



Article Concentration of Microplastics in Road Dust as a Function of the Drying Period—A Case Study in G City, Korea

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Abstract: Microplastics (MPs) are plastic particles < 5 mm in diameter, which are detected in air, soil, and water, causing various environmental problems. In total, 37.3% of MPs are generated from point pollution sources and 62.7% from non-point sources; most of the non-point-source MPs are from vehicles tires, road-marking paint, and bitumen used in road pavements. In this study, the concentrations of MPs generated from roads in Goyang city, South Korea, were examined in terms of the drying period (0, 1, 2, or 3 d). Road dust sampling was performed at the kerb and quantitative and qualitative analyses were conducted for each sample. The MP concentrations were 552 (±39) MPs g⁻¹ for a 0 d drying period and 1530 (±602) MPs g⁻¹ for a 3 d drying period, confirming that the MP concentration in road dust increased with the increase in drying period. Among the detected substances, black particles accounted for the highest proportion (72%) and were found to be bitumen and tyre particles. This study also confirmed that the MPs accumulated on roads were washed away when rainfall exceeded a certain amount.

Keywords: microplastics; road dust; tyre wear; non-point pollution

1. Introduction

Microplastics (MPs) are plastic particles (<5 mm in size) that are either intentionally manufactured or generated from the breakdown of plastic debris [1]. MPs, which act as carriers of chemicals and pollutants, are detected in air, soil, and water, posing a threat to the environment and human health [2–5].

The sources of MPs are divided into point (e.g., effluent from factories and sewage treatment plants) and non-point (e.g., runoff from parking lots, roads, farms, etc.) pollution sources; 37.3% of MPs are generated from point sources, while 62.7% are from non-point sources [6]. Point pollution sources are usually treated emission sources that can be controlled. However, non-point pollution sources flow into aquatic environments from various sources, such as wind, stormwater, road runoff, snow dumping, and mismanaged waste [7,8], and they are difficult to manage.

Road dust is the main non-point pollutant source of MPs, primarily originating from tyres, bitumen, and road-marking paint [9,10]. Tyres consist of 60% styrene butadiene rubber and are combined with various additives and natural rubber [9,11]. Styrene-butadiene-styrene, which is a thermoplastic elastomer, is mainly used in bitumen, while styrene-isoprene-styrene, ethylene-vinyl-acetate, polyamide, and polyacrylate are mainly added to road-marking paint to improve its performance [9,12].

Road dust, which is one of the major MP emission sources that enters the environment, is recognized by many researchers, and various countries such as Iran, India, Japan, Vietnam and Nepal have studied MPs contained in road dust (Table 1) [13–15]. These previous studies reported the contamination status based on particle quantification (particles per weight or area). From the MPs accumulated on road surfaces, 90–95% was expected to



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Copyright: © 2022 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). flow into soil, rivers, and oceans via rainfall runoff [16–19]. Once MPs flow into the aquatic environment, their treatment becomes difficult as 90% of them accumulate in sediments [20]. Therefore, MPs in road dust should be dealt with before it is washed away by rainfall.

Location	Sampling Methods	MPs and MRs Concentration (MPs g^{-1})	Reference
Iran	Sweeping with brush	87.6 ± 78.2	[13]
India	Sweeping with brush	2.27 ± 0.9	[14]
Japan Vietnam Nepal	Vacuum cleaner	2.5 ± 2.7 4.1 ± 3.5 3.9 ± 3.5	[15]

Table 1. A comparison of microplastics concentration in road dust studies.

As MPs contained in road dust mainly flowed into the aquatic environment by rainfall, we hypothesized that MPs on the road would accumulate as the drying period increases. Hence, this study aimed to identify the correlation between the concentration of MPs in road dust and the drying period. In addition, the characteristics of MPs such as the morphology of MPs (size, form and color etc.) and the chemical composition of MPs detected in road dust were investigated.

