

Article

Investigating the Effect of Different Meteorological Conditions on MAX-DOAS Observations of NO₂ and CHOCHO in Hefei, China

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Received: 9 May 2019; Accepted: 22 June 2019; Published: 27 June 2019



Abstract: In this work, a ground-based remote sensing instrument was used for observation of the trace gases NO2 and CHOCHO in Hefei, China. Excessive development and rapid economic growth over the years have resulted in the compromising of air quality in this city, with haze being the most prominent environmental problem. This is first study covering observation of CHOCHO in Hefei (31.783° N, 117.201° E). The observation period of this study, i.e., July 2018 to December 2018, is divided into three different categories: (1) clear days, (2) haze days, and (3) severe haze days. The quality of the differential optical absorption spectroscopy (DOAS) fit for both CHOCHO and NO2 was low during severe haze days due to a reduced signal to noise ratio. NO₂ and CHOCHO showed positive correlations with PM2.5, producing R values of 0.95 and 0.98, respectively. NO₂ showed strong negative correlations with visibility and air temperature, obtaining R values of 0.97 and 0.98, respectively. CHOCHO also exhibited strong negative correlations with temperature and visibility, displaying R values of 0.83 and 0.91, respectively. The average concentration of NO₂, CHOCHO, and PM2.5 during haze days was larger compared to that of clear days. Diurnal variation of both CHOCHO and NO₂ showed a significant decreasing trend in the afternoons during clear days due to photolysis, while during haze days these two gases started to accumulate as their residence time increases in the absence of photolysis. There was no prominent weekly cycle for both trace gases.

Keywords: MAX-DOAS; CHOCHO; NO2; meteorological conditions; haze days

1. Introduction

China has been ranked near the bottom of the Global Environmental Sustainability Index due to its persistent and extensive air pollution [1]. Widespread air pollution in China can be attributed to a variety of factors: high output from production and manufacturing, an extensive rise in the number of automobiles, a massive economic boom, and high population growth. Poor air quality results in the degradation of human health as well as negative impacts on terrestrial ecosystems and the built environment.

Nitrogen dioxide (NO₂) and glyoxal (CHOCHO) have significant roles in varying the chemistry of the troposphere [2]. Glyoxal may originate from natural as well as anthropogenic activities [3–5]. The molecule is the smallest alpha-dicarbonyl with the highest predominance in the troposphere [6].



oxidation cycles [5,7,8] and acts as an indicator for secondary organic aerosol (SOA) formation in the atmosphere [4,7,9]. In addition, CHOCHO is also produced as an oxidation product for alkyne, isoprene, and various aromatic hydrocarbons. The concentrations of CHOCHO are not directly impacted by vehicular discharges [5] because they are only in fractions.

The formation of tropospheric ozone and the destruction of stratospheric ozone are catalyzed by nitrogen dioxide (NO₂) along with other species in the troposphere [10]. The oxidation of nitrogen dioxide forms nitric acid under favorable circumstances in the atmosphere [11]. It increases the risk and incidence of infections related to the respiratory tract [12]. Combustion of fossil fuels in urban settlements has been regarded as the main source of nitrogen dioxide [13]. Additionally, nitrogen dioxide may act as a precursor for the formation of aerosols under certain meteorological conditions. Nitrate aerosols constitute a significant proportion of fine particulates in the urban environment.

In the lower atmosphere, a substantial driving factor towards pollutant distribution in terms of chemical behavior and residence time is believed to be the meteorological condition of the locality. There exist several pieces of evidence in the literature that highlight the effect of various meteorological factors on the distribution of trace gases and aerosols in the atmosphere [14]. The overall tropospheric profile including gaseous pollutants and a variety of particles along with their meteorological parameters present a dynamic and multidimensional picture of the atmosphere, aiding towards a better understanding of pollutant sources, features, and sinks, as well as their dependence on weather conditions.

Multi-axis differential optical absorption spectroscopy (MAX-DOAS) is a type of passive spectroscopy system that has been in use to observe tropospheric trace gases over the past decade. There have been some reported studies on monitoring air quality in China using MAX-DOAS instruments. Long-term observations have resulted in the demonstration of monthly, weekly, and diurnal cycles of SO₂, NO₂, and HCHO in China, and these long-term based observations have been published in Ma et al. [15], Wang et al. [16], and Hendrick et al. [17]. Concentrations of tropospheric nitrogen dioxide were obtained using MAX-DOAS mounted at four different locations in Shanghai in 2010 during the Shanghai World Expo [18]. A MAX-DOAS system was deployed in the Eastern part of China to monitor NO₂, HCHO, and SO₂ in the Yangtze River Delta (YRD) region. Shanghai, Hefei, and Nanjing were designated for long-term monitoring. Vertical profiles and vertical column densities (VCDs) of tropospheric trace gases were monitored. Minimum values of SO₂ and NO₂ were observed at noon, whereas maxima for HCHO occurred during noontime. The broad-spectrum concentration of the pollutants was observed to gradually decrease from Shanghai to Hefei [19]. The effect of haze and non-haze conditions on the retrieval of CHOCHO from MAX-DOAS observations in Beijing has been investigated [20].

