM_{2.5} along the Yangtze River is weakened; (iv) the simulation of PM_{2.5} from virtual point sources showed that Mt. LS could have an effect of up to 55% on the PM_{2.5} concentration in Nanjing. This study showed that the local complex topographies have an obvious effect on the local wind field and the concentration of PM2.5. Therefore, it is important to consider the influence of local topographies and land cover types when predicting local wind field and air quality.

Keywords: WRF; CALMET; CALPUFF; PM_{2.5} concentration; wind field; complex terrain features

1. Introduction

Due to rapid development, air pollution has attracted more public attention [1,2]. Frequent haze pollution is closely related to adverse meteorological conditions and high pollutant emissions [3–6]. Evidence has shown that air quality in China has been improved a lot owing to the large decrease in anthropogenic emissions, and meteorological conditions are recently becoming key factors on air quality [7,8]. Severe air pollution characterized by high concentrations of PM_{2.5} can occur more frequently over complex terrains even with relatively weak anthropogenic emissions, due to comprehensive impacts of underlying features on planetary boundary layer (PBL) conditions [9]. In addition, new particles may form under different levels of air pollutants combined with meteorological effects [10]. Therefore, it is important to study quantitatively the impacts of complex terrain features on the transportation and distribution of air pollutants.

Meteorological conditions in the PBL are the main factors affecting air quality, especially near-surface meteorological elements, such as wind speed, wind direction, relative humidity, temperature, and mixed-layer thickness, which have a direct influence on the diffusion and transport of pollutants in the PBL [11]. Among all those meteorological factors, wind speed and wind direction affect pollutants more frequently. Complex terrains usually exhibit complex effects (dynamically and thermally) on local wind fields; therefore,



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Copyright: © 2023 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). local air quality can also be affected. Pollution is often accompanied by weak near-surface wind which could lead to the accumulation of pollutants, while wind direction determines the transmission and diffusion of pollutants [12,13]. Over complex terrains with mountains, valleys, and multiple land covers, wind speed and wind direction in the PBL are strongly affected by dynamical and thermal processes (non-geostrophic winds are usually formed), thus weakening/strengthening the ability of the atmosphere to remove local atmospheric pollutants [14,15]. However, current quantitative analysis of how complex terrains affect local wind field and air quality is still insufficient, and numerical simulation results often have large errors due to the insufficient representation of local terrain features.

Due to the temporal and spatial limitations of the observation data, numerical models have become a widely used method to study the diffusion and transport of pollutants under complex topographies. Most studies focus on the analysis of the atmospheric boundary layer structure under different pollution processes [16] and the impacts of different boundary layer parameterization schemes [17,18]. However, it is difficult for numerical models to accurately capture the atmospheric boundary layer structure and wind field over complex topographies due to the parameterized turbulence calculation in the PBL. For example, the Weather Research and Forecasting (WRF) model usually overestimates near-surface wind speed [19,20], which could be due to the difficulty in accurately describing surface roughness under complex topographies [21]. In addition, horizontal and vertical grid resolutions also affect the simulation of near-surface wind [22]. Studies have shown that it is difficult to simulate the wind profile with WRF below 100 m [22], and large wind direction errors often occur when the wind speed is low [23]. The bias of wind simulation under complex topographies can be reduced when a large-eddy simulation is introduced [9]. However, such a procedure not only requires more accurate topographic information and more detailed land cover types [24,25], but also needs large computational resources. In recent years, the coupling of mesoscale models and small regional models has been widely used in near-surface wind field simulation. For example, studies have shown that the coupling of WRF and CALMET (the meteorological diagnosing part of the air quality dispersion model CALPUFF) can better simulate near-surface wind fields [26]. Furthermore, such a procedure could save lots of computational resources compared to large-eddy simulation.

