

## Article

# Effects of Short-Term Tillage Managements on CH<sub>4</sub> and N<sub>2</sub>O Emissions from a Double-Cropping Rice Field in Southern of China

Haiming Tang \* , Chao Li, Lihong Shi, Kaikai Cheng, Li Wen, Weiyan Li and Xiaoping Xiao

Farming Ecology Department, Hunan Soil and Fertilizer Institute, Changsha 410125, China; hnchaoli0419@163.com (C.L.); shilihong1981@hunaas.cn (L.S.); chengkaikai@stu.hunau.edu.cn (K.C.); wenli@hunaas.cn (L.W.); lwy86810@163.com (W.L.); hntfsxxping@163.com (X.X.)

\* Correspondence: tanghaiming66@163.com; Tel.: +86-731-8469-6102; Fax: +86-731-8469-1581

**Abstract:** Soil carbon (C) content plays an important role in maintaining or increasing soil quality and soil fertility. However, the impacts of different tillage and crop residue incorporation managements on greenhouse gas (GHG) emissions from paddy fields under the double-cropping rice (*Oryza sativa* L.) system in southern China still need further study. Therefore, a field experiment was conducted to determine the impacts of different short-term (5-years) tillage and crop residue incorporation managements on soil organic carbon (SOC) content, SOC stock, and GHG emissions from paddy fields under the double-cropping rice system in southern China. The field experiment included four tillage treatments: rotary tillage with all crop residues removed as a control (RTO), conventional tillage with crop residue incorporation (CT), rotary tillage with crop residue incorporation (RT), and no-tillage with crop residue retention (NT). These results indicated that SOC stock in paddy fields with CT, RT, and NT treatments increased by 4.64, 3.60, 3.50 Mg ha<sup>-1</sup> and 4.68, 4.21, and 4.04 Mg ha<sup>-1</sup> in 2019 and 2020, respectively, compared with RTO treatment. The results showed that early rice and late rice yield with CT treatment increased by 7.22% and 19.99% in 2019 and 6.19% and 6.40% in 2020, respectively, compared with RTO treatment. A two-year (2019–2020) investigation of GHG results indicated that methane emissions from paddy fields with NT treatment were decreased, but nitrous oxide emissions from paddy fields were increased. The lowest mean global warming potential (GWP) and per yield GWP carbon dioxide were found with NT treatment, compared to RT and CT treatments. Therefore, it was a beneficial practice for maintaining SOC stock and decreasing GHG mitigation under the double-cropping rice system in southern China by applying no-tillage with crop residue retention management.

**Keywords:** rice; tillage; crop residue; soil organic carbon; greenhouse gas; global warming potential



**Citation:** Tang, H.; Li, C.; Shi, L.; Cheng, K.; Wen, L.; Li, W.; Xiao, X. Effects of Short-Term Tillage Managements on CH<sub>4</sub> and N<sub>2</sub>O Emissions from a Double-Cropping Rice Field in Southern of China. *Agronomy* **2022**, *12*, 517. <https://doi.org/10.3390/agronomy12020517>

Academic Editor: Carmelo Maucieri

Received: 9 January 2022

Accepted: 14 February 2022

Published: 18 February 2022

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## 1. Introduction

Soil organic carbon (SOC) content plays an important role in cycling carbon (C) in terrestrial ecosystems and maintaining methane (CH<sub>4</sub>) and nitrous oxide (N<sub>2</sub>O) emissions from agricultural soil [1]. In a previous study, these results demonstrated that soil quality and soil fertility were mainly affected by SOC content, as higher SOC content represents a significant contribution to reducing C emission through C sequestration and greenhouse gas (GHG) emissions from agricultural soil [2]. Therefore, it was a beneficial strategy for increasing soil productivity and SOC content by reducing CH<sub>4</sub> and N<sub>2</sub>O emissions from agricultural soil.

Some results indicated that SOC content and its stock were sensitive to changes with field management practices, including crop system, tillage, crop residue, and fertilizer regime [3,4]. It has been confirmed that SOC content and its stock were increased by the combined application of tillage with crop residue incorporation management under

long-term field experiment conditions. Smith et al. [5] found that SOC stock with no-tillage (NT) treatment was higher than that of the other tillage treatments. Some results indicated that SOC content at the 0–30 cm soil layer with conventional tillage (CT) and NT treatments increased, while C storage (0–5 cm) at the surface layer with NT treatment increased [6–8]. However, Chen et al. [9] indicated that SOC stock at the 0–20 cm soil layer was not significantly impacted by tillage and crop residue management. However, the impacts of different short-term tillage treatments on SOC content and its stock in paddy fields still need further study, including NT, CT, RT, and crop residue return to paddy fields.