The city of Hefei, which supports a population of 683 people/km², is the capital city of Anhui Province. Excessive development and rapid economic growth over the years have resulted in the compromising of air quality in this city, with haze being the most prominent environmental problem. The prevalence of haze has dramatically impacted public health and has cast negative impressions on transportation networks and production systems. The city usually witnesses hazy days during the autumn and winter months of the year. This can be understood by taking into account the crop burning practices which occur during these months.

In addition, the location of the city makes it more prone to cold and dry air, with relatively gentle winds resulting in less chance for dust and other particulates to be diffused, causing the formation of haze.

Keeping in mind the aforementioned facts, Hefei may be considered a potential site for the observation of pollutants like NO₂ and CHOCHO. However, there have been no studies which have reported on the monitoring of CHOCHO and effects of different meteorological conditions on the retrieval of NO₂ and CHOCHO.

In this study we primarily focused on retrieval of CHOCHO and NO₂ from MAX-DOAS observations. Time series and weekly and diurnal cycles of VCDs of NO₂ and CHOCHO were generated. The dependence of CHOCHO and NO₂ VCDs on different meteorological parameters is discussed. As a quick overview of the arguments discussed above, the current research claims the unique contribution of monitoring the air quality of Hefei, China.

2. Materials and Methods

2.1. Instrument

Ground-based observations of several trace gases were carried out using a remote sensing MAX-DOAS instrument. The instrument contains a spectrometer which has a resolution of 0.6 nm and a spectral range of 300–490 nm. The major source of light employed by the instrument is dispersed sunlight, while spectral observations can be taken at different viewing angles depending on the concerned atmospheric species [21]. An inbuilt processor automatically controls the day-to-day measurements of the instrument. The spectral measurements were recorded at elevation viewing angles of 1°, 2°, 3°, 4°, 5°, 6°, 8°, 10°, 15°, 30°, and 90°. The average time span for single measurement was 60 s.

2.2. Observation Site

Anhui Province is situated in Eastern Central China with Hefei (31.783° N, 117.201° E) as the capital. The climate of the region is subtropical and humid with the existence of four distinct seasons: summer, fall, winter, and spring. The locality is dominated by southeasterly winds during summer and northwesterly during the winter season. Following a general trend, Hefei has experienced a boom in growth and development over the past two decades. At present, the city supports a permanent population of about 7.7 million. For the current study, the MAX-DOAS instrument was fitted on the building of the Hefei Environmental Protection Bureau. The observation site was almost in the center of city. Figure 1 shows the location of the observation site for MAX-DOAS measurements in Hefei, China.



Figure 1. Location of Hefei observation site for multi-axis differential optical absorption spectroscopy (MAX-DOAS) measurements in China.

2.3. DOAS Analysis

The application of DOAS provides differential slant column densities (dSCDs) for a variety of trace gases [22]. The current study analyzes MAX-DOAS spectra by employing QDOAS software developed by the Royal Belgian Institute for Space Aeronomy (BIRA-IASB) [23]. The spectra were employed to correct the measurement spectra prior to further analysis. A variety of absorption cross sections for trace gas [24–28], a low order polynomial, a Fraunhofer reference spectrum, and a Ring spectrum were included in the DOAS fit. Comprehensive details for the DOAS fit settings have been depicted in Table 1.

Parameter	Data Source	Trace Gases	
		NO ₂	СНОСНО
Wavelength		425–490 (nm)	438–457 (nm)
NO ₂	298 K, Vandaele et al. [24]	1	\checkmark
NO ₂	220 K, Vandaele et al. [25]	1	\checkmark
O ₃	223K, Serdyuchenko et al. [26]	1	\checkmark
CHOCHO	296K, Volkamer et al. [8]	х	\checkmark
O_4	293 K, Thalman and Volkamer [27]	1	\checkmark
	296 K, HITEMP (High-temperature		
H ₂ O	spectroscopic absorption parameters),	1	\checkmark
	Rothman et al. [28]		
Ring	Calculated with QDOAS	1	\checkmark
Polynomial degree		5	5

Table 1. The parameter settings used i	in MAX-DOAS observations.
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Figure 2 displays a characteristic DOAS spectral fitting of the spectrum obtained at a viewing angle of 30° on 11 November 2018. The fitting shows obvious absorption structures and low residuals, revealing that the spectral fitting is of good quality.