2. Materials and Methods

2.1. Study Site

The road dust samples were collected from the road side in the downtown area of Goyang city, South Korea, where the traffic volume was measured (Figure 1a,b). The road had eight lanes (four tracks each with lanes in both directions), and the sampling point was a location with high traffic density in rush hours and these geo-located images were captured from Google maps. The annual average daily traffic (AADT) at the sampling point is approximately 24,027 vehicles/d and the average driving speed is 60 km/h. The sampling site is cleaned with a road sweeper once a week.

2.2. Sampling

Road dust was sampled from 17 May to 2 June 2021. As we hypothesised that road dust would be washed away after a rain event, the sample when the road surface had dried after rain was determined as "the 0-d drying period" (Figure S1). The average temperature was 17.4 °C, and 6 h after the rain stopped, we sampled the road dust. The sample "the 1-d drying period" was collected at the same time the next day; sampling was performed at designated locations for drying periods d-0, 1, 2 and 3 (Figure 1c). Samples on different days for each drying period were analysed twice (Table 2). Samples were collected by sweeping one block in front of the kerb using a wooden brush (Figure S2). The sampled road dust was sifted through a 1 mm sieve to remove leaves, stones and branches. It was then stored in a sterilized 50 mL tube and transported to a laboratory. Road dust samples were collected at the same sampling time, and the drying period was recorded.



Figure 1. Locations of: (**a**) Goyang city, (**b**) sampling point, and (**c**) road dust sampling point (drying period (D.P.)). (Courtesy: Google Maps).

Drying Period (d)	Sampling Date	Sampling Time	Amount of Dust Collected (g)
0	17 May 2021	14:35	6.5
0	25 May 2021	14:00	11.1
1	18 May 2021	14:00	12.5
1	26 May 2021	14:00	11.6
2	19 May 2021	14:00	9.4
2	3 June 2021	14:00	21.3
2	20 May 2021	14:00	12.7
3	24 May 2021	14:00	12.2
	Drying Period (d) 0 0 1 2 3 3	Drying Period (d) Sampling Date 0 17 May 2021 25 May 2021 25 May 2021 1 18 May 2021 2 19 May 2021 2 3 June 2021 3 20 May 2021	$\begin{array}{ c c c } \label{eq:Drying Period} \\ \begin{tabular}{ c c c } \hline Drying Period \\ \hline (d) \end{tabular} & Sampling Date & Sampling Time \\ \hline 10 & 17 May 2021 & 14:35 \\ \hline 25 May 2021 & 14:00 \\ \hline 12 & 18 May 2021 & 14:00 \\ \hline 26 May 2021 & 14:00 \\ \hline 26 May 2021 & 14:00 \\ \hline 20 May 2021 & 14:00 \\ \hline 3 June 2021 & 14:00 \\ \hline 3 & 24 May 2021 & 14:00 \\ \hline \end{array}$

Table 2.	Road	dust	sampl	ling o	date	and	dry	ing	period	ι.
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2.3. Sample Pretreatment

The sieved samples were dried in a drying oven at 70 °C for 24 h. Aliquots of 1 g samples were obtained, and each sample was replicated three times. The organic matter in these samples was then oxidized using hydrogen peroxide (H₂O₂, 35%). After the oxidation, each sample was sieved by mesh filter (pore size 100 μ m). The sieved samples were transferred to a beaker with a ZnCl₂ solution (density: 1.6 g/cm³) and allowed to stand for density separation for 2 h, and then the supernatant was decanted into another cleaned beaker. This operation was repeated three times [21]. The density-separated samples were filtered using a track-etched polycarbonate filter (20 μ m pore size and 47 mm diameter) (GVS, Sanford, ME, USA) and dried in a desiccator for 24 h.