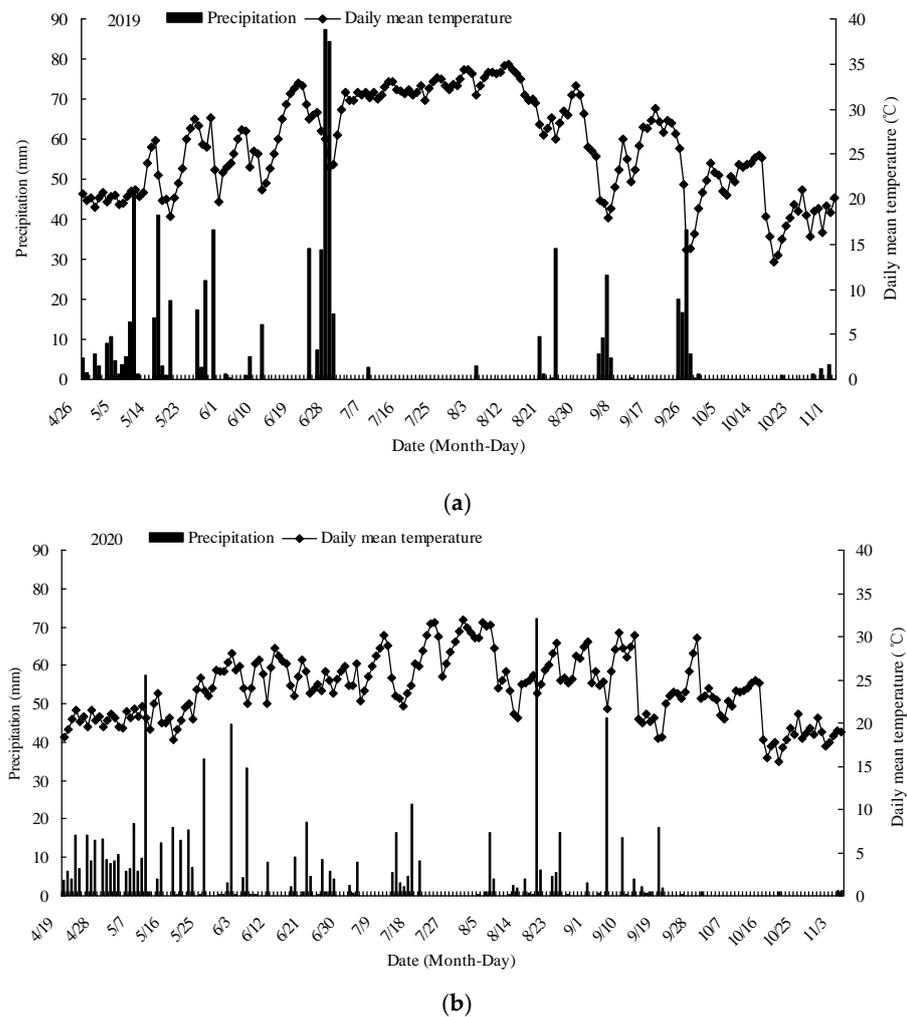
GHG emissions from agricultural soil mainly included CH<sub>4</sub>, N<sub>2</sub>O, and carbon dioxide (CO<sub>2</sub>). Some results demonstrated that CH<sub>4</sub> and N<sub>2</sub>O emissions from paddy fields were obviously changed under different tillage and crop residue return conditions [9,10]. NT was a beneficial practice in increasing SOC stock and reducing GHG emissions from paddy fields [5]. Soil C sequestration and crop yield were improved under the incorporation of crop residue conditions [11]. However, some studies indicated that the benefits of increasing SOC stock and reducing GHG emissions from paddy fields were limited with the application of no-tillage and crop residue practices. Dendooven et al. [12] showed that SOC stock and GHG emissions were not obviously changed under tillage with crop residue incorporation conditions. Furthermore, some results indicated that N<sub>2</sub>O emission from paddy fields with CT treatment was lower than that of NT treatment with crop residue incorporation [10,13]. Tang et al. [14] showed that rice yield was decreased with NT practice, compared with CT and RT practices. Global warming potential (GWP) was generally considered a vital indicator for the effects of GHG emissions, with the GWP of CO<sub>2</sub> defined as 1. There was a close relationship between decreasing GWP emissions and enhancing rice yield in paddy ecosystems [15]; some results suggested that GWP of GHG from paddy fields was obviously promoted by tillage with crop residue incorporation management [10,12].

Double-cropping rice (*Oryza sativa* L.) planting systems (early rice and late rice) are mainly cropping systems in southern China [16]. It is generally believed that tillage (CT, RT, and NT) and crop residue practices play an important role in maintaining or increasing soil quality and soil fertility in paddy fields. In our previous study, the results showed that soil properties were obviously influenced by the combined application of short-term tillage with crop residue management, including soil pH, soil bulk density, and soil microbes, which in return influence SOC content and rice yield [17]. However, related information about the effects of different tillage with crop residue incorporation treatments on SOC stock and GHG emissions from the double-cropping rice field in southern China needed further study. Therefore, a field experiment including different tillage with crop residue incorporation treatments was set up in southern China. The object of this study was: (1) to explore the impacts of different short-term tillage managements on SOC content and SOC stock under a double-cropping rice system; and (2) to measure the characteristics of CH<sub>4</sub> and N<sub>2</sub>O emissions from paddy fields, per yield GWP with different tillage managements in southern China.

## 2. Materials and Methods

### 2.1. Sites and Cropping System

This field experiment was located in Ningxiang City (28°07' N, 112°18' E) of Hunan Province, China, and began in November 2015. The cropping system of the field experiment was Chinese milk vetch (*Astragalus sinicus* L.), early rice, and late rice (*Oryza sativa* L.). The daily precipitation and daily mean temperature of the paddy field during the experimental period are shown in Figure 1. The type of soil in the paddy field was Stagnic Anthrosols, and it was developed from Quaternary red earth. Soil physicochemical characteristics at plough layer (0–20 cm) before this field experiment were as follows: total nitrogen (N) 2.14 g kg<sup>-1</sup>, available N 192.20 mg kg<sup>-1</sup>, total phosphorous (P) 0.82 g kg<sup>-1</sup>, available P 13.49 mg kg<sup>-1</sup>, total potassium (K) 13.21 g kg<sup>-1</sup>, available K 81.91 mg kg<sup>-1</sup>, soil organic carbon (SOC) 22.07 g kg<sup>-1</sup>.



**Figure 1.** Daily precipitation and daily mean temperature of paddy field during experimental period. (a) was the 2019, (b) was the 2020.

## 2.2. Experimental Design

This field experiment included four tillage treatments: rotary tillage with all crop residues removed as a control (RTO), conventional tillage with crop residue incorporation (CT), rotary tillage with crop residue incorporation (RT), and no-tillage with crop residue retention (NT). Each tillage treatment was laid out in a random complete block design with three repeats. The other more detail related information about cropping system, tillage management, applied with total number of Chinese milk vetch and rice straw, inorganic fertilizer, dates of transplanting and harvesting of rice were described according to Tang et al. [17].

## 2.3. Soil Sample and Greenhouse Gas Collection

The field experiments were conducted by combined application of tillage with crop residue incorporation management from 2015 to 2019. Soil samples were collected from paddy fields at maturity stages of late rice in November 2019 and 2020. Six soil cores were collected from the paddy field in the form of a composite soil sample, and three repeats were collected from each tillage treatment at the time of soil sample collection. Organic material, small stone, and rice roots were removed from the soil, and soil samples were air dried and sieved through a 0.15 mm mesh to investigate the SOC content [18].

N<sub>2</sub>O and CH<sub>4</sub> gas samples were collected using the static chamber technique at 09:00–11:00 during the early rice and late rice whole growth periods. N<sub>2</sub>O and CH<sub>4</sub> gas

samples were collected weekly after early or late rice seedling transplant to paddy fields, respectively. More detailed information about greenhouse gas sample collection can be found in Zhang et al. [10].