Nanjing, one of the largest developed cities in China, is located on the east coast of the middle latitude continent and the lower reaches of the Yangtze River. The widest part of the Yangtze River in Nanjing can reach 2 km, and the Ningzhen Mountains and Mt. LaoShan (LS) are located along the Yangtze River as shown in Figure 1. The average altitude of Mt. LS is about 200 m, and the highest peak is 442 m. Due to large anthropogenic emissions, it is of great scientific significance to study the diffusion and transmission law of pollutants under complex topographies to improve the air quality of Nanjing and maintain the coordinated development of environment and economy. Based on meteorological observations during 2017–2019 in Nanjing, it is found that the dominant wind direction of the Pukou Station (PK) in Nanjing is very different from that of the other station (Figure 1a,b). By comparing the hourly observation data of the PK and the Liuhe Station (LH) in Nanjing from 2017 to 2019, it is found that a wind direction difference of greater than 90° and between 45° and 90° accounted for 13.5% and 17.9%, respectively. The difference in wind direction between the two stations is probably due to the complex terrain features. Therefore, in order to accurately simulate the wind field and the transport and diffusion of PM_{2.5} in Nanjing, the impacts of model grid and terrain height data resolutions, land cover types, and the Mount LS and Yangtze River topographies are studied in this study.

The aim of the study is to discuss the impacts of different model grid and topographic elevation resolutions, and land cover types, on local wind field simulation as well as the transport and distribution of $PM_{2.5}$. In this manuscript, data and the model configurations are described in Section 2. Then, the model performances with different configurations are evaluated in Section 3.1, and the impacts of complex terrain features on local wind field and



the distribution of PM_{2.5} concentration are discussed in Sections 3.2 and 3.3. Conclusions are made in Section 4.

Figure 1. (a) Wind rose of LH Station from 2017 to 2019; (b) wind rose of PK Station from 2017 to 2019; (c) the simulation area of CALMET. The red dots are the locations of the LH and PK Stations, the red pentagrams are the locations of the virtual point sources, the mountainous areas are labeled with yellow text, and the shading is the topographic height (the dark blue area represents the Yangtze River).

2. Data and Methods

2.1. Data

The meteorological observation data used in this study are the hourly observation data of wind speed and direction from the national meteorological stations PK (118.62° E, 32.05° N) and LH (118.83° E, 32.35° N) in Nanjing, obtained from the China Meteorological Data Service Centre of the National Meteorological Information Centre as shown in Figure 1, which is also used to evaluate the model performance. The wind observations were measured at 10 m above the ground using a ZQZ-TFR wind sensor, which has a sampling frequency of 1 s⁻¹; the sampling accuracy is 0.1 m s⁻¹ and 0.01° for wind speed and wind direction, respectively.

The initial and boundary conditions for WRF simulation are provided by the Global Data Assimilation System with a horizontal resolution of 0.25° and a temporal resolution of 6 h. The topography height data used in WRF and CALMET are updated using the Shuttle Radar Topography Mission (SRTM-3) with a horizontal resolution of approximately 100 m. In addition, the symbols and units used in the manuscript are shown in Table S3.

2.2. Model Configurations

In this study, WRF version 4.2, jointly developed by the National Centers for Environmental Prediction (NCEP), the National Center for Atmospheric Research (NCAR), and other departments [27], was used to generate the first meteorological guesses for the CAL-MET diagnostic model. CALMET is a meteorological module of the unsteady Lagrange model CALPUFF for downscale analysis, which includes a diagnostic wind field generator containing objective analysis and parameterized treatments of slope flows, kinematic terrain effects, terrain blocking effects, a divergence minimization procedure, and a micrometeorological model for overland and overwater boundary layers. Configurations of WRF are shown in Table 1. Firstly, two sets of WRF simulations, namely the WRF1000 (the innermost domain has a horizontal resolution of 1 km) and WRF300 (the innermost domain has a horizontal resolution of 300 m), are configured with three layers of grid nesting with different horizontal resolutions. For WRF300, topographic correction for surface winds to represent extra drag from sub-grid topography and enhanced flow at hill tops is turned on (topo_wind = 1). The outermost grid covers the Yangtze River Delta and its surrounding areas, and the innermost grid covers the entire study area.

