



Article Investigating Water Storage Dynamics in the Litter Layer: The Impact of Mixing and Decay of Pine Needles and Oak Leaves

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Abstract: Little is known about how the degree of mixing various forest-forming species affects forest floor hydrology. We evaluated the water storage capacity of the resulting litter layer by mixing the litterfall of Scots pine and sessile oak and studying their decomposition time. We prepared 90 artificial samples containing pure pine litter, pure oak litter, and mixed pine–oak litter with varying shares of pine needles. These samples were subjected to 15 months of decomposition in soil. After every three months of decay, some samples were removed from the soil, and their water storage capacity, bulk density, and C:N ratio were evaluated. Our findings indicate that samples with the greatest water storage capacity had a low C:N ratio and a predominant share of oak leaves. Conversely, samples with a high C:N ratio and a predominant share of pine needles had the lowest water storage capacity. After 12 and 15 months of decomposition, the water storage capacity increased by more than 52% compared to the initial water capacity of the samples. The highest increase in water storage capacity (>40%) was observed in samples with 80 and 100% of pine needles. Our findings suggest that introducing mixed-species stands, with deciduous species as the predominant component, can yield several ecological benefits, such as an increased ability to store water in forest floor.

Keywords: forest hydrology; forest floor; pine-oak mixed forest; water properties; C:N ratio

1. Introduction

In the last 40–50 years, drought conditions have become more prevalent globally due to climate change [1], and further increment of drought frequency is predicted [2–4]. In areas where climate change will cause more droughts in the future, the composition of forests should be gradually changed to species exhibiting greater tolerance to drought stress [5]. Some studies have indicated that repeated droughts in the same area could cause changes in the species composition of forests towards reduced diversity and the emergence of single-species tree stands [6,7]. On the other hand, during the last decade, forest management has supported mixing tree species with various resistance and ecological traits to mitigate adverse effects on stand growth and reduce drought stress [8]. For example, in drought-prone sites in the Mediterranean region, the pine–beech mixed forest was indicated to be an effective stand model in view of climate change [9]. According to Kotlarz et al. [10], *Quercus robur* monocultures are more sensitive to drought than mixed stands with *Q. robur* as the



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Copyright: © 2024 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). dominant species, while the presence of a ~20% admixture of *Q. robur* and *Pinus sylvestris* significantly reduced deterioration of vegetation indices during short-term drought.

Many studies have been conducted on the resistance of mixed stands to water stress during drought, focusing mainly on the hydraulic properties of trees as a reaction to drought events [11–14]. However, trees possess numerous belowground adaptations (e.g., root architecture and depth) that likely interact with soil properties and soil biota to deal with drought effects [15–17]. Thus, taking into account only aboveground traits of trees without consideration of their belowground traits may lead to incorrect projections concerning consequences of drought.

Little is known about the influence of mixing various tree species on soil hydrology, and consequently the indirect impact of trees inside mixed stands on drought prevention. Stand species composition may affect soil physical and chemical properties through litterfall constituting the litter layer on the soil surface, which comprises the boundary between the atmosphere and mineral soil in forest ecosystems [18]. The litter layer is the first forest soil horizon that intercepts throughfall, and its properties determine the amount of water that evaporates, is stored, or percolates into the mineral soil after each rainfall. The hydrological importance of the litter layer results from its control over the transfer of water and energy between the sub-canopy atmosphere and the mineral soil [19]. The litter layer exhibits low thermal conductivity and acts as an insulator between the atmosphere and mineral soil, reducing the temperature, temperature fluctuations, and the thermal gradient in mineral soil [20].