## 2.4. Laboratory Analysis

### 2.4.1. Soil Bulk Density

Soil bulk density (BD) at 0–20 cm of paddy field were conducted at maturity stages of late rice in November 2019 and 2020. The soil BD of the paddy field was investigated using metallic cores of a known volume, and the investigation of soil BD was performed according to the method described by Blake and Hartge [19].

### 2.4.2. SOC and SOC Stock

The SOC content of each soil sample was investigated using the rapid titration method. More detailed information about this method was described by Ellert and Bettany [20].

SOC stock at 0–20 cm of the paddy field was calculated by multiplying soil BD, SOC content, and thickness of the soil layer, and the units were expressed as Mg ha<sup>-1</sup>. The SOC stock of the paddy field was calculated using the following equivalent:

$$M_e = M_s \times conc \times 0.001 \quad (1)$$

$$M_{s_i} = \rho_{b_i} \times T_i \times 10,000 \quad (2)$$

where  $M_e$  was SOC stock (Mg ha<sup>-1</sup>),  $conc$  was SOC content at 0–20 cm (kg Mg<sup>-1</sup>),  $M_{s_i}$  was soil mass at 0–20 cm (kg ha<sup>-1</sup>),  $\rho_{b_i}$  was soil BD at 0–20 cm (Mg·m<sup>-3</sup>),  $T_i$  was depth of soil layer (m); 10,000 was the coefficient that area units of m<sup>2</sup> converted into ha, 0.001 was the coefficient that mass units of kg converted into Mg.

### 2.4.3. N<sub>2</sub>O and CH<sub>4</sub> Emissions

CH<sub>4</sub> and N<sub>2</sub>O emissions flux and cumulative were investigated according to the method described by Zhang et al. (2013) [10]. Briefly, N<sub>2</sub>O and CH<sub>4</sub> emissions flux of gas samples were investigated with a gas chromatograph equipped with an electron capture detector (ECD) and flame ionization detector (FID). N<sub>2</sub>O and CH<sub>4</sub> emissions flux were investigated using a stainless-steel column, Porapak Q (80/100 mesh), and 13XMS column (60/80 mesh), with ECD and FID at 330 °C and 200 °C, respectively. The GWP of CH<sub>4</sub> and N<sub>2</sub>O from the paddy field were calculated according to the method described by Thelen et al. [21].

## 2.5. Rice Yield

At maturity stages of early rice and late rice, rice yield with all tillage treatments was investigated in each plot.

## 2.6. Statistical Analysis

All data in the present manuscript were expressed as the mean ± standard error. All statistical analysis for each item with all tillage treatments were conducted using SPSS statistical software (v3.11). The data of each investigated item with different tillage treatments were compared by using one-way analysis of variance (ANOVA) following the standard procedure at the 5% probability level.

## 3. Results

### 3.1. SOC Content and SOC Stock

Soil bulk density (BD) of paddy fields with CT, RT, NT, and RTO treatments ranged from 1.10 to 1.22 g cm<sup>-3</sup>, and this result showed that soil BD with NT treatment was significantly higher ( $p < 0.05$ ) than that of CT and RT treatments in 2019 and 2020 (Table 1). Compared with NT treatment, soil BD with CT and RT treatments were significantly decreased (1.10 and 1.10 g m<sup>-3</sup> for CT, 1.14 and 1.15 g m<sup>-3</sup> for RT in 2019 and 2020,

respectively). There was no significant difference ( $p > 0.05$ ) in soil BD between NT and RTO treatments in either year.

**Table 1.** Impacts of different short-term tillage managements on soil BD, SOC, and its stock of paddy field under the double-cropping rice system (0–20 cm).

Years	Treatments	BD (g cm <sup>-3</sup> )	SOC (g kg <sup>-1</sup> )	SOC Stock (Mg ha <sup>-1</sup> )
2019	CT	1.10 ± 0.03 b	23.54 ± 0.68 a	51.79 ± 1.50 a
	RT	1.14 ± 0.04 b	22.26 ± 0.64 ab	50.75 ± 1.47 ab
	NT	1.21 ± 0.04 a	20.93 ± 0.61 b	50.65 ± 1.46 ab
	RTO	1.17 ± 0.03 ab	20.15 ± 0.58 b	47.15 ± 1.36 b
2020	CT	1.10 ± 0.03 b	23.59 ± 0.67 a	51.90 ± 1.48 a
	RT	1.15 ± 0.03 b	22.36 ± 0.63 ab	51.43 ± 1.46 ab
	NT	1.22 ± 0.04 a	21.01 ± 0.60 b	51.26 ± 1.45 ab
	RTO	1.17 ± 0.03 ab	20.18 ± 0.56 c	47.22 ± 1.35 b

CT: conventional tillage with crop residue incorporation; RT: rotary tillage with crop residue incorporation; NT: no-tillage with crop residue retention; RTO: rotary tillage with all crop residues removed as a control. BD: soil bulk density; SOC: soil organic carbon. Different lowercase letters in the same column indicated a significant difference at the 0.05 level.

The SOC content and SOC stock of paddy fields were obviously changed under short-term (5-year) continuous crop residue incorporation conditions (Table 1). These results indicated that SOC content and SOC stock of paddy field with CT, RT, and NT treatments were significantly higher ( $p < 0.05$ ) than that of RTO treatment, storing as much as 3.60 to 4.68 Mg ha<sup>-1</sup> more C than that of the RTO treatment in 2019 and 2020. Meanwhile, CT treatment significantly increased SOC content and SOC stock in the paddy field. These results indicated that SOC content and SOC stock of paddy field with CT, RT, and NT treatments were increased, but there was no significant difference ( $p > 0.05$ ) in SOC stock between RTO treatment and RT and NT treatments in 2019 and 2020.