Figure 2. Characteristic DOAS spectral fitting of the spectrum obtained at a viewing angle of 30° on 11 November 2018 for (**a**) NO₂ and (**b**) CHOCHO. Legend: dSCD, differential slant column density.

Vertical column densities were generated using an air mass factor (AMF) [29]. For the current study, a differential air mass factor (dAMF) was applied.

$$VCD_{trop} = \frac{dSCD_{\alpha}}{dAMF_{\alpha}} \tag{1}$$

The difference in AMF between $\alpha \neq 90^{\circ}$ and $\alpha = 90^{\circ}$ is actually referred to as dAMF.

$$dAMF_{\alpha} = AMF_{\alpha} - AMF_{90^{\circ}} \tag{2}$$

$$VCD_{trop} = \frac{dSCD_{\alpha}}{AMF_{\alpha} - AMF_{oo^{\circ}}}$$
(3)

AMF can be estimated using a geometrical approximation approach [30,31].

$$AMF = \frac{1}{\sin(\alpha)} \tag{4}$$

Equation (4) then becomes:

$$VCD_{trop} = \frac{dSCD_{\alpha}}{1/sin(\alpha) - 1}$$
(5)

3. Results and Discussion

3.1. Meteorological Conditions

The observation period of this study, i.e., July 2018 to December 2018, has been divided into three different categories of meteorological conditions based on meteorological parameters. These three categories are (1) clear days, (2) haze days, and (3) severe haze days. The clear days are days with visibility greater than 10 km and a PM2.5 concentration less than 70 μ g/m³. The haze days are days with visibility less than 10 km and greater than 5 km and where the PM2.5 concentration is greater than 70 μ g/m³ and less than 115 μ g/m³. The severe haze days are days with visibility less than 5 km and a PM2.5 concentration greater than 115 μ g/m³ [32,33]. Table 2 shows a summary of different meteorological conditions. The data for meteorological parameters like temperature, humidity, and visibility was downloaded from (http://www.wunderground.com/). The data for meteorological parameters was obtained from a weather station installed at Hefei airport. The data for PM2.5 was downloaded from (http://beijingair.sinaapp.com/).

Table 2. Different meteorological conditions.

Meteorological Condition	Visibility (km)	PM2.5 (μg/m ³)
Clear days	>10	<70
Haze days	>5 and <10	>70 and <115
Severe haze days	<5	>115

3.2. Impact of Meteorological Conditions on DOAS Fit of CHOCHO and NO₂

The measurements of MAX-DOAS in the ultraviolet as well as the visible spectral range largely rely on the intensity of sunlight. The excellence of the DOAS fit is characterized by the structure of the residual left after subtracting numerous absorbers using a numerical least square fitting method [22,34] which is usually expressed using the root mean square (RMS). It is a measure of mean "instrument error" which is largely subjected to many specific parameters like limitations of the instruments (such as dark current and spectral resolution, etc.), along with limitations in illustrating the actual state of the atmosphere (for instance aerosols and scattering processes, etc.).

Figure 3 shows the average RMS and dSCD errors during different weather conditions. A higher RMS is observed during severe haze and haze days while on clear days the RMS is low. A similar trend is observed for the dSCD error with higher values during severe haze and haze days and lower values during clear days. These results can be related to the fact that during haze days the intensity of light is low, resulting in a lower signal to noise ratio. The quality of the DOAS fit is affected during severe haze days, which can be observed from the RMS and dSCD errors for both CHOCHO and NO₂. This can result in an underestimation of CHOCHO and NO₂ levels. By contrast, the RMS and dSCD errors during haze and clear days are in an acceptable range and do not have any significant effect on the levels of CHOCHO and NO₂.



Figure 3. (**a**) NO₂ RMS (**b**) NO₂ dSCD error, (**c**) CHOCHO RMS and (**d**) CHOCHO dSCD error under different weather conditions.