Table 1.	Configurations	of	WRF.
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The Settings of the Study Area			
Experiment name	WRF1000	WRF300	
Initial condition	$GDAS/FNL 0.25^{\circ} \times 0.25^{\circ}$ (DOI: 10.5065/D65Q4T4Z)		
Projection/central geolocation	Lambert/(119° E, 33° N)		
Topography	GMTED2010 (30s, ~1 km)	/SRTM (~100 m)	
Integration step	60 s	30 s	
	D01 (9 km), 229 × 279	D01 (7.5 km), 92×94	
Three layers of nesting	D02 (3 km), 276 × 312	D02 (1.5 km), 106×151	
	D03 (1000 m), 123 × 177	D03 (300 m), 266 × 311	
Vertical resolution	37 layers from near ground to 100 hPa (4 layers between 1013 hPa and 1000 hPa)		
Microphysical process	New Thompson et al. scheme		
Radiation scheme	ŔRTMG		
Cumulus parameterization scheme	None		
Near-surface layer scheme	Eta Similarity		
Land surface process	Noah Land Surface Model		
Boundary layer + topography correction	Boundary layer + topography correction Yonsei University scheme + Topo_wind = 0 Yonsei University scheme + Topo_v		

The innermost results of the WRF simulations were then used as first guesses for the CALMET model for further meteorological field downscaling. The simulation region of CALMET is shown in Figure 1c with a horizontal resolution of 150 m and 25 unevenly distributed vertical layers from near ground to 2000 m (with a resolution of 20 m below 200 m). In order to optimize the wind simulation, three WRF-CALMET simulations with different first guesses and land-use types (LU) were configured, namely WRF1000-CALMET150, WRF300-CALMET150, and WRF300-CALMET150-LU. The WRF1000-CALMET150 and WRF300-CALMET150 experiments used WRF default USGS land cover types (1999 version), while WRF300-CALMET150-LU updated the land cover types to 2021 based on the 30 m resolution land cover data of China developed by Yang and Huang [28], and the relevant parameters of different land cover types were updated correspondingly. The modified parameters include: surface roughness, albedo, Bowen ratio, soil heat flux parameter, and leaf area index.

In addition, in order to reflect the influence of local topographies on wind field and the distribution of $PM_{2.5}$ concentration, the Yangtze River and Mount LS within the study area were respectively "removed" in some simulations (for LS, the terrain height of grids within the LS region was reduced to 10 m, so that Mt. LS became a flat region; for the Yangtze River, the terrain heights of grids within the Yangtze River and its coastal areas were all set to 10 m, and the land type was changed to built-up land). Three virtual emission point sources (located in the Jiangbei Chemical Industry Park, Qixia District, and Gulou District, respectively, as shown in Figure 1c) were respectively configured to test the transport and diffusion of $PM_{2.5}$. The base height, stack elevation, exit diameter, exit velocity, exit temperature, and emission rate were 50 m, 5 m, 3 m, 10 m s⁻¹, 363 K, and 1000 tons per year, respectively. CALPUFF was adopted to simulate $PM_{2.5}$ transport and diffusion, and eight groups of experiments were set up. The descriptions of each experiment are shown in Table 2. In this study, January, April, July, and October of 2019 were selected as simulation periods, representing winter, spring, summer, and autumn, respectively. Another group of experiments was conducted to compare the dispersion forecasts by using the WRF1000 and

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WRF300_CALMET150 meteorology, which also recommends higher precision simulation (Text S3 and Figure S4).

Table 2. CALPUFF experiment setup.