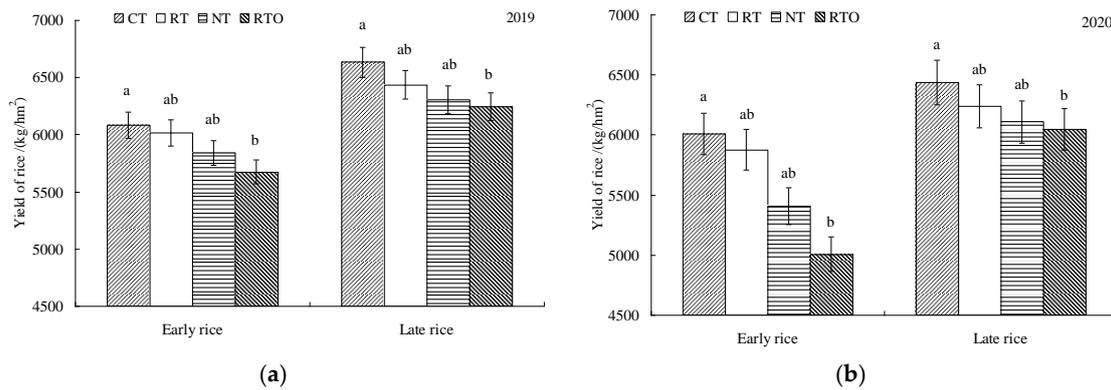
### 3.2. Early Rice and Late Rice Yield

The results indicated that early rice and late rice yield with RTO treatment were lower than that of CT, RT, and NT treatments, and double-cropping rice yield with RTO treatment was significantly lower ( $p < 0.05$ ) than that of CT treatment. There was no significant difference ( $p > 0.05$ ) in double-cropping rice yield between NT treatment and CT and RT treatments in 2019 and 2020 (Figure 2). Compared to RTO treatment, early rice and late rice yield with CT treatment increased by 7.22% and 19.99% in 2019 and 6.19% and 6.40% in 2020. The results indicated that double-cropping rice yield with RT and NT treatments was increased, but there was no significant difference ( $p > 0.05$ ) in double-cropping rice yield between RT, NT, and RTO treatments.

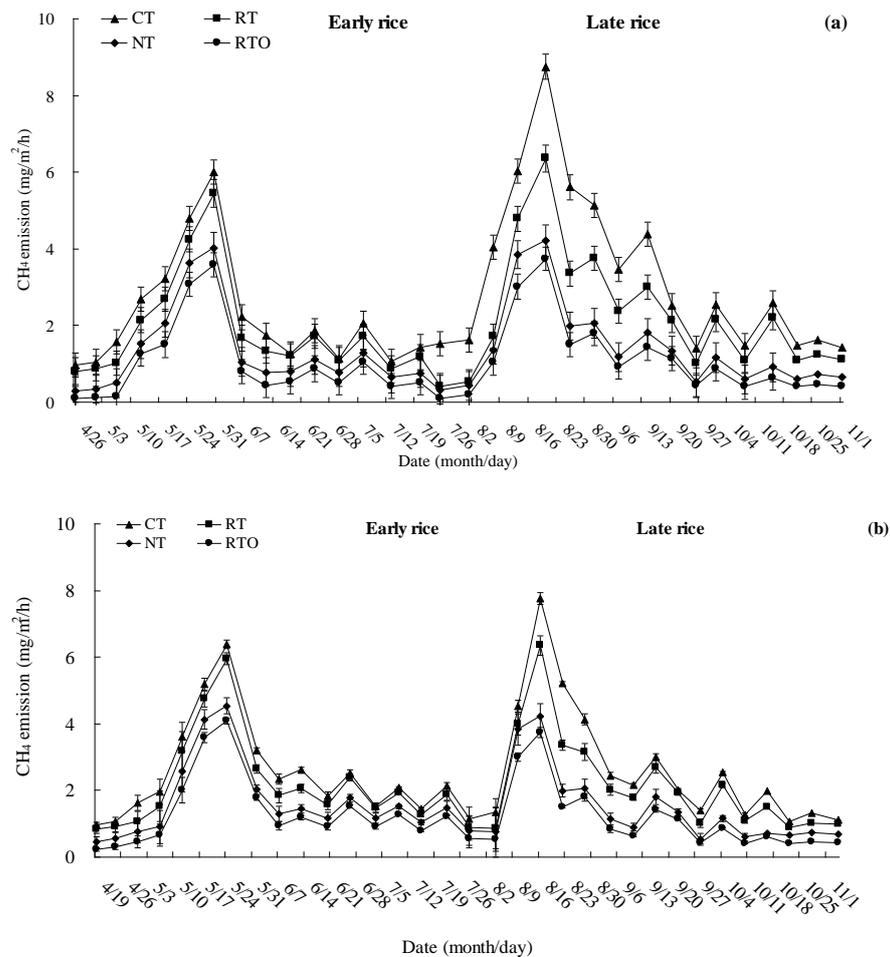
### 3.3. CH<sub>4</sub> Emission Flux

These results indicated that CH<sub>4</sub> emission flux from paddy field with RTO treatment was significantly lower than that of CT, RT, and NT treatments. During the early rice growth stage, CH<sub>4</sub> emission flux was lower at the early stage, but CH<sub>4</sub> emission flux peaked at 27 d and 28 d after rice seedling transplanting to paddy fields in 2019 and 2020, respectively, and then dropped to a lower level (Figure 3a). During the early rice growth stage, the order of CH<sub>4</sub> emission flux from paddy field with all tillage treatments was as follows: RTO < NT < RT < CT.

During the late rice growth stage, characteristics of CH<sub>4</sub> emission flux from paddy fields with all tillage treatments were similar to the early rice growth stage. CH<sub>4</sub> emission flux from the paddy field was lower following rice seedling transplant to the paddy field, but CH<sub>4</sub> emission flux peaked at 25 d and 23 d after rice seedling transplant to the paddy field in both years, respectively. These results showed that the order of CH<sub>4</sub> emission flux from the paddy field with all tillage treatments was as follows: RTO < NT < RT < CT (Figure 3b).



**Figure 2.** Impacts of different short-term tillage managements on early rice and late rice yield under the double-cropping rice system. CT: conventional tillage with crop residue incorporation; RT: rotary tillage with crop residue incorporation; NT: no-tillage with crop residue retention; RTO: rotary tillage with all crop residues removed as a control. Error bars represent standard error of the mean. Different lowercase letters indicated a significant difference at the 0.05 level. The same as below. (a) was the 2019, (b) was the 2020.