Some studies have demonstrated that litter and detritus layers can retain significant amounts of throughfall, thus affecting water flow into mineral soil and water supply to vegetation [21–23]. Greiffenhagen et al. [24] stated that these layers contain about 20% of the total water quantity available to plants up to a depth of 1 m with high species-dependent variability. The hydrological properties of the litter layer are influenced by the type of vegetation and the stand species composition. For example, Gerrits [25] showed that cedar litter has only about one half of the water storage capacity (1.0 mm) of beech litter (1.8 mm). Ilek et al. [26] stated that spruce litter (*Picea abies*) retains nearly three times more water than beech litter (Fagus sylvatica) and almost twice as much as fir litter (Abies alba). For pine litter (Pinus radiata) and eucalyptus litter (Eucalyptus spp.), Putuhena and Cordery [23] recorded the water storage capacity of 2.3 and 1.4 mm, respectively. However, little is known about the effect that the degree of mixing of forest-forming species has on the litter hydrology and the ways in which water properties of the litter layer change with time, i.e., during decomposition processes taking place in the surface soil horizon. Thus, our study aimed to evaluate (1) the effect of mixing litterfall of Scots pine and sessile oak on the water storage capacity of the resulting litter layer, and (2) how the mixing degree of both species changes the water storage capacity and bulk density while progressing the decomposition process.

So far, no research has been conducted on the relationship between the effect of mixing the litterfall of various trees and water storage capacity. We analyzed how differences in the chemical composition of needles and leaves affect water retention capacity to better understand water retention dynamics in the context of mixed coniferous–deciduous forests. Knowledge concerning the influence of stand species composition on litter hydrology is important to ensure greater understanding of the water cycle in forest ecosystems, and thus it can be used in forest management practice under climate change conditions.

2. Materials and Methods

2.1. Study Site and Sampling

A field experiment was set up in the Experimental Forest of the Poznań University of Life Sciences in Murowana Goślina (Poland) in the Potasze forest district (division 73a) situated in a temperate climate zone (Figure 1A). The average annual temperature at the site is 8.5 °C and the average annual precipitation is 500 mm. Dominant canopy trees include sessile oak (*Quercus petraea* [Matt.] Liebl.) and, in some places, Scots pine (*Pinus sylvestris* L.) [www.bdl.lasy.gov.pl/portal/mapy-en (accessed on 20 December 2023)]. Soil

in the study area is classified as Albic Luvisols (Neocambic) [27]. Research material in the form of pine needles and oak leaves was collected in uniform pine or oak fragments of the stand. Due to this, it was possible to collect needles or leaves from the litter layer, free from admixtures of other tree species. Needles and leaves were collected using a brush at the end of November 2020, promptly after the fall of fresh organic matter.



Figure 1. (**A**) Location of pine needle and oak leaf sample collection and the field experiment on the decomposition of artificial litter samples. (**B**) Examples of artificial pine and oak–pine litter samples prepared inside square plastic containers. (**C**) Top view of the lid of a plastic container with artificial litter dug into the soil. (**D**) View of all artificial litter samples tested for the influence of decomposition time on water storage capacity of the forest floor.

2.2. Experiment Design

The collected pine needles and oak leaves were air-dried in laboratory conditions. Then, 90 artificial samples were prepared, containing litter consisting of pine needles (100% pine and 0% oak), oak leaves (100% oak and 0% pine), or needles and leaves in various degrees of mixing, i.e., with a mass fraction of pine needles of 20, 40, 60 and 80% (Figure 1B). We prepared 18 samples for each variant involving pine needles and/or oak leaves in the litter. The samples were prepared in square plastic containers, designed so that rainwater could enter the container through the top and flow out through its bottom during the field experiment. Fine plastic mesh was placed on the holes cut in the bottom of the containers and in the covers, which on the one hand allowed for water to flow freely through the samples and also prevented various types of contaminants from entering the containers, including litterfall reaching the forest floor (Figure 1C). Each sample contained 20 g of organic matter. When weighing 20 g of needles and leaves, their initial moisture content was considered, which was approximately 10% (moisture content was determined after drying a portion of the organic material at 105 °C). The weighed material was first placed in beakers with water and left until it fell to the bottom, i.e., until it reached a density of