Experiment Name	Settings of Topographies and Virtual Point Sources		
CTRL_HGY	Real topographies, virtual point source in Jiangbei Chemical Industry Park		
CTRL_QX	Real topographies, virtual point source in Qixia District		
CTRL_GL	Real topographies, virtual point source in Gulou District		
LS_HGY	"Remove" LS, virtual point source in Jiangbei Chemical Industry Park		
LS_QX	"Remove" LS, virtual point source in Qixia District		
LS_GL	"Remove" LS, virtual point source in Gulou District		
YZR_HGY	"Remove" the Yangtze River and its coastal areas, virtual point source is in Jiangbei Chemical Industry Park		
YZR_QX	"Remove" the Yangtze River and its coastal areas, virtual point source in Qixia District		
YZR_GL	"Remove" the Yangtze River and its coastal areas, virtual point source in Gulou District		

3. Results and Discussion

3.1. Comparative Analysis of Downscaled Wind Field from Different Experiments

As the PK station is located within Mt. LS, and the dominant wind direction is quite different from the other station in Nanjing, the simulated wind at PK was compared with observations to evaluate the model's performance. The comparisons between hourly simulated wind speed and observed wind speed at the PK station in different experiments are shown in Figure 2. In general, increasing horizontal resolutions and the precision of the underlying surface do not significantly improve the simulation of wind speed, and CALMET generally performs better than WRF. The root-mean-square error (RMSE) of wind speed of WRF1000_CALMET150 fluctuates in the range of 1.03–1.23 m/s. However, the RMSE of wind speed fluctuates in the range of 1.57–3.72 m/s and 1.64–3.72 m/s after the improvement of horizontal resolution (WRF300_CALMET150_LU), respectively. Meanwhile, the correlation coefficients between simulated wind speed and observed wind speed in the three experiments are 0.47–0.65, 0.21–0.54, and 0.21–0.54.

The evaluation results of wind direction are shown in Table 3, Text S2 and Figure S3. It indicates that improving the model horizontal resolution, topographic elevation, and land cover types can significantly improve the simulation of wind direction. The RMSE of wind direction in WRF1000_CALMET150 fluctuates in the range of 102.7–166.6°, while those in WRF300_CALMET150 and WRF300_CALMET150_LU fluctuate in the range of 59.6–77.6° and 56.0–71.6°. Meanwhile, compared with WRF1000_CALMET150, the absolute biases (Abias) of WRF300_CALMET150 and WRF300_CALMET150_LU are reduced by 33.3-60.5% and 35.9-66.1%, respectively. These promising results showed that the resolutions of model grids and topographic elevation are critical to the simulation of wind direction, especially over regions with complex terrains. This is because WRF is a mesoscale model which can hardly deal with small-scale features, such as PBL turbulence, slope wind, etc., unless with the use of the large-eddy simulation. However, CALMET is a diagnostic model with advanced abilities describing small-scale features; therefore, CALMET is expected to be better than WRF in wind field simulation. For terrains that are relatively flat, further improvement of WRF grid resolution did not show much positive effect on wind simulation (Tables S1 and S2). This is because the theoretical finest horizontal resolution to use PBL parameterizations is 1 km. Further improvement of horizontal resolution could cause unrecognized bias. In addition, these similar statistics for the LH station are shown in Text S4, Tables S1 and S2.









Figure 2. Cont.

Wind Speed(m/s)



Figure 2. Comparisons of downscaled wind speed from different experiments: (**a**) WRF1000_CA LMET150; (**b**) WRF300_CALMET150; (**c**) WRF300_CALMET150_LU.

Month	Index	WRF1000_CALMET150		WRF300_CALMET150		WRF300_CALMET150_LU	
		WRF	CALMET	WRF	CALMET	WRF	CALMET
January	Abias	100.4°	95.5°	53.3°	51.8°	53.3°	46.6°
	RMSE	140.5°	135.5°	70.9°	69.7°	70.9°	63.2°
April	Abias	75.3°	72.6°	48.3°	47.2°	48.3°	41.0°
	RMSE	102.9°	102.7°	64.3°	62.5°	64.3°	56.5°
July	Abias	90.1°	90.4°	55.3°	60.3°	55.3°	54.4°
	RMSE	127.9°	130.2°	71.6°	77.6°	71.6°	70.2°
October	Abias	112.6°	119.6°	44.4°	48.3°	44.5°	40.5°
	RMSE	162.1°	166.6°	59.6°	64.1°	59.6°	56.0°

3.2. Influences of Different Topographies

As discussed above, increasing the horizontal resolution of the first guess of CAL-MET and improving the precision of the underlying surface can significantly improve the simulation of wind direction. Therefore, the above meteorological fields were used as the benchmark to quantify the complex topographies effects by "removing" Mt. LS and the Yangtze River in the study area, respectively. The differences in monthly mean wind speed and wind vector in the four seasons of 2019 are shown in Figures 3 and 4.