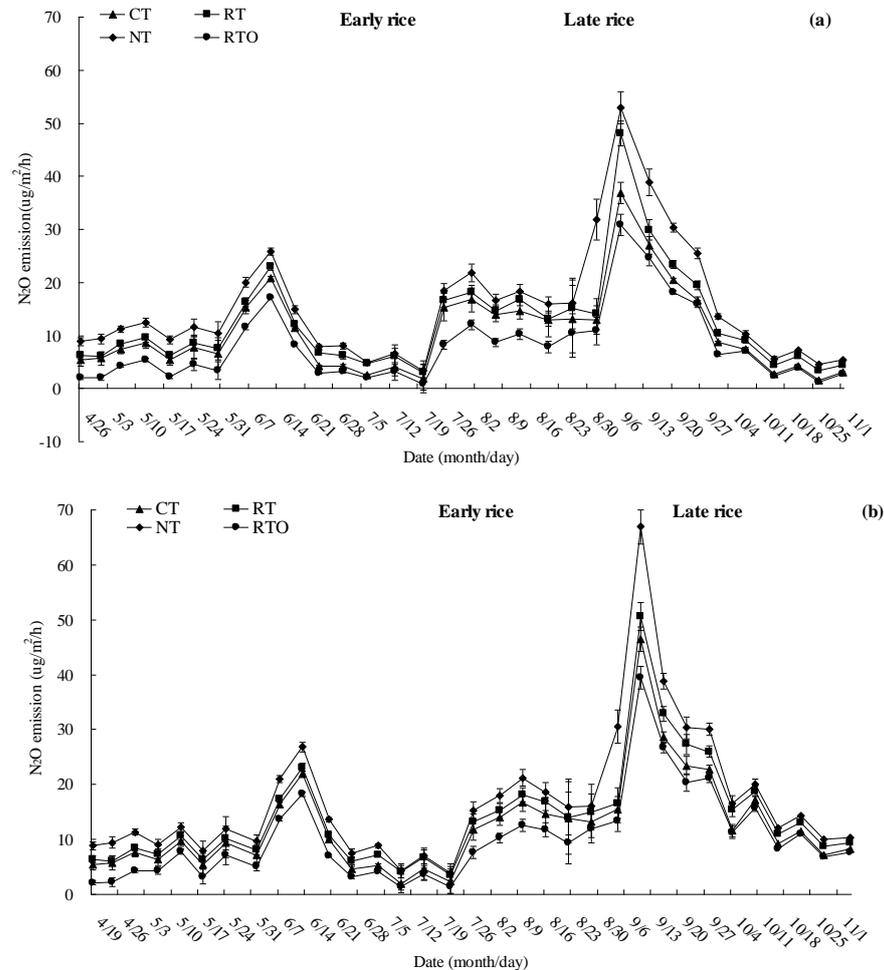


**Figure 3.** Impacts of different short-term tillage managements on CH<sub>4</sub> emission flux from double-cropping rice field in 2019 (a) and 2020 (b).

### 3.4. N<sub>2</sub>O Emission Flux

These results showed that N<sub>2</sub>O emission flux from the paddy field increased after rice seedling transplant to the paddy field; the highest peak of N<sub>2</sub>O emission flux from the

paddy field was observed at the aeration stage (Figure 4). These results indicated that the order of  $N_2O$  emission flux from the paddy field with all tillage treatments was as follows:  $NT > RT > CT > RTO$  (Figure 4a). In 2019, the average  $N_2O$  emission flux from the paddy field with CT, RT, NT, and RTO treatments was  $7.42$ ,  $8.71$ ,  $10.96$ , and  $4.85 \mu g m^{-2} h^{-1}$ , respectively. In 2020, the average  $N_2O$  emission flux from paddy fields with all tillage treatments was  $8.11$ ,  $9.26$ ,  $11.32$ , and  $5.85 \mu g m^{-2} h^{-1}$ , respectively.



**Figure 4.** Impacts of different short-term tillage managements on  $N_2O$  emission flux from double-cropping rice field in 2019 (a) and 2020 (b).

These results indicated that the highest peak of  $N_2O$  emission flux from the paddy field was observed at 42 d and 45 d after rice seedling transplanting to the paddy field in both years, respectively (Figure 4b). In 2019, the average  $N_2O$  emission flux from paddy field with CT, RT, NT, and RTO treatments was  $13.44$ ,  $15.69$ ,  $19.59$ , and  $10.69 \mu g m^{-2} h^{-1}$ , respectively. In 2020, the average  $N_2O$  emission flux from the paddy field with all tillage treatments was  $16.01$ ,  $18.10$ ,  $21.60$ , and  $13.71 \mu g m^{-2} h^{-1}$ , respectively.

### 3.5. Cumulative Emissions of $N_2O$ and $CH_4$ from Paddy Field

This result indicated that cumulative  $CH_4$  emission from paddy fields with RT, NT, and RTO treatments was significantly lower ( $p < 0.05$ ) than that of the CT treatment. The order of cumulative  $CH_4$  emission from the paddy field with all tillage treatments (average of two years) were as follows:  $CT (12.03 g m^{-2}) > RT (9.52 g m^{-2}) > NT (6.52 g m^{-2}) > RTO (5.14 g m^{-2})$ . Cumulative  $CH_4$  emission from paddy field with CT, RT, and NT treatments (average of two years) increased by 137.28%, 86.67%, and 27.14%, respectively, compared with RTO treatment (Table 2).

**Table 2.** Impacts of different short-term tillage managements on CH<sub>4</sub> and N<sub>2</sub>O emissions from double-cropping rice field (g m<sup>-2</sup>).

Years	Treatments	CH <sub>4</sub>			N <sub>2</sub> O		
		Early Rice	Late Rice	Total	Early Rice	Late Rice	Total
2019	CT	4.50 ± 0.12 a	7.69 ± 0.22 a	12.19 ± 0.35 a	0.015 ± 0.001 c	0.032 ± 0.001 c	0.047 ± 0.002 c
	RT	3.81 ± 0.11 b	5.31 ± 0.15 b	9.12 ± 0.26 b	0.018 ± 0.001 b	0.037 ± 0.001 b	0.055 ± 0.002 b
	NT	2.66 ± 0.08 c	3.28 ± 0.10 c	5.95 ± 0.17 c	0.022 ± 0.001 a	0.046 ± 0.001 a	0.068 ± 0.002 a
	RTO	2.03 ± 0.06 d	2.56 ± 0.07 d	4.59 ± 0.13 d	0.010 ± 0.001 d	0.025 ± 0.001 d	0.035 ± 0.001 d
2020	CT	5.35 ± 0.15 a	6.53 ± 0.18 a	11.87 ± 0.34 a	0.017 ± 0.001 c	0.040 ± 0.001 c	0.057 ± 0.002 c
	RT	4.66 ± 0.11 b	5.26 ± 0.15 b	9.92 ± 0.28 b	0.019 ± 0.001 b	0.046 ± 0.001 b	0.065 ± 0.002 b
	NT	3.53 ± 0.10 c	3.55 ± 0.10 c	7.08 ± 0.21 c	0.024 ± 0.001 a	0.054 ± 0.001 a	0.078 ± 0.002 a
	RTO	2.87 ± 0.08 d	2.81 ± 0.08 d	5.68 ± 0.16 d	0.012 ± 0.001 d	0.035 ± 0.001 d	0.047 ± 0.001 d