>1 g cm⁻³ (~7 days). This time was considered to be the moment of filling the internal capillarity of needles and leaves [28]. Once this level was reached, the needles and leaves were placed in square containers and immersed in water again. Then, we mixed the material immersed in water to ensure random distribution of needles and leaves within each container (especially in the mixed oak–pine samples). Afterwards, the container was quickly pulled out of the water with the organic material at the bottom. In the next stage, containers with prepared litter samples were subjected to a gravitational drainage process lasting approximately 1 h. After completing this process, the containers with the samples were weighed to determine the wet mass of the samples. Next, we calculated the volume of each sample by measuring the height of the wet sample formed in the container and multiplying the height by the surface area of the container's base. Based on these measurements, we calculated the initial water storage capacity (mm) and bulk density (g cm⁻³) of the artificial litter samples (with undecomposed organic matter) according to the following formulas [26]:

Water storage capacity =
$$[(Wet mass - Dry mass)/Volume] \times 10,$$
 (1)

$$Bulk \ density = Dry \ mass / Volume, \tag{2}$$

where 10 is a factor of conversion into mm of H_2O .

The prepared samples were transported to the forest, where they were placed in the soil in such a way that the lids of the containers were at the level of the forest floor (Figure 1D). After every three months of decomposition, 18 samples were removed from the soil (3 samples per each variant) and transported to the laboratory. These samples were immersed in water for seven days. After removing the samples from the water and completing the gravity drainage process, we calculated the water storage capacity and bulk density of the samples at a given degree of decomposition analogously to Formulas 1 and 2. We also estimated the percentage changes in water storage capacity (Δ WSC), bulk density (Δ BD), mass and volume losses of samples every three months of decay (the percentage changes in all variables were determined in relation to their initial value, i.e., determined for undecomposed samples). We assumed a maximum decomposition time of 15 months (5 sampling dates).

2.3. Chemical Analyses

We determined the carbon and nitrogen contents, the C:N ratio, and the percentages of lignin, cellulose, extractives, and ash for pure oak and pine samples. These tests were carried out on fresh samples and at each of the five stages of decomposition. Carbon and nitrogen contents, as well as the C:N ratio, were additionally determined for samples with different shares of pine needles (20-80%) at each stage of decay. Before the determination of chemical components of leaves and needles, the samples were ground in a Fritsch Pulverisette 15 laboratory mill (Fritsch GmbH, Idar-Oberstein, Germany). Carbon and nitrogen contents were determined using the Leco analyzer (Leco, St. Joseph, MI, USA). Based on carbon and nitrogen contents, the C:N ratio was calculated. Cellulose content was determined according to Seifert's method [29] using a mixture of acetyl acetone, 1.4-dioxane and hydrochloric acid to isolate cellulose. Acid-insoluble lignin content was assessed according to the T 222 om-06 standard TAPPI method [30] using 72% sulfuric acid to hydrolyze and solubilize carbohydrates. Extractives soluble in alcohol (96% ethanol) were determined according to the T 204 cm-97 standard [31], whereas ash was tested according to T 211 om-02 [32]. All the chemical analyses were repeated with three replicates for each sample of leaves and needles.

2.4. Statistical Analyses

Statistical analysis and associated graphics were performed in Statistica 13.3 PL (Stat-Soft Inc., Tulsa, OK, USA) and the R programming language (R Core Team, Vienna, Austria, 2020) in R Studio (RStudio Team, Boston, MA, USA, 2020). Significant differences were tested by one-way ANOVA and post hoc Tukey's test after checking the normality of distribution with the Shapiro–Wilk test and the equality of variance with Levene's test. In the case of the non-parametric nature of the data, the Kruskal–Wallis test was applied. We adopted a general linear model (GLM) to investigate the effect of the initial share of pine needles in the litter layer and the time of the decomposition process on the C:N ratio, percentage changes in the bulk density and water storage capacity of the samples. Principal component analysis (PCA) was employed to examine the dependence between the share of pine needles in litter samples and their properties. All the tests were performed at a significance level of 0.05.