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Figure 3. Differences in wind field between "removing" Mt. LS and CTRL: (**a**–**d**) monthly mean wind speed differences; (**e**–**h**) monthly mean wind vector differences (colored arrows).





As shown in Figure 3a–d, the variations of wind speed after "removing" Mt. LS range from –25.0% to 29.3%. The wind speed is increased at the foot of Mt. LS, but is weakened at the ridge area. This phenomenon indicates that Mt. LS has a strong barrier effect on the upslope wind. As shown in Figure 3e, f, after the "removal" of Mt. LS, the upslope wind towards the ridge is strengthened. Especially on the southeast side of Mt. LS, not only the upslope wind, but also the northeast wind along the southeast side of Mt. LS, is strengthened.

As shown in Figure 4a–d, after "removing" the Yangtze River and its coastal areas, the variations of wind speed are between -22.0% and 19.7%. The wind speed in the Yangtze River channel is generally enhanced, with the greatest enhancement occurring at the south of Baguazhou (an island in the Yangtze River) where the Yangtze River splits, while the wind speed on the land, especially the south part of Baguazhou, is slightly weakened. As shown in Figure 4e,f, after "removing" the Yangtze River, the variations of wind vector in most regions in the study area are less than 0.1 m/s. It is worth noting that a strong transport channel is observed at the south of Baguazhou from southeast to west. The results showed that the Yangtze River can weaken the transport of $PM_{2.5}$ from southeast to northwest to some extent.

3.3. Impacts of Mt. LS and the Yangtze River on PM_{2.5} Transport

In order to study the influence of local topographies on wind field and PM_{2.5} transport, the Yangtze River and Mt. LS were respectively "removed" as described in Section 2, three virtual point emission sources (Figure 1c) were respectively set, and CALPUFF was adopted to simulate PM_{2.5} transport.

Figures S1 and S2 are cases where PM_{2.5} transportations were obviously affected by terrain features (22:00 LST on January 10th to 00:00 LST on 11 January 2019 for Mt. LS, and from 22:00 LST on April 3rd to 2:00 on 4 April 2019 for the Yangtze River). As shown in Figure S1, before Mt. LS is "removed" (CTRL_HGY, CTRL_QX, and CTRL_GL), the transport of PM_{2.5} is affected by the wind field and transported southwest along the southeast side of Mt. LS, and tends to be blocked by Mt. LS. The transport of $PM_{2.5}$ in experiments CTRL_HGY and CTRL_QX is more obvious. In experiment CTRL_GL, the transport of PM_{2.5} across the Yangtze River is blocked by Mt. LS and then moves towards the southwest. After Mt. LS is "removed", the concentration of PM_{2.5} in the Mt. LS direction of different experiments increases by about 0.2–0.3 μ g/m³, which is about 20% higher than those CTRL experiments. Similar to the above conclusions, there have been some other studies on the relationship between terrains, wind, and pollutant concentration. By studying the relationship between diurnal wind and transport of atmospheric aerosol in the Columbia River gorge of Oregon and Washington, Mark et al. [29] find that the concentration of pollutants is correlated with the prevailing wind. The study of Julian et al. [9] indicates that the central basin-shape section of the valley is poorly ventilated, and hence air pollution there would originate mostly from local emission sources. Black carbon particulates are being transported and deposited all round the year in the Himalayas and the surrounding regions; pre-monsoon and monsoon seasons contributed to the largest amounts of deposition [30]. As shown in Figure S2, before the "removal" of the Yangtze River (CTRL_QX and CTRL_GL), PM_{2.5} is basically transported from the point sources to the east, and is blocked by Mt. LS after crossing the Yangtze River. The concentration of $PM_{2.5}$ clearly decreases when reaching Mt. LS, and tends to be transported to the southwest, but this tendency is not obvious. After the Yangtze River is "removed" (YZR_QX and YZR_GL), the transport of $PM_{2.5}$ to the west is weakened, thus leading to a decrease in $PM_{2.5}$ concentration. However, in experiment YZR_QX, the concentration of PM_{2.5} increases at the south of Baguazhou where the Yangtze River splits, which is related to the enhanced east wind in this area. In experiment YZR_GL, the concentration of PM_{2.5} also increases in a small area over the Yangtze River, which may be caused by the weakening of the transport ability of PM_{2.5} across the river after the Yangtze River is "removed". During other periods, such topographic effects are similar to those in Figures S1 and S2 when the background