2.4. Quantitative and Qualitative Analyses of MPs

For the pretreated samples, the number of MPs (>100 μ m) and their sizes and shapes were examined using a stereoscopic microscope (Discovery V8, Carl Zeiss, Oberkochen, Germany). For the quantitative analysis of MPs, 1 g of road dust sample was aliquoted, and, after pretreatment, the number of MPs contained in 1 g was analysed. The unit was number of detected MPs per gram (MPs/g). Fibres, films, flakes, paint particles, and black particles (i.e., tyres and bitumen) were counted [22]. In addition, the morphology of the detected particles was analysed using scanning electron microscopy (SEM, S-3700N, Hitachi, Ltd., Ibaraki, Japan) at 2.0 kV and energy dispersive X-ray spectroscopy (EDX, Xflash detector 5010, Bruker, Billerica, MA, USA). The particles presumed to be MPs detected through the microscope were analysed using Fourier-transform infrared spectroscopy (FTIR; Spotlight 400, Perkin Elmer, Shelton, CT, USA) in the reflective mode. After the measurement, we used the results that matched the measured spectrum by >70% determined through library searching. Furthermore, to identify tyre components, measurements were performed using Raman spectroscopy (DXR 3xi, Thermo Scientific, Waltham, MA, USA).

3. Results and Discussion

3.1. Concentration of MPs in Road Dust as a Function of the Drying Period

The concentrations of MPs in the continuously collected samples (17–20 May 2021), revealed that the longer the drying period, the higher the MP concentration (Figure 2). In sample 0–1, 552 \pm 39 MPs g⁻¹ were detected, compared to the amount detected in the sample 3-1 drying period, where MPs increased by 2.7 times (1530 \pm 602 MPs g⁻¹). When the drying period was 0 d, the concentration of MPs was low because they were washed away from the roads by rainfall. As the drying period increased, the MP concentration increased due to the accumulation of MPs generated on the roads [10,23].



Figure 2. Microplastic (MP) concentrations in road dust with an increasing drying period.

Despite the same drying period, the concentration of MPs in road dust varied significantly depending on the traffic volume and rainfall. For the samples in the 0 d drying period, the concentrations of MPs contained in road dust were 552 ± 39 MPs g⁻¹ for sample 0–1 and 2106 ± 5 MPs g⁻¹ for sample 0–2, respectively (Table 2). There was an evident difference in the MP concentration between the two samples (sample 0–1 and 0–2) due to the occurrence of heavy rainfall (63.5 mm); thus, a relatively lower MP concentration was detected in the road dust as the accumulated MPs on the road were washed away by rainfall [24]. For sample 0–2, there was no rainfall event for 3 d prior to measurement, and

a rainfall event of only 4.5 mm occurred in the morning (06.00–09.00) on the sampling day. Therefore, it could be assumed that the MPs were not sufficiently washed away from the road by rainfall. The flow of MPs toward the kerb of the road was accelerated by rainfall; therefore, this suggested that the concentration of MPs in the road dust accumulated at the kerb increased [25,26]. This confirmed that a rainfall event exceeding a certain amount is required to wash away the MPs on the road through the drain located at the kerb. The road dust analysis was conducted eight times, which revealed that the average concentration of MPs in road dust was 1363 ± 495 MPs g⁻¹. In a similar study in South Korea, 102–303 particles g⁻¹ of tyre tread were detected in road dust [27]. Compared to this previous study, the daily traffic volume was approximately three times higher in this study. Therefore, traffic volume could be one factor affecting the amount of MPs. As a result of this study, we confirmed that MP concentrations in road dust are highly influenced by drying period, rainfall and traffic volume. Additional factors in MP concentration were temperature and wind [21,28,29].

3.2. Morphology of Microplastics Detected in Road Dust

Considering the shape of the MPs detected in road dust, black particles presumed to be tyre and bitumen contributed the highest proportion ($67 \sim 79\%$), followed by paint ($7 \sim 17\%$), fibre ($6 \sim 15\%$), flake ($5 \sim 7\%$) and film ($1 \sim 3\%$) (Figure 3). Although there are variations in samples, it was difficult to find a correlation between the detected shape ratio and the drying period or MP concentration.