3. Results

The initial water storage capacity (*WSC*₀) of litter samples with undecomposed organic matter ranged from 0.41 to 1.12 mm (0.63 \pm 0.02 mm on average) (Figure 2A). The *WSC*₀ of pure pine litter was, on average, 64% greater than that of pure oak litter. We observed that the 20 and 40% admixtures of pine needles in the litter did not significantly increase its water capacity compared to those of pure oak litter. In turn, even a 20% admixture of oak leaves in the litter caused a reduction in its water storage capacity on average by ~22%. With a 40% share of oak leaves, *WSC*₀ of the litter was ~37% lower than that of pure pine litter (Figure 2A). The initial bulk density (*BD*₀) of artificial litter samples increased with the share of pine needles (Figure 2B). The *BD*₀ of pure pine litter was 165% greater than that of pure oak litter samples. We observed a strong linear relationship between *WSC*₀ and *BD*₀ of litter samples (Figure 2C).



Figure 2. (A) The initial water storage capacity (WSC_0), (B) the bulk density (BD_0) of artificial litter samples prepared from fresh fallen pine needles and oak leaves, and (C) the relationship between these two variables. Different letters indicate significant differences between samples with varying shares of pine needles (Tukey's test, p < 0.05).

Freshly fallen, undecomposed oak leaves and pine needles differed in their chemical properties. Pine needles contained, on average, 7.5% more carbon, 50% more nitrogen, 70% more cellulose, and 133% extractives than undecomposed oak leaves, which, compared to the needles, had a higher C:N ratio (by 45% on average) and contained approximately 48% and 164% more lignin and ash, respectively (Table 1). The chemical properties of pine needles and oak leaves changed depending on their decomposition time. While the carbon content remained relatively constant in both species during the 15-month decomposition process, the nitrogen, lignin, and ash contents increased, on average, by ~92, 115, and 190% in pine and by ~63, 43, and 79% in oak leaves (Table 1). The cellulose content did not change significantly in oak leaves during the decomposition process. In the case of pine, after 15 months of decomposition, the cellulose content decreased by ~61%. In both

species, we also observed a decrease in the C:N ratio and extractive contents (greater in pine than in oak) with the time of the decomposition process. Interestingly, despite the initial differences in the contents of cellulose and lignin in pine needles and oak leaves, after approximately nine months of decomposition, the levels of these compounds in both species began to become similar.

Table 1. Chemical properties of pure pine and oak leaf samples progressing over the decomposition time (mean values). Different letters indicate significant differences between decomposition times (Kruskal–Wallis test, p < 0.05).

Variable	C (%)		N (%)		C:N		Cellulose (%)		Lignin (%)		Extractives (%)		Ash (%)	
Decomposition Time (Months)	Pine	Oak	Pine	Oak	Pine	Oak	Pine	Oak	Pine	Oak	Pine	Oak	Pine	Oak
0	47.1 ^a	43.9 ^{ab}	1.2 ^a	0.8 ^a	39.9 ^a	57.7 ^a	44.4 ^a	26.1 ^a	24.2 ^a	35.8 ^a	30.3 ^a	13.0 ^a	2.2 ^a	5.8 ^a
3	50.0 ^a	43.7 ^{ab}	1.5 ^{ab}	0.9 ^{ab}	33.1 ^{ab}	47.0 ^{ab}	44.1 ^a	24.8 a	28.4 ^{ab}	38.8 ^{ab}	17.1 ^{ab}	10.8 ^a	2.5 ^{ab}	8.1 ^{ab}
6	50.8 ^a	44.8 ^b	1.8 ^{ab}	1.2 ^{ab}	28.2 ^{ab}	38.9 ^{ab}	39.9 ^a	27.0 ^a	36.6 ^{ab}	42.3 ^{ab}	19.7 ^{ab}	9.7 ^a	3.5 ^{ab}	8.4 ^{ab}
9	49.7 ^a	43.5 ab	2.6 ^b	1.3 ^{ab}	19.0 ^b	32.7 ^{ab}	29.7 ^a	27.0 ^a	47.9 ^{ab}	45.4 ^{ab}	13.6 ^b	9.0 ^a	4.4 ^b	10.5 ^{ab}
12	49.0 ^a	42.0 ab	2.5 ^{ab}	1.4 ^b	19.3 ^{ab}	30.0 ^b	27.4 ^a	27.4 ^a	50.7 ^{ab}	47.4 ^{ab}	14.8 ^{ab}	8.6 ^a	4.2 ^{ab}	11.8 ^b
15	47.8 ^a	38.0 ^a	2.5 ^{ab}	1.3 ^a	19.0 ^b	28.4 ^b	27.6 ^a	26.4 ^a	51.9 ^b	51.1 ^b	15.1 ^{ab}	8.8 ^a	4.2 ^{ab}	10.4 ^{ab}