wind field is dominated by east or northeast wind, which will not be elaborated on here (Text S1).

In order to analyze the overall influence of topographies on $PM_{2.5}$ transport in different seasons, the distributions of monthly mean $PM_{2.5}$ concentration differences in different seasons after changing the topographies are depicted in Figures 5 and 6, for Mt. LS and the Yangtze River, respectively. The impact of "removing" Mt. LS on $PM_{2.5}$ transport is mainly concentrated in a fan-shaped area in the southwest of Nanjing, which increases the $PM_{2.5}$ concentration along the Mt. LS direction. The affected areas are similar in each month, and are concentrated in the Mt. LS region. The concentration of $PM_{2.5}$ simulated by experiment LS_GL decreases in areas over the Yangtze River (especially in April). This decrease in $PM_{2.5}$ concentration could be attributed to the changed local circulations; for example, the weakened mountain-valley circulation during nighttime could lower the $PM_{2.5}$ is more divergent, which indicates that $PM_{2.5}$ transport over the Yangtze River on $PM_{2.5}$ is more divergent, which indicates that $PM_{2.5}$ transport over the Yangtze River and its surrounding areas is significantly affected by the wind direction. This also confirms that the accuracy of wind field simulation is extremely important for the diffusion and transport of $PM_{2.5}$ under complex topographies.



Figure 5. Variations of the spatial distribution of monthly mean PM_{2.5} concentration simulated by CALPUFF after "removing" Mt. LS. First row: LS_HGY-CTRL_HGY; second row: LS_QX-CTRL_QX; third row: LS_GLCTRL_GL.



Figure 6. Similar to Figure 5, but for the "removal" of Yangtze River.

In order to quantify the influence of Mt. LS (since the impact of Mt. LS on $PM_{2.5}$ is well distributed) on $PM_{2.5}$ concentration level, the monthly average increases/decreases in $PM_{2.5}$ concentration in percentages are summarized in Table 4. The percentage variation of $PM_{2.5}$ concentration is calculated as the monthly domain average increase or decrease of $PM_{2.5}$ due to the "removal" of Mt. LS divided by the average $PM_{2.5}$ concentration simulated in the study area. After "removing" Mt. LS, all three experiments (LS_HGY, LS_QX, and LS_GL) result in a greater increase than decrease in $PM_{2.5}$ concentration in the study area, with the largest increase found in LS_GL in January (55.0%) and the largest decrease found in LS_GL in April (-33.7%). Therefore, the transport of $PM_{2.5}$ from emissions in Gulou District are more effectively transported to the west, which significantly increases the concentration of $PM_{2.5}$ in the Mt. LS area (Figure 5).

Table 4. Monthly domain averaged percentage variation of $PM_{2.5}$ concentration by "removing" Mt. LS (%).