RTO: rotary tillage with all crop residues removed as a control; CT: conventional tillage with crop residue incorporation; RT: rotary tillage with crop residue incorporation; NT: no-tillage with crop residue retention. Different lowercase letters in the same column indicated a significant difference at the 0.05 level.

These results demonstrated that cumulative N<sub>2</sub>O emission from paddy fields with CT, RT, and RTO treatments were significantly lower ( $p < 0.05$ ) than that of NT treatment, during early rice and late rice whole growth stages. The results indicated that the order of cumulative N<sub>2</sub>O emission from the paddy field with all tillage treatments (average of two years) were NT (0.073 g m<sup>-2</sup>) > RT (0.060 g m<sup>-2</sup>) > CT (0.052 g m<sup>-2</sup>) > RTO (0.041 g m<sup>-2</sup>). Cumulative N<sub>2</sub>O emission from paddy field with CT, RT, and NT treatments (average of two years) increased by 27.78%, 47.72%, and 80.12%, compared to RTO treatment, respectively (Table 2).

### 3.6. Comprehensive GWP of CH<sub>4</sub> and N<sub>2</sub>O from Paddy Field

According to GWP, the contribution of CH<sub>4</sub> emission to global warming was higher than that of N<sub>2</sub>O emission from paddy fields (Table 3). These results indicated that the GWP of CH<sub>4</sub> emission from the paddy field with CT treatment was significantly higher ( $p < 0.05$ ) than that of RT, NT, and RTO treatments, and with the trend CT > RT > NT > RTO in both years. Meanwhile, the results showed that GWP of N<sub>2</sub>O emission from paddy field with CT, RT, and RTO treatments were significantly lower ( $p < 0.05$ ) than that of NT treatment, and with the trend NT > RT > CT > RTO in both years.

**Table 3.** GWP of CH<sub>4</sub> and N<sub>2</sub>O emissions from paddy field and per yield GWP with different short-term tillage managements.

Years	Treatments	CH <sub>4</sub> Emission (g m <sup>-2</sup> )	N <sub>2</sub> O Emission (g m <sup>-2</sup> )	GWP of CH <sub>4</sub> (kg CO <sub>2</sub> ha <sup>-1</sup> )	GWP of N <sub>2</sub> O (kg CO <sub>2</sub> ha <sup>-1</sup> )	GWP of CH <sub>4</sub> and N <sub>2</sub> O (kg CO <sub>2</sub> ha <sup>-1</sup> )	Early and Late Rice Grain Yield (kg ha <sup>-1</sup> )	Per yield GWP CO <sub>2</sub> (kg kg <sup>-1</sup> )
2019	CT	12.19 ± 0.35 a	0.047 ± 0.002 c	3051.77 ± 88.10 a	140.26 ± 5.85 c	3192.02 ± 92.14 a	12,718.5 ± 367.2 a	0.25 ± 0.01 a
	RT	9.12 ± 0.26 b	0.055 ± 0.002 b	2283.19 ± 65.91 b	164.13 ± 4.73 b	2447.32 ± 70.64 b	12,454.5 ± 359.5 ab	0.20 ± 0.01 b
	NT	5.95 ± 0.17 c	0.068 ± 0.002 a	1489.58 ± 43.01 c	202.92 ± 4.04 a	1692.51 ± 48.85 c	12,150.0 ± 350.7 ab	0.14 ± 0.01 c
	RTO	4.59 ± 0.13 d	0.035 ± 0.001 d	1149.11 ± 33.17 d	104.45 ± 3.01 d	1253.55 ± 36.18 d	11,922.0 ± 344.2 b	0.11 ± 0.01 d
2020	CT	11.87 ± 0.34 a	0.057 ± 0.002 c	2971.65 ± 85.78 a	170.10 ± 5.74 c	3141.75 ± 90.01 a	12,343.9 ± 356.3 a	0.25 ± 0.01 a
	RT	9.92 ± 0.28 b	0.065 ± 0.002 b	2483.47 ± 71.69 b	193.97 ± 4.93 b	2677.44 ± 76.68 b	12,112.6 ± 349.7 ab	0.22 ± 0.01 b
	NT	7.08 ± 0.21 c	0.078 ± 0.002 a	1772.48 ± 51.16 c	232.77 ± 4.22 a	2005.24 ± 57.36 c	11,515.9 ± 332.4 ab	0.17 ± 0.01 c
	RTO	5.68 ± 0.16 d	0.047 ± 0.001 d	1421.99 ± 41.04 d	140.26 ± 3.19 d	1562.24 ± 44.23 d	11,055.8 ± 319.2 b	0.14 ± 0.01 d

RTO: rotary tillage with all crop residues removed as a control; CT: conventional tillage with crop residue incorporation; RT: rotary tillage with crop residue incorporation; NT: no-tillage with crop residue retention. Different lowercase letters in the same column indicated a significant difference at the 0.05 level.

The order of the GWP of N<sub>2</sub>O and CH<sub>4</sub> emissions from paddy fields with all tillage treatments (average of two years) were as follows: CT (3166.89 kg CO<sub>2</sub>-equivalent ha<sup>-1</sup>) > RT (2562.38 kg CO<sub>2</sub>-equivalent ha<sup>-1</sup>) > NT (1848.88 kg CO<sub>2</sub>-equivalent ha<sup>-1</sup>) > RTO (1407.90 kg CO<sub>2</sub>-equivalent ha<sup>-1</sup>). Compared with RTO treatment, GWP of N<sub>2</sub>O and CH<sub>4</sub> emissions from the paddy field with CT, RT, and NT treatments (average of two years) increased by 127.87%, 83.31%, and 31.69%, respectively (Table 3). Meanwhile, the results

showed that per yield GWP CO<sub>2</sub> with RT, NT, and RTO treatments were significantly lower ( $p < 0.05$ ) than that of CT treatment and had the trend CT > RT > NT > RTO in both years.