3.3. Time Series of NO₂, CHOCHO, and Meteorological Parameters

Observations of the trace gases NO₂ and CHOCHO were performed from July 2018 to 31 December 2018 in Hefei, China. QDOAS software was used for analysis of the spectrum to obtain dSCDs. VCDs were generated from these dSCDs using a geometric approximation approach. Data for meteorological parameters like temperature, humidity, and visibility was downloaded from (http://www.wunderground.com/). Data for PM2.5 was downloaded from (http://beijingair.sinaapp.com/). This website belongs to the national environmental monitoring network. Figure 4 shows time series of NO₂ and CHOCHO.



Figure 4. Time series of (**a**) NO₂ and (**b**) CHOCHO from 1 July 2018 to 31 December 2018. Legend: VCDs, vertical column densities.

3.4. Dependence of Trace Gases on Meteorological Parameters

In the lower atmosphere, a substantial driving factor of pollutant distribution in terms of chemical behavior and residence time is the meteorological condition of the locality. Hence, the relation of trace gases, i.e., NO₂ and CHOCHO, with different meteorological parameters is discussed here. Figure 5 shows monthly variations of PM2.5, temperature, visibility, NO₂, and CHOCHO. The average concentration of PM2.5 is seen to increase gradually from July to December. A similar trend is observed for the concentration of NO₂ and CHOCHO, whereas air temperature and visibility show a decreasing trend. The decrease in visibility due to an increase in PM2.5 is obviously to have resulted in the more frequent occurrence of haze days in the month of December.



Figure 5. Monthly variations in (a) CHOCHO, (b) NO₂, (c) temperature, (d) visibility, and (e) PM2.5.

The increasing trend of CHOCHO and NO₂ VCDs can be explained by the fact that photolysis is a major sink of both these trace gases. During the winter months when haze conditions start to occur more frequently due to an increase in PM2.5, visibility decreases. The rate of photolysis decreases, as visibility is low. This results in the accumulation of NO₂ and CHOCHO because their major sink, i.e., the rate of photolysis, is very low. The observed increase in CHOCHO from October may also be related to heating during winter. This finding is consistent with glyoxal retrieved over Beijing and Northern China from OMI (Ozone monitoring instrument) satellite data [35,36]. The increase in the concentration of NO₂ may also be related to an increase in the burning of fossil fuels during the winter months. Figure 6 shows correlation plots of NO₂ and CHOCHO with meteorological parameters.



Figure 6. Correlation plots of NO2 and CHOCHO with PM2.5, temperature, and visibility.

NO₂ shows strong negative correlations with temperature and visibility, producing R values of 0.97 and 0.98, respectively. CHOCHO also exhibits strong negative correlations with temperature and visibility, giving R values of 0.83 and 0.91, respectively. However, both CHOCHO and NO₂ show positive correlations with PM2.5, having R values of 0.95 and 0.98, respectively. Table 3 shows the average concentration of NO₂, CHOCHO, and PM2.5 during clear days, haze days, and severe haze days. It can be observed that the average concentration of NO₂ and CHOCHO is higher during haze days and heavy days compared to clear days, which is due to reduced photolytic activity, with photolytic activity acting as a major sink for NO₂ and CHOCHO. During heavy haze days the values are lower when compared with those of haze days.

The most likely reason for this trend is the fact that the quality of the DOAS fit is affected during severe haze days due to higher RMS and dSCD errors which result in the underestimation of NO_2 and CHOCHO during severe haze days.

Meteorological Condition	NO ₂ (molecules/cm ²)	CHOCHO (molecules/cm ²)	Visibility (km)	PM2.5 (μg/m ³)
Clear days	1.92×10^{16}	4.70×10^{14}	22	36
Haze days	$4.38 imes 10^{16}$	$7.62 imes 10^{14}$	8.8	103
Severe haze days	3.30×10^{16}	5.25×10^{14}	3.9	151

Table 3. Average concentrations of NO₂, CHOCHO, PM2.5, and visibility under different meteorological conditions.

3.5. Diurnal Variation

Atmospheric trace gases may come from a variety of natural sources of chemical and biological nature as well as from agricultural and industrial practices. The concentration of such gases in the atmosphere and their variation depends upon emission sources, transport routes, and removal mechanisms [37]. In urban centers, the concentration of trace gases is directly or indirectly linked to anthropogenic activities. Specific trends can be seen when looking at the diurnal cycle of these trace

gases in the atmosphere at different times of the day. The diurnal cycle is critical for understanding the atmospheric profile of these pollutants as well as vital for understanding the emission sources and atmospheric chemistry of these trace gases. The study of atmospheric trace gases can be employed to address various contemporary environmental problems [38]. In this study, the mean diurnal variation of both trace gases was calculated. Figure 7 represents the diurnal variation of CHOCHO and NO₂.