The C:N ratio of samples decreased clearly during the first months of decomposition and stabilized after approximately 9 months of decay, reaching an average of ~24 (Figure 3A). We observed that C:N decreased when the initial proportion of pine needles in the samples increased, i.e., the C:N of pure pine samples was approximately 32% lower than that of pure oak samples (Figure 3B). The influence of decomposition time and the share of pine needles in the samples on the C:N ratio was confirmed by GLM analysis (Table 2).



Figure 3. The C:N ratio of litter samples (**A**) progressing with the time of organic matter decomposition (regardless of the share of pine needles) and (**B**) in relation to the initial share of pine needles (regardless of decomposition time). Different letters indicate significant differences in C:N ratio between decomposition times (Tukey's test, p < 0.05) or shares of pine needles (Kruskal–Wallis test, p < 0.05).

Litter mass loss progressed with the decomposition process in all variants of the initial content of pine needles in the samples (Table 3). After 15 months of decomposition, the mass loss amounted to an average of 63%, with the largest loss (~72%) found in samples with 40% pine needles and the lowest in samples with 80 and 100% oak leaves. After the first 3 months of decomposition, the lowest mass losses were recorded in samples with a predominant share of oak, while the greatest (over 20%) were found in samples containing 80 and 100% of pine needles.

Table 2. General linear model analysis (GLM) for the C:N ratio, percentage changes in the bulk density (Δ BD) and water storage capacity (Δ WSC) of the litter layer in terms of initial bulk density and water storage capacity of artificial litter samples in terms of the share of pine needles and decomposition time. Significance effects (p < 0.05) are shown in bold.

Variable	C	N	ΔΙ	BD	ΔWSC		
Vulluoic	F	p	F	p	F	p	
Share of pine needles	122.40	0.0000	41.58	0.0000	9.67	0.0000	
Decomposition time	438.32	0.0000	10.34	0.0000	38.08	0.0000	
Share x Decomposition time	3.02	0.0005	3.39	0.0001	1.79	0.0428	

Table 3. Mean percentage losses of litter mass and volume (grey color) of samples in terms of decomposition time and initial share of pine needles. Values with the same letter are not significantly different between decomposition times or shares of pine needles (Kruskal–Wallis test, p < 0.05).

Share of Pine Needles (%)	Decomposition Time (Months)											
	3		6		9		12		15		Mean	
0	8.3	4.4	33.3	26.9	50.8	55.6	56.7	72.7	55.7	70.3	40.2 ^a	46.0 ^a
20	2.5	7.3	36.7	23.9	58.3	64.7	57.5	66.7	56.7	79.8	42.3 ^a	45.6 ^a
40	7.5	4.8	39.2	25.7	62.5	67.1	60.0	68.8	71.7	74.0	48.2 ^a	48.1 ^a
60	11.7	4.5	38.3	22.2	56.7	51.3	67.5	69.1	68.3	65.0	48.5 ^a	40.6 ^a
80	20.0	8.1	37.5	16.2	57.5	35.7	68.3	59.3	65.0	56.3	49.7 ^a	35.1 ^a
100	24. 2	4.0	41.7	14.5	59.2	26.7	61.7	38.7	65.8	44.5	50.5 ^a	25.7 ^a
Mean	12.4 ^a	2.9 ^a	37.7 ^a	20.2 ^a	57.5 ^b	50.2 ^b	61.9 ^b	64.7 ^b	63.2 ^b	62.8 ^b	46.6	40.2

Besides litter mass loss, we also observed changes (on average ~40%) in the volume of samples over the decomposition process (Table 3). After the first 3 months of decomposition, the recorded volume losses in samples varying in the share of pine needles were below 10%. After 15 months of decay, volume loss amounted to an average of 63%, with the greatest loss (>70%) found in samples with a predominant share of oak leaves (60–100%) and the lowest (~45%) in pure pine samples.