#### 4. Discussion

##### 4.1. Effects of Tillage Management on SOC Content and SOC Stock

These results indicated that soil BD was obviously changed under different short-term tillage conditions, especially when combined with crop residue incorporation. Soil BD of double-cropping rice fields decreased with crop residue incorporation management. This may be attributed to the fact that the formation of macro-aggregates and macro-pores was promoted with the cementation of secreted polysaccharides and organic acids by soil microorganisms during the process of decomposition of the applied crop residue [4,22]. This decrease in soil BD in the paddy field could reflect a higher level of rice root and rhizodeposition crop residue input under tillage practice conditions compared to those without crop residue input (Table 1). Meanwhile, these results showed that the lowest soil BD was correlated with the highest SOC content in the CT and RT treatments; NT treatment had a moderate level of SOC and the highest soil BD, and moderate soil BD and the lowest SOC content was found in the RTO treatment. Lack of crop residue input was the main reason for the increase of soil BD at the surface layer because crop residue stabilizes the soil aggregate against breaking, dispersion, and collapse [23], which agrees with these results, indicating that an increase in soil SOC content was related to a decrease in soil BD under combined tillage and crop residue incorporation conditions (Table 1).

Some results proved that SOC content and SOC stock of paddy fields were obviously influenced by different tillage practices [16,17]. In the present study, the results demonstrated that SOC content and SOC stock with NT treatment were obviously higher than that of RTO treatment. Our results were similar to the previous finding that SOC content and SOC stock with crop residue input management were increased [4,6], which were closely related to soil C content, cropping system, and crop residue management [22]. Our results proved that SOC content and SOC stock of paddy field with CT and RT treatments were higher than that of NT and RTO treatments, suggesting a higher C level, root residue, and exudates returning to the paddy field under applied crop residue conditions (Table 1). On the other hand, a lower C decomposition rate and higher organic matter incorporation were also another reason for the increased SOC stock of paddy fields [24]. Our results were in agreement with previous findings based on long-term field experiment conditions [9,17]. These results suggested that SOC content and SOC stock of paddy fields were closely correlated with an increase in soil quality, resulting in an increase in rice yield. In this study, there was a close relationship between higher rice yield and higher soil quality (SOC content and SOC stock), as expressed in paddy fields with RT and CT treatments (Figure 2). RT and CT treatments had the highest rice yield, as well as higher soil quality (higher value in SOC content and SOC stock, lower value in soil BD). Therefore, it was a beneficial strategy for sustaining or enhancing soil quality and rice yield in the double-cropping rice system with RT and CT treatments. RT and CT treatments were beneficial management for reducing soil mechanical resistance of the paddy field. However, there was the lowest rice yield and lower soil quality (e.g., lower SOC content and SOC stock, and higher soil BD) with RTO treatment. Therefore, it was an effective strategy for sustaining or enhancing soil quality and rice yield under the double-cropping rice system in southern China by combining rotary tillage and conventional tillage with crop residue input management.

##### 4.2. Effects of Tillage Management on CH<sub>4</sub> and N<sub>2</sub>O Emissions

CH<sub>4</sub> and N<sub>2</sub>O emissions from paddy fields were mainly influenced by different field practices, including tillage, fertilizer regime, crop residue, irrigation, and so on. Some results proved that CH<sub>4</sub> and N<sub>2</sub>O emissions from agricultural soil were closely related with tillage practice, CH<sub>4</sub> and N<sub>2</sub>O emissions from paddy fields were obviously enhanced under combined tillage and crop residue conditions [10]. In this study, the results showed that CH<sub>4</sub> and N<sub>2</sub>O emissions flux, cumulative CH<sub>4</sub> and N<sub>2</sub>O emissions from paddy field

with NT, RT, and CT treatments were much larger than that of RTO treatment, which was in agreement with the previous results [9]. The reason for this phenomenon may be that: (i) soil microorganism activities were promoted with crop residue returning to paddy field for that provide more carbon and energy source for soil microorganism activities; (ii) methanogens and nitrifying bacteria were also increased, which provide more substrate for CH<sub>4</sub> and N<sub>2</sub>O emissions from paddy fields. Meanwhile, these results indicated that CH<sub>4</sub> and N<sub>2</sub>O emissions from paddy fields with all tillage treatments during the early rice whole growth stage were larger than that of the late rice growth stage, suggesting that CH<sub>4</sub> and N<sub>2</sub>O production was higher than that of the CH<sub>4</sub> and N<sub>2</sub>O consumption rate under the higher temperature conditions (Figure 1), and could explain why CH<sub>4</sub> and N<sub>2</sub>O emissions from paddy fields were increased during the late rice whole growth stage than during the early rice whole growth stage. On the other hand, soil microbial activities were improved under higher temperature conditions. Furthermore, relatively high temperatures were beneficial for promoting the decomposition process of crop residue, producing a larger number of organic compounds and resulting in the promotion of CH<sub>4</sub> and N<sub>2</sub>O emissions from paddy fields.

In a previous study, the results proved that NT treatment was a beneficial soil practice for reducing CH<sub>4</sub> emission from paddy fields [25]. Our results indicated that CH<sub>4</sub> emission flux from paddy fields with NT treatment was lower than that of CT and RT treatments, which was similar to the previous findings [10]. The reason was mainly attributed to increased crop residue (e.g., rice root) incorporation into paddy field, adding tillage practice in the CT and RT plots, which would promote carbon decomposition and mineralization and provide more carbon substrates for soil microorganism activities. On the other hand, root biomass among different tillage treatments might mainly affect carbon decomposition [25]. Furthermore, NT treatment with no-tillage practice inhibited SOM decomposition and provided a lower carbon substrate for soil methanogen activities compared to conventional tillage and rotary tillage treatments [25]. In the present study, NT treatment decreased SOC content in the paddy field compared to RT and CT treatments, suggesting that CH<sub>4</sub> emission from the paddy field with no-tillage practice decreased. Some results indicated that aerobic methanotrophs and anaerobic methanogens were obviously influenced by the rice root soil environment [14]. RT and CT treatments destroyed soil structure and reduced gas diffusivity at the plough layer; therefore, soil CH<sub>4</sub> uptake and CH<sub>4</sub> oxidation were decreased [26]. However, the soil compaction increased with NT treatment, prolonging the CH<sub>4</sub> transfer access, reducing CH<sub>4</sub> emission into the atmosphere from the paddy field, and decreasing the amount of CH<sub>4</sub> transferred to the rice rhizosphere that was absorbed and emitted by rice plants [5,27]. Therefore, CH<sub>4</sub> emission from paddy fields was increased under RT and CT conditions.