Figure 7. Diurnal cycles for (a) CHOCHO and (b) NO₂ VCDs monitored in Hefei, China.

Diurnal variation of NO₂ VCDs is observed to follow a typical pattern during clear days. Higher values occur during the morning and evening whereas lower values occur during noontime. In the morning NO₂ levels increase with a peak at around 10 a.m. local time, which can be attributed to increased traffic-related NOx emissions. The lower values during noontime can be attributed to the enhanced rate of photolysis, along with the oxidation of NO₂ by OH radicals. However, this reduction in NO₂ levels is very small because our observation period took place during the winter season, when the rate of photolysis is very low. The level of NO₂ VCDs again starts to rise during the late afternoon because rate of photolysis decreases and NO₂ begins to accumulate as the load of traffic

again increases during the evening. It is worth noting that our results are in agreement with other MAX-DOAS measurements for megacities [17–19,39].

For haze and severe haze days there is no reduction in the level of NO_2 during the afternoon. This is obviously because the major sink for NO_2 , i.e. photolysis, is absent during these days, which results in a slight increase in residence time of the trace gas. During severe haze days, a lower concentration of NO_2 is observed as compared to haze days. These results may be ascribed to an underestimation of trace gases during severe haze days due to a lower signal to noise ratio.

During clear days, due to a faster photolysis rate, a certain decrease in CHOCHO emissions is observed during the afternoon; haze and heavy haze days do not depict such a trend. CHOCHO concentrations during severe haze days are lower than during haze days, which can be explained by taking into account the fact that DOAS retrieval for CHOCHO is very sensitive and is underestimated during severe haze days.

3.6. Weekly Cycles

Human activities can be categorically divided according to weekly cycles. During weekends, because of reduced industrial activities and transport, the emissions of atmospheric pollutants normally decline [40] in comparison to week days, with peak anthropogenic activities at industrial as well as personal levels (including more use of public and private transport). This factor may have a significant effect on weekly cycles of various atmospheric species. However, in China, because of continuous industrial activity throughout the week and no formal weekly breaks, the weekend effect for various atmospheric species has been observed to be non-significant [15,18,19,41]. Figure 8 shows the results of our weekly cycle. There is no meaningful conclusion which can be drawn from these results. These results are consistent with the findings of previous studies.



Figure 8. Weekly cycles for (a) NO₂ and (b) CHOCHO VCDs monitored in Hefei, China.

4. Conclusions

In this work, MAX-DOAS observations for NO₂ and CHOCHO were performed from 1 July 2018 to 31 December 2018 in Hefei, China. Hefei is the capital of Anhui Province and has a population of around 7.80 million. Hefei is becoming a potential site for air quality monitoring. Because of the development and growth of its economy its air quality is deteriorating and haze days are occurring more frequently. There have been fairly sparse studies reporting the monitoring of CHOCHO throughout China. This is first study to observe CHOCHO in Hefei. NO₂ is well known for its significance in tropospheric chemistry. The observation time of our study was divided into three different categories based on meteorological parameters and PM2.5 levels. These categories were named clear days, haze days, and severe haze days. The excellence of the DOAS fit was not good during the severe haze days

for both CHOCHO and NO₂. RMS and dSCD errors were higher during severe haze days due to a low signal to noise ratio. This therefore resulted in the underestimation of NO₂ and CHOCHO during severe haze days. NO₂ and CHOCHO showed positive correlations with PM2.5, giving R values of 0.95 and 0.98, respectively. NO₂ showed strong negative correlations with visibility and air temperature, displaying R values of 0.97 and 0.98, respectively. CHOCHO also exhibited strong negative correlations with temperature and visibility, producing R values of 0.83 and 0.91, respectively. These outcomes can be accredited to the fact that photolysis is the main sink for both NO₂ and CHOCHO. An increase in PM2.5 concentration caused a reduction in visibility and hence resulted in the reduction of the rate of photolysis. The average concentrations of NO₂, CHOCHO, and PM2.5 during haze days were higher when compared to those of clear days. Diurnal variations of both CHOCHO and NO₂ showed significant decreasing trends during clear days due to photolysis, while during haze days the trace gases started to accumulate as the residence time of these gases increased. There was no prominent weekly cycle for both trace gases because of continuous industrial activity throughout the week and no formal weekly breaks; the weekend effect for various atmospheric species has been observed to be non-significant.