As a result of the loss of mass and volume of samples along with the advancement of the decomposition process, there were also changes in bulk density (Δ BD). Interestingly, throughout the entire period of the study, the bulk density of the tested samples fluctuated. We observed that the bulk density of samples (regardless of the pine and oak mixing degree) in the first nine months of decay decreased, on average, by ~11%, while after 12 and 15 months of decay the bulk density increased by about 12% (Figure 4A). In samples with a predominant pine share, density mainly decreased. In samples with a predominant proportion of oak leaves, there was primarily an increase in density in relation to the initial bulk density of undecomposed litter (Figure 4B). GLM analysis showed that changes in the bulk density of the samples were influenced by both decomposition time and the initial share of pine needles in the samples (Table 2).

In general, samples with a low C:N ratio and a predominant share of oak leaves exhibited the greatest water storage capacity (WSC). The lowest WSC was found in samples with a high C:N ratio and a predominant share of pine needles (Figure 5). We observed changes in water storage capacity (Δ WSC) in relation to the initial water storage capacity (WSC₀) with decomposition time. Regardless of pine needle share, the Δ WSC increased, on average, by ~23% in the first 6 months of decay (Figure 6A), while after 12 and 15 months the water storage capacity was over 52% greater than WSC₀. The highest increase in water storage capacity (>40%) occurred in samples with a predominant share of oak leaves, while the lowest (about 28% on average) was recorded in samples containing 80 and 100% pine needles (Figure 6B). GLM analysis confirmed the influence of the initial share of pine needles and decomposition time on Δ WSC. We found a clear linear relationship between

 Δ WSC and Δ BD (Figure 7). In most samples with a predominant share of pine needles, their density mainly decreased, even by about 50%. In turn, the bulk density of samples with a predominant proportion of oak leaves was primarily increased (even by about 80%), and a decrease in density did not exceed 40%. Throughout the entire range of changes in the bulk density of the tested samples, we noted an increase in water capacity compared to the WSC₀. The greatest increases in WSC (reaching 80%) were found in samples with a predominant amount of oak leaves. Δ WSC of samples with a predominant pine share did not exceed 60%.



Figure 4. Percentage changes in bulk density (Δ BD) of artificial litter samples progressing with the time of organic matter decomposition (**A**) and in relation to the initial share of pine needles (**B**). Different letters indicate significant differences in Δ BD between decomposition times or shares of pine needles (Kruskal–Wallis test, *p* < 0.05).



Figure 5. Dependence of water storage capacity (WSC) of the litter layer on the degree of organic matter decomposition (C:N) and the initial share of pine needles in the samples.



Figure 6. Percentage changes in water storage capacity (Δ WSC) of artificial litter samples progressing with decomposition time (**A**) and in relation to the initial share of pine needles in samples (**B**). Different letters indicate significant differences in Δ WSC between decomposition times or shares of pine needles (Tukey's test, *p* < 0.05).



Figure 7. Dependence of percentage changes in water storage capacity (Δ WSC) of samples on changes in their bulk density (Δ BD) during the 15-month decomposition process in relation to the initial share of pine needles in the samples. The red line indicates no change in bulk density.