Drying day	Sampling date	Shape Black particle (tire, bitumen) Paint Fiber Flake Film	Amount of MPs (MPs/g)
0	2021/05/17	67% 15% 8% 7%	552 (± 39)
1	2021/05/18	73% 11% 6% 7%	807 (± 62)
2	2021/05/19	76% 9% 5%	1218 (± 309)
3	2021/05/20	79% 7% 5%	1530 (± 602)
0	2021/05/25	68% 11% 14% 5%	2106 (± 5)
1	2021/05/26	71% 11% 12% 5%	1587 (± 224)
2	2021/06/03	<u>68%</u> <u>17%</u> <u>8%</u> <u>6%</u>	1430 (± 50)
3	2021/05/24	68% 11% 15% 5%	1678 (± 251)

Figure 3. Microplastic (MP) concentrations by sampling date.

A study conducted in Australia reported that black fragments represented 75% of all detected MPs [30]. Furthermore, an analysis of stormwater wetland in Sweden showed that black fragments, which were considered to be tyre particles emitted from automobiles, were dominant in the detected MPs [31]. Tyre particles detected in road dust are attributed to the mechanical abrasion of tyres [23], and the SEM analysis results in this study revealed severe surface wear due to friction on the black particles (Figure S3).

The second highest detected shape was that of paint particles, which are mainly generated by physical impacts and weathering on house and building surfaces [32,33]. In addition, when road-marking paint was detected and analysed, 200–300 μ m glass beads were detected in the paint (Figure S4). These glass beads are generally used in road surface markings to provide retro-reflectivity [34] and occur along with the weathered road paint particles. Red, blue, and green colours were mainly observed in the detected flakes and



films and were determined to be substances that originated from road facilities and vehicles (Figure 4).

Figure 4. Microscope analysis of microplastic shapes: fibre, flake, film, paint, black particles (tyre, bitumen).

The sizes of the detected MPs were divided into MPs (fibre, flake, and film) and black particles (tyre and bitumen) (Figure 4). MPs had the highest distribution in the 200–300 μ m range, and 80% were \leq 700 μ m. Moreover, most of the black particles were in the 100–200 μ m range and 80% were \leq 400 μ m (Figure 5).

The size range of tyre tread particles detected in stormwater sediment was >500 μ m, and other studies showed similar distributions, with 85% ranging 50–350 μ m [30,35]. However, Järlskog et al. found that the particle size of road dust sampled at the kerb was mostly in the 10–100 μ m range [22]. In this study, only particles >100 μ m were analysed; therefore, MPs and black particles <100 μ m in size in road dust should be analysed in future studies.



Figure 5. Size range of microplastics: MPs (fibre, flake and film); black particles (tyre and bitumen).

3.3. Chemical Composition of Microplastic Detected in Road Dust

The black particles most detected in the road dust samples were analysed through SEM/EDX, FT-IR and Raman for a qualitative analysis. They were completely black, and the shape was elongated, round and irregular under an optical microscope.

As a result of SEM analysis, a rough surface of black particles was presumed to be caused by breakage or wear on the road. The elements of the black particles were confirmed by EDX [36] and contained S, Na/Zn, Ca, K, Mg, Fe and Al, which are the main components contained in tyres (Figure 6 and Figure S8). The main components of tyre are rubber, which comprises 40–45%, and filler materials (carbon black(C), silica (Si) and chalk (CaCO₃)) 30–35%. Vulcanization agents (sulphur (S), zinc oxide (ZnO)) account for 2–5% (Table S1). Therefore, Si is most strongly detected, but all elements must be detected [11]. In this study, a decrease in the detection value was observed when there were stones or other substances around the black particles during EDX analysis. Therefore, accurate density separation is important to obtain accurate results in EDX mapping analysis. As a result of FT-IR analysis, black particles were confirmed that exhibited the highest proportion in road dust were poly(butadiene: styrene) components (Figure 7a). Moreover, carbon black was detected in the Raman analysis (Figure 7b). Carbon black is a component of tyres, and poly(butadiene: styrene) is mainly used to enhance rubber for tyres and bitumen; thus, the black particles were categorised as bitumen or tyre particles [23,37,38].