Author Contributions: Conceptualization, Z.J. and C.L.; formal analysis, Z.J.; methodology, C.L., W.T., and H.L.; resources, C.L.; validation, W.T. and C.X.; writing, review, and editing, Z.J., K.U., and C.L.

Funding: This research was funded by grants from the National Key Research and Development Program of China (2018YFC0213104), the National Natural Science Foundation of China (41722501, 91544212, 51778596, and 41575021) and the National Key Research and Development Program of China (2016YFC0203302).

Acknowledgments: We are indebted to Aimon Tanvir for their useful communications.

Conflicts of Interest: The authors declare no conflict of interest.

References

- 1. Liu, J.; Diamond, J. China's environment in a globalizing world. *Nature*. 2005, 435, 1179. [CrossRef] [PubMed]
- Hong, Q.; Liu, C.; Chan, K.L.; Hu, Q.; Xie, Z.; Liu, H.; Si, F.; Liu, J. Ship-based MAX-DOAS measurements of tropospheric NO₂, SO₂, and HCHO distribution along the Yangtze River. *Atmos. Chem. Phys.* 2018, 18, 5931–5951. [CrossRef]
- 3. Stavrakou, T.; Müller, J.F.; Smedt, I.D.; Roozendael, M.V.; Kanakidou, M.; Vrekoussis, M.; Wittrock, F.; Richter, A.; Burrows, J.P. The continental source of glyoxal estimated by the synergistic use of spaceborne measurements and inverse modelling. *Atmos. Chem. Phys.* **2009**, *9*, 8431–8446. [CrossRef]
- Fu, T.M.; Jacob, D.J.; Wittrock, F.; Burrows, J.P.; Vrekoussis, M.; Henze, D.K. Global budgets of atmospheric glyoxal and methylglyoxal, and implications for formation of secondary organic aerosols. *Geophys.Res. Atmos.* 2008, 113, D15. [CrossRef]
- Wittrock, F.; Richter, A.; Oetjen, H.; Burrows, J.P.; Kanakidou, M.; Myriokefalitakis, S.; Volkamer, R.; Beirle, S.; Platt, U.; Wagner, T. Simultaneous global observations of glyoxal and formaldehyde from space. *Geophys.Res. Lett.* 2006, *33*, 16. [CrossRef]
- 6. Myriokefalitakis, S.; Vrekoussis, M.; Tsigaridis, K.; Wittrock, F.; Richter, A.; Brühl, C.; Volkamer, R.; Burrows, J.P.; Kanakidou, M. The influence of natural and anthropogenic secondary sources on the glyoxal global distribution. *Atmos. Chem. Phys.* **2008**, *8*, 4965–4981. [CrossRef]
- Sinreich, R.; Volkamer, R.; Filsinger, F.; Frieß, U.; Kern, C.; Platt, U.; Sebastián, O.; Wagner, T. MAX-DOAS detection of glyoxal during ICARTT 2004. *Atmos. Chem. Phys.* 2007, 7, 1293–1303. [CrossRef]
- 8. Volkamer, R.; Molina, L.T.; Molina, M.J.; Shirley, T.; Brune, W.H. DOAS measurement of glyoxal as an indicator for fast VOC chemistry in urban air. *Geophys.Res. Lett.* **2005**, *32*, 8. [CrossRef]
- 9. Vrekoussis, M.; Wittrock, F.; Richter, A.; Burrows, J.P. Temporal and spatial variability of glyoxal as observed from space. *Atmos. Chem. Phys.* **2009**, *9*, 4485–4504.
- 10. Crutzen, P.J. The influence of nitrogen oxides on the atmospheric ozone content. *Q. J. R. Meteorol. Soc.* **1970**, *96*, 320–325. [CrossRef]
- 11. Seinfeld, J.H.; Pandis, S.N. *Atmospheric Chemistry and Physics: From Air Pollution to Climate Change*, 2nd ed.; John Willey & Sons. Inc.: New York, NY, USA, 2006.