A projection of the variables on the factor plane clearly demonstrated the relationship of the share of species in the litter with their properties (Figure 8). PCA analysis explained 84% of the variability of the examined features. Factor 1 is related to water storage capacity, while Factor 2 is related to the quality of organic matter expressed by the C:N ratio. PCA analysis confirmed distinctiveness of the litter with a predominance of pine needles (80 and 100% of pine needles). These two groups were characterized by the highest initial water storage capacity, while the samples with the lowest proportion of needles were characterized by the greatest changes in water storage capacity (Δ WSC) and changes in their bulk density (Δ BD).



Figure 8. Projection of variables on the plane of the first and second PCA factors, where ML is mass loss of litter samples during decay, VL is volume loss of litter samples during decay, C is carbon content, N is nitrogen content, WSC₀ is initial water storage capacity of undecomposed litter, BD₀ is initial bulk density of undecomposed litter samples, WSC is the water storage capacity of semi-decomposed samples, Δ WSC is the percentage change in water storage capacity in relation to WSC₀, Δ BD is the percentage change in bulk density regarding BD₀.

4. Discussion

Our 15-month experiment confirmed that the degree of mixing of pine needles and oak leaves and their decomposition rate are very closely related to water storage capacity. Particular sublayers of the forest floor usually vary in their water storage capacity due to differences in bulk density, porosity, and advancement of organic matter decay, i.e., the semi-decomposed detritus layer usually has greater density and water storage capacity than the undecomposed litter layer [26,33]. It corresponds with our results. We observed that at each stage of the decomposition process, the water storage capacity of samples increased compared to undecomposed litter samples, regardless of pine needle share (Figures 6 and 7). The greatest increase in water storage capacity was found in samples with a predominant share of oak leaves (Figures 6B and 7). It could be related to the fact that during decomposition, the leaves begin to stick more closely to one another. The relatively large surface area of oak leaves may cause more water to accumulate between adjacent leaves, making it difficult to drain, which probably increased the water storage capacity of samples with a high proportion of oak leaves. The differences in the shape of individual litter components may affect the distribution of water flow channels in the forest floor [34]. Thus, the presence of pine needles between the leaves may prevent the leaves from sticking to one another (especially when wet), which reduces the water storage capacity of the litter layer, while on the other hand it may result in increased water infiltration into the soil profile.

Some studies indicate differences in the water storage capacity of the litter between single-species stands, which results from the differences in the morphology of dead organic debris contained in the litter [26]. Leaf litter typically retains more water than coniferous litter [34,35]. According to [36], mixed litter exhibits higher water storage capacity than pure coniferous litter; however, mixed coniferous–broadleaved litter usually shows a lower water storage capacity than mixed coniferous litter [33]. In turn, in ref. [37], it was stated that an admixture of pine needles in oak litter has a slight effect on bulk density and total porosity of litter (compared to pure oak litter) while increasing macroporosity and decreasing the water storage capacity and degree of litter saturation. The differences in bulk density, total porosity and macroporosity of pure and mixed litter layers reflect the initial material's diversity prior to the decomposition process. It appears that the capacity of the litter layer for retention of water might be affected by the origin of organic matter

from different tree species. Therefore, it is essential to carry out additional research at a local level to gain greater insight into the effect of mixing tree species on the ability of the litter layer to hold water. This knowledge will help devise effective forest management strategies under climate change conditions.

Our results showed that the water storage capacity of mixed litter does not always exceed that of the pure litter layer. In the case of undecomposed litter, the greatest water storage capacity was recorded for pure pine litter. Admixture of oak leaves caused reduction in water storage capacity (Figure 2A). On the other hand, pure oak undecomposed litter had the lowest water storage capacity, and the admixture of pine needles caused no significant increase in its water storage capacity. In the case of semi-decomposed samples, the lowest water storage capacity was found in pure pine samples, while an admixture of oak leaves only slightly increased water storage capacity (Figure 6B). Pure oak leaf semi-decomposed samples had the highest water storage capacity, and the admixture of pine did not significantly change its water storage capacity. Thus, it may be concluded that the water storage capacity of the litter layer in mixed pine–oak stands depends more on the degree of organic matter decomposition than on the proportion of mixing in the litterfall of the two tree species.