Point Source	Mean Concentration Variation	January	April	July	October
HGY	Average increase	24.2	21.4	22.0	22.5
HGY	Average decrease	-11.8	-11.5	-13.8	-13.9
QX	Average increase	25.4	23.1	18.3	31.1
QX	Average decrease	-10.1	-14.9	-14.5	-10.7
GL	Average increase	55.0	47.4	40.0	32.7
GL	Average decrease	-25.2	-33.7	-27.0	-31.7

4. Conclusions

Nanjing is located in a hilly area with the Yangtze River running through it, which has formed a very complex topography. The observations from meteorological stations LH and PK (from 2017 to 2019) showed that there are significant differences in dominant wind direction between the two stations, which may be affected by the local topographies. In this study, WRF and CALMET were used to conduct downscale analysis to study the impacts of different grid and topographic elevation resolutions, and land cover types, on local wind field simulation. In addition, the CALPUFF model was used to study the influence of Mt. LS and the Yangtze River on the transport of PM_{2.5} from three virtual point sources at different locations in Nanjing. The main conclusions are as follows:

- Improving the horizontal resolution of WRF to 300 m, updating the resolution of topographic elevation to ~100 m, and turning on the topographic wind correction can effectively improve the simulation results of wind direction, but there is no obvious improvement in wind speed.
- 2. CALMET can further improve the simulation results of wind speed and direction. After updating the land cover types, the wind speed and direction improve more obviously under complex topographies. Mt. LS mainly blocks the upslope wind, and the ridge wind is weakened after "removing" Mt. LS. The Yangtze River mainly blocks the transport of PM_{2.5} from urban areas to the northwest of the river. After "removing" the Yangtze River, the transport channel from the south to the north of the river is more obvious at the south of Baguazhou where the Yangtze River splits.
- 3. According to the simulation results of PM_{2.5} transport of virtual point sources, Mt. LS acts as a barrier, blocking PM_{2.5} diffusion and forcing PM_{2.5} transport to the south or southeast, and its influence on atmospheric PM_{2.5} level caused by PM_{2.5} emitted from Gulou District can reach 55%. The influence of the Yangtze River on PM_{2.5} transport is relatively divergent and has no obvious characteristics.

This study shows that the complex topographies have an obvious effect on the local wind field and the transport of atmospheric $PM_{2.5}$. In the process of simulating local $PM_{2.5}$, it is important to consider the influence of local topographies and land cover types on wind field. In addition, the observation of wind at meteorological Station PK basically reflects the influence of Mt. LS on the local wind field, but it is quite different from the dominant wind direction in Nanjing.

Supplementary Materials: The following supporting information can be downloaded at: https: //www.mdpi.com/article/10.3390/atmos14050761/s1, Text S1: The impacts of Mt. LS and the Yangtze River on PM2.5 transport; Text S2: Comparisons of simulated and observed wind direction from different experiments; Text S3: The PM2.5 concentration and wind differences between WRF1000_CALMET1000 and WRF300_CALMET150_LU from experiments with three different virtual emission point sources; Text S4: comparisons of simulated and observed wind from different experiments at Liuhe Station; Figure S1: Spatial distributions of PM2.5 concentrations and the differences between before and after Mt. LS is "removed" from experiments with different point sources (ug/m³); Figure S2: Similar to Figure S1, but for the "removal" of the Yangtze River; Figure S3: Comparisons of simulated and observed wind direction from different experiments: WRF1000_CALMET150; WRF300_CALMET150; WRF300_CALMET150_LU; Figure S4: The PM2.5 concentration and wind differences between WRF1000_CALMET1000 and WRF300_CALMET150_LU from experiments with three different virtual emission point sources (ug/m³): Gulou District; Jiangbei Chemical Industry Park; Qixia District; Table S1: Verified wind speed of Liuhe Station; Table S2: Verified wind direction of Liuhe Station; Table S3: The symbols and units used in this manuscript.

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Data Availability Statement: The meteorological observation data used in this study are the hourly observation data of wind speed and direction from the national meteorological stations PK and LH in Nanjing obtained from the China Meteorological Data Service Centre of the National Meteorological Information Centre (http://data.cma.cn/dataService/cdcindex/datacode/A.0012.0001/show_value/normal.html, access on 21 April 2023). The initial and boundary conditions for WRF simulation are provided by the Global Data Assimilation System (GDAS/FNL, DOI: 10.5065/D65Q4T4Z) with a horizontal resolution of 0.25° and a temporal resolution of 6 h.

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