- 12. Chaloulakou, A.; Mavroidis, I.; Gavriil, I. Compliance with the annual NO₂ air quality standard in Athens. Required NOx levels and expected health implications. *Atmos. Environ.* **2008**, *42*, 454–465. [CrossRef]
- 13. Noxon, J.F. Stratospheric NO₂ in the Antarctic winter. *Geophys.Res. Lett.* **1978**, *5*, 1021–1022. [CrossRef]
- Sahu, L.K.; Tripathi, N.; Yadav, R. Contribution of biogenic and photochemical sources to ambient VOCs during winter to summer transition at a semi-arid urban site in India. *Environ. Pollut.* 2017, 229, 595–606. [CrossRef] [PubMed]
- 15. Ma, J.Z.; Beirle, S.; Jin, J.L.; Shaiganfar, R.; Yan, P.; Wagner, T. Tropospheric NO₂ vertical column densities over Beijing: results of the first three years of ground-based MAX-DOAS measurements (2008–2011) and satellite validation. *Atmos. Chem. Phys.* **2013**, *13*, 1547–1567. [CrossRef]
- 16. Wang, Y.; Ying, Q.; Hu, J.; Zhang, H. Spatial and temporal variations of six criteria air pollutants in 31 provincial capital cities in China during 2013–2014. *Environ. Int.* **2014**, *73*, 413–422. [CrossRef] [PubMed]
- 17. Hendrick, F.; Müller, J.F.; Clémer, K.; Wang, P.; Mazière, M.D.; Fayt, C.; Gielen, C.; Hermans, C.; Ma, J.Z.; Pinardi, G.; et al. Four years of ground-based MAX-DOAS observations of HONO and NO₂ in the Beijing area. *Atmos. Chem. Phys.* **2014**, *14*, 765–781. [CrossRef]
- Chan, K.L.; Hartl, A.; Lam, Y.F.; Xie, P.H.; Liu, W.Q.; Cheung, H.M.; Lampel, J.; Pöhler, D.; Li, A.; Xu, J.; et al. Observations of tropospheric NO₂ using ground based MAX-DOAS and OMI measurements during the Shanghai World Expo 2010. *Atmos. Environ.* 2015, *119*, 45–58. [CrossRef]
- 19. Tian, X.; Xie, P.; Xu, J.; Li, A.; Wang, Y.; Qin, M.; Hu, Z. Long-term observations of tropospheric NO₂, SO₂ and HCHO by MAX-DOAS in Yangtze River Delta area, China. J. Environ. Sci. **2018**, 71, 207–221. [CrossRef]
- 20. Javed, Z.; Liu, C.; Khokhar, M.F.; Xing, C.; Tan, W.; Subhani, M.A.; Rehman, A.; Tanvir, A. Investigating the impact of Glyoxal retrieval from MAX-DOAS observations during haze and non-haze conditions in Beijing. *J. Environ. Sci.* **2019**. [CrossRef]
- 21. Plane, J.M.; Saiz-Lopez, A. UV-visible differential optical absorption spectroscopy (DOAS). In *Analytical Techniques for Atmospheric Measurement*; Oxford publisher: Oxford, UK, 2006; pp. 147–188.
- 22. Platt, U.; Stutz, J. Differential absorption spectroscopy. In *Differential Optical Absorption Spectroscopy*; Springer: Berlin/Heidelberg, Germany, 2008; pp. 135–174.
- 23. Danckaert, T.; Fayt, C.; Van Roozendael, M.; De Smedt, I.; Letocart, V.; Merlaud, A.; Pinardi, G. *QDOAS Software User Manual*; Belgian Institute for Space Aeronomy: Brussels, Belgium, 2013.
- 24. Vandaele, A.C.; Hermans, C.; Simon, P.C.; Carleer, M.; Colin, R.; Fally, S.; Merienne, M.F.; Jenouvrier, A.; Coquart, B. Measurements of the NO₂ absorption cross-section from 42,000 cm⁻¹ to 10,000 cm⁻¹ (238–1000 nm) at 220 K and 294 K. *J. Quant. Spectrosc. Radiat. Transf.* **1998**, *59*, 171–184. [CrossRef]
- Vandaele, A.C.; Hermans, C.; Fally, S. Fourier transform measurements of SO₂ absorption cross sections: II.: Temperature dependence in the 29,000–44,000 cm⁻¹ (227–345 nm) region. *J. Quant. Spectrosc. Radiat. Transf.* 2009, 110, 2115–2126. [CrossRef]
- 26. Serdyuchenko, A.; Gorshelev, V.; Weber, M.; Chehade, W.; Burrows, J.P. High spectral resolution ozone absorption cross-sections–Part 2: Temperature dependence. *Atmos. Meas. Tech.* **2014**, *7*, 625–636. [CrossRef]
- 27. Thalman, R.; Volkamer, R. Temperature dependent absorption cross-sections of O₂–O₂ collision pairs between 340 and 630 nm and at atmospherically relevant pressure. *Phys. Chem. Chem. Phys.* **2013**, *15*, 15371–15381. [CrossRef] [PubMed]
- Rothman, L.S.; Gordon, I.E.; Barber, R.J.; Dothe, H.; Gamache, R.R.; Goldman, A.; Perevalov, V.I.; Tashkun, S.A.; Tennyson, J. HITEMP, the high-temperature molecular spectroscopic database. *J. Quant. Spectrosc. Radiat. Transf.* 2010, 111, 2139–2150. [CrossRef]
- 29. Solomon, S.; Schmeltekopf, A.L.; Sanders, R.W. On the interpretation of zenith sky absorption measurements. *Geophys. Res. Atmos.* **1987**, *92*, 8311–8319. [CrossRef]
- 30. Hönninger, G.; Friedeburg, C.V.; Platt, U. Multi axis differential optical absorption spectroscopy (MAX-DOAS). *Atmos. Chem. Phys.* **2004**, *4*, 231–254. [CrossRef]
- 31. Celarier, E.A.; Brinksma, E.J.; Gleason, J.F.; Veefkind, J.P.; Cede, A.; Herman, J.R. Validation of ozone monitoring instrument nitrogen dioxide columns. *J. Geophys. Res. Atmos.* **2008**, *113*, D15. [CrossRef]
- 32. Zheng, G.J.; Duan, F.K.; Su, H.; Ma, Y.L.; Cheng, Y.; Zheng, B.; Zhang, Q.; Huang, T.; Kimoto, T.; Chang, D.; et al. Exploring the severe winter haze in Beijing: the impact of synoptic weather, regional transport and heterogeneous reactions. *Atmos. Chem. Phys.* **2015**, *15*, 2969–2983. [CrossRef]
- 33. Duan, L.; Xiu, G.; Feng, L.; Cheng, N.; Wang, C. The mercury species and their association with carbonaceous compositions, bromine and iodine in PM2.5 in Shanghai. *Chemosphere* **2016**, *146*, 263–271. [CrossRef]