The N<sub>2</sub>O emission in our study was negligible (0.035–0.078 g m<sup>-2</sup>) during early rice and late rice growth stages in 2019 and 2020. This agrees with previous findings indicating that N<sub>2</sub>O emission from paddy fields was also lower [7,18]. N<sub>2</sub>O emission from paddy fields was the result of soil anaerobic denitrification and aerobic nitrification activities [26]. Zhang et al. [10] showed that N<sub>2</sub>O emission from paddy fields with NT management were increased, which was mainly affected by anaerobic conditions related to wet and compact soil. However, our results proved that N<sub>2</sub>O emission from paddy fields with NT treatment were slightly higher than those of RT and CT treatments. The results were similar to those of previous studies in which the adoption of reduced tillage or no-tillage management increased N<sub>2</sub>O emission [5]. Higher N<sub>2</sub>O emission from paddy fields with NT treatment might be mainly attributed to an increase in soil compaction [5], a higher denitrification rate [28], and a higher soil nutrient content (e.g., N, SOC) [10]. However, some results showed no significant difference in N<sub>2</sub>O emission between NT and CT treatments [29] or more N<sub>2</sub>O emission from paddy field with conventional tillage treatment [30]. The difference in previous results might closely relate to the climate conditions and soil physicochemical characteristics of the field experiment. However, the impacts of different tillage

management on related soil anaerobic denitrifying and aerobic nitrifying bacteria in the double-cropping rice field still need further study based on long-term conditions.

Global warming potential (GWP) is usually regarded as an important factor in estimating the potential impacts of CH<sub>4</sub>, N<sub>2</sub>O, and CO<sub>2</sub> on global climate change. Chen et al. [9] proved that the GWP of CH<sub>4</sub> and N<sub>2</sub>O with NT treatment was lower than that of RT and CT treatments in a double-cropping rice field. In this study, the results showed that NT treatment was an effective practice to reduce GWP (27.97% and 41.58% compared to RT and CT treatments, respectively) from the paddy field. Ussiri et al. [26] also indicated that GWP with NT treatment was as low as 50% of that of moldboard and chisel tillage treatments. Clearly, compared with N<sub>2</sub>O emission, CH<sub>4</sub> emission was the main component of GWP in all tillage treatments, which accounted for 95.10%, 93.02%, 88.20%, and 91.35% with CT, RT, NT, and RTO treatments (average of two years), respectively (Table 3), which were similar to the report of Cheng et al. [31], confirming that CH<sub>4</sub> emission accounted for mainly the percentage of GWP from paddy fields. Therefore, it was necessary to use suitable tillage with crop residue practice to decrease CH<sub>4</sub> emission instead of N<sub>2</sub>O emission from paddy fields to reduce GWP under the double-cropping rice system. Meanwhile, our results proved that the range of per yield GWP CO<sub>2</sub> of the double-cropping rice system with all tillage treatments was from 0.11 to 0.25 kg kg<sup>-1</sup>, which were within the range of the previous study results [32]. However, the range of per yield GWP CO<sub>2</sub> in our study was lower than that of previous results [33], in which GHG emissions from paddy fields were increased by continuous waterlogging practices, while alternating wetting and drying irrigation practices were used in our study. Therefore, applying alternating wetting and drying irrigation practices might reduce GHG emissions from paddy fields [34]. At present, there is still a need to further study the impacts of different tillage treatments on per yield GWP CO<sub>2</sub> of double-cropping rice systems. In this study, our results indicated that the order of per yield GWP CO<sub>2</sub> with all tillage treatments was similar to the sequence of GWP (Table 3) and rice yield, suggesting that NT treatment was a suitable practice for reducing per yield GWP CO<sub>2</sub> of the double-cropping rice system in southern China.

## 5. Conclusions

In summary, our study indicated that SOC stock and greenhouse gas (GHG) emissions from double-cropping rice fields were strongly affected by short-term tillage management. CT treatment significantly increased SOC content and SOC stock at the 0–20 cm soil layer of double-cropping rice fields. NT treatment obviously decreased CH<sub>4</sub> emission, although it increased N<sub>2</sub>O emission from the paddy field significantly. These results showed that the GWP of CH<sub>4</sub> and N<sub>2</sub>O emissions were significantly decreased with short-term NT treatment, due to a reduction in CH<sub>4</sub> emission. Among the CT, RT, and NT treatments, per yield GWP CO<sub>2</sub> was lowest with NT treatment (0.14 and 0.17 kg kg<sup>-1</sup> in 2019 and 2020, respectively). Therefore, these results suggest that NT could serve as an effective management strategy for reducing GWP of CH<sub>4</sub> and N<sub>2</sub>O emissions from the double-cropping rice field in southern China. However, there was still a need to further explore the impacts of different tillage managements on related soil microorganism mechanisms in the double-cropping rice field based on long-term conditions.

**Author Contributions:** H.T. and X.X. designed the experiments; C.L., K.C. and L.W. performed the experiments; L.S. and W.L. analyzed the data; H.T. wrote the manuscript. All authors have read and agreed to the published version of the manuscript.

**Funding:** This work was supported by the National Natural Science Foundation of China (31872851) and Innovative Research Groups of the Natural Science Foundation of Hunan Province (2021CX36, 2021CX37).

**Institutional Review Board Statement:** Not applicable.

**Informed Consent Statement:** Not applicable.

**Data Availability Statement:** The data reported in this study is contained within the article.

**Acknowledgments:** We would like to thank the staff of Ningxiang Agricultural Technology Extension Center for field management.

**Conflicts of Interest:** The authors declare no conflict of interest.

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