When coloured MPs were analysed, we detected substances in them that are used in coating particles (ethyl cellulose and COVINAX), and rubber for anti-vibration and impact (poly(norbornene)), indicating that they were generated from vehicle coatings, brake pads and bumpers [23] (Figures S5–S7). Furthermore, polyethylene was mainly detected in the shape of fibres. Fibres are generated from various sources (such as synthetic clothing, ropes, etc.) and introduced into roads through the atmosphere [39,40]. In addition, among the components detected in road dust, some substances are commonly used as outer wall materials (i.e., materials used in moulding and sphere-shaped polystyrene used as building interior materials). This indicated that MPs in road dust were not only from the accumulation of MPs generated by vehicles but also from the atmospheric deposition of MPs generated in cities [41].

As 50% of the MPs generated on roads flow into aquatic environments without treatment, it is necessary to reduce road dust through periodic road cleaning and additional non-point source treatments are required for mitigation [42]. In particular, treatment measures for road dust accumulated at the kerb are also required because the generated road dust moves to the kerb [28]. According to the study by Järlskog et al., road dust contains a large number of black particles <100 μ m [22]. These smaller MPs have a higher potential to affect human health, including detrimental respiratory and cardiovascular impacts [17,34,43]. Although respiratory dysfunction due to MP inhalation has not yet been verified in any report, residents possibly inhale MPs every day [19]. From a public health perspective, the quantification of MPs contained in road dust can assist in estimating their impact on human health [44].



Figure 6. Optical microscope and SEM/EDX mapping of particles of road dust sample (red circle: black particle, yellow arrow: soil and stone).



Figure 7. Analyses of black particles by (**a**) Fourier-transform infrared spectroscopy (FTIR) and (**b**) Raman analysis.

4. Conclusions and Implications

The following conclusions have been obtained from this study:

- (1) In the samples with 0 to 3 d drying periods, the number of MPs contained in road dust increased with the increase in the drying period because MPs accumulated on the road. In this study, a total of eight analyses were performed, two times for each drying period (d 0, 1, 2 and 3), and the average concentration of road dust was 1363 (± 495) MPs g⁻¹.
- (2) In this study, 72% of the MPs in road dust were black particles (i.e., tyre and bitumen particles). When the components of the black particles were analysed, carbon black used in tyres and poly(butadiene: styrene) used in bitumen and tyre were detected, the black particles were presumed to be from tyres and bitumen.
- (3) The size detection limit of this study was 100 μm, and most (80%) of the black particles detected had sizes <400 μm. As smaller black particles (MPs < 100 μm) in road dust may adversely affect human health, even though respiratory dysfunction owing to MP inhalation has not been verified, their quantification may contribute to assessing their impacts on public health.</p>

A limitation of this study was that tyres hardly reflect light by material and colour, so it is difficult to measure components by spectroscopic analysis (FT-IR and Raman spectroscopy). Therefore, it is difficult to define whether all of the black particles observed with a microscope are tyre particles or bitumen. Our results have shown that MPs in road dust accumulate on roads as the drying period increases. Future studies could monitor stormwater runoff to estimate the microplastic flow into water environments and quantify particles <100 μ m in size.

Supplementary Materials: The following are available online at https://www.mdpi.com/article/ 10.3390/su14053006/s1, Figure S1: Sampling date and daily precipitation; Figure S2: Sampling (a) site, (b) point, (c) road dust sample, (d) sampled road dust; Figure S3: SEM image of black particles (x 100); Figure S4: SEM image of glass beads (a) x 30, (b) 100, and (c) SEM-EDX analysis; Figure S5: Coating material FT-IR analysis (Ethyl cellulose (matching score: 0.708), COVINAX (matching score: 0.822)); Figure S6: Rubber for anti-impact and vibration FT-IR analysis (Poly(norbornene) (matching score: 0.835)); Figure S7: Polyethylene FT-IR analysis (Polyethylene (low density) (matching score: 0.760)); Figure S8: SEM-EDX analysis of tyre from road dust; Table S1: Tyre chemical component.

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