- 34. Wagner, T.; Burrows, J.P.; Deutschmann, T.; Dix, B.; Friedeburg, C.V.; Frieß, U.; Hendrick, F.; Heue, K.P.; Irie, H.; Iwabuchi, H.; et al. Comparison of box-air-mass-factors and radiances for Multiple-Axis Differential Optical Absorption Spectroscopy (MAX-DOAS) geometries calculated from different UV/visible radiative transfer models. *Atmos. Chem. Phys.* 2007, 7, 1809–1833. [CrossRef]
- 35. Alvarado, L.M.A.; Richter, A.; Vrekoussis, M.; Wittrock, F.; Hilboll, A.; Schreier, S.F.; Burrows, J.P. An improved glyoxal retrieval from OMI measurements. *Atmos. Meas. Tech.* **2014**, *7*, 4133. [CrossRef]
- 36. Wang, Y.; Tao, J.; Cheng, L.; Yu, C.; Wang, Z.; Chen, L. A Retrieval of Glyoxal from OMI over China: Investigation of the Effects of Tropospheric NO₂. *Remote Sens.* **2019**, *11*, 137. [CrossRef]
- 37. Rasmussen, R.A.; Khalil, M.A.K. Atmospheric trace gases: trends and distributions over the last decade. *Science* **1986**, 232, 1623–1624. [CrossRef] [PubMed]
- 38. Wilkniss, P.E.; Lamontagne, R.A.; Larson, R.E.; Swinnerton, J.W.; Dickson, C.R.; Thompson, T. Atmospheric trace gases in the southern hemisphere. *Nat. Phys. Sci.* **1973**, *245*, 45–47. [CrossRef]
- Gratsea, M.; Vrekoussis, M.; Richter, A.; Wittrock, F.; Schönhardt, A.; Burrows, J.; Kazadzis, S.; Mihalopoulos, N.; Gerasopoulos, E. Slant column MAX-DOAS measurements of nitrogen dioxide, formaldehyde, glyoxal and oxygen dimer in the urban environment of Athens. *Atmos. Environ.* 2016, 135, 118–131. [CrossRef]
- 40. Cleveland, W.S.; Graedel, T.E.; Kleiner, B.; Warner, J.L. Sunday and workday variations in photochemical air pollutants in New Jersey and New York. *Science* **1974**, *186*, 1037–1038. [CrossRef] [PubMed]
- 41. Beirle, S.; Platt, U.; Wenig, M.; Wagner, T. Weekly cycle of NO₂ by GOME measurements: a signature of anthropogenic sources. *Atmos. Chem. Phys.* **2003**, *3*, 2225–2232. [CrossRef]



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