The properties of detritus in the forest floor depend on the rate of decomposition processes, which in turn is important in water retention [38]. Its degree of decay is heterogeneous both vertically (gradually increasing closer to the mineral soil horizon) and horizontally, with decomposition being dependent on such drivers as climate and canopy type [39]. The dominant factor determining decomposition processes is the quality of the supplied detritus, especially litter [40–42]. Organic matter decomposition processes include leaching, microbial colonization, and fragmentation [43]. During detritus decomposition, microbial nutrient processes, i.e., immobilization and mineralization, play an important role in changes in detritus nutrient contents and stream nutrient concentrations [44,45]. Our research confirmed a strong relationship between water storage capacity and the C:N ratio expressing the quality of organic matter. A high ratio corresponds to a higher organic C content and a lower N content, suggesting greater stability of organic C in the soil, while a low ratio corresponds to a lower organic C content and a higher N content, indicating a greater degree of decomposition of organic matter in the soil [46].

The bulk density of the forest floor containing decomposing organic matter usually increases with depth [26]. However, in our study, we observed both an increase and a decrease in bulk density over 15 months of decay (Figures 4 and 7). The low density of oak undecomposed litter (Figure 2B) could have been caused by the fact that leaf litter has a large surface area and often "curls", creating air spaces within the litter layer. This produces lesser mass per volume compared to that of a needled pine litter layer [47]. The increase in the density of semi-decomposed samples with a predominance of oak leaves was probably related to a lower loss of mass and a greater loss of volume in these samples during decomposition (Table 3). In turn, the decreasing density of semi-decomposed samples with a predominant share of pine needles was probably related to a greater decrease in mass than in volume in the initial phase of decomposition.

In general, the water storage capacity of the forest floor is related to bulk density, i.e., the water storage capacity grows with an increase in density [48]. In our investigations, despite decreases in density in the initial stage of decomposition, the water storage capacity of semi-decomposed litter increased. This suggests that the hydrological properties of the litter layer are closely related not only to its physical properties, but also to the change in its chemical properties during decomposition. Thus, the water storage capacity of the litter layer should also be considered at the molecular level of the leaf and needle structure. This is because needles and leaves are chemically complex molecular structures with specific hygroscopic properties that affect their water-holding capacity [49–51].

When examining the influence of the chemical composition of leaves and needles on their water storage capacity, it should be considered that the chemical structure of the outer and inner layers in leaf and needle tissue differs considerably. Due to the abovementioned chemical differences in individual layers of leaves and needles, they probably differ in terms of hygroscopic properties. As a result, degradation of individual layers of leaf and needle tissue (first external, then internal) affects the ability of the littler layer to store water during layer decay [52]. The leaves and needles are covered with an outer layer of extractive substances, e.g., waxes [53–56], which consists of hydrophobic organic compounds, mainly straight-chain aliphatic hydrocarbons [57,58]. This composition of waxes makes them responsible for maintaining wettability of the leaf surface [59].

In our study, significant reduction in the percentage of extractive substances both for leaves and needles was noted after the first 3 months of their decay. A significant part of detected extractives are waxes, found in the outer layer of needle and leaf tissue [55]. According to the literature concerning wettability of leaves and needles, after removing wax from the surface of aging leaves and needles, the contact angle decreases significantly and the leaves and needles become easily wettable [60]. The presence of waxes on the leaf surface tends to reduce their wettability [61]. However, in the case of the tests described, this significant loss of extractive substances (waxes) was not accompanied by an increase in water storage capacity. This may suggest that the extractive substances do not significantly affect the tested properties. However, it is worth focusing on the chemical components of leaves and needles that appear under the layer of waxes, lignin and cellulose. These components, due to differences in their structure, vary in their hydrophobicity, e.g., cellulose is more and lignin less hydrophilic [62].

The sorption of moisture by each cell wall polymer depends on not only its hydrophilic nature but also accessibility of water to the polymer's hydroxyl groups. Most, if not all, of the hydroxyl sites in lignin are accessible to moisture. The non-crystalline portion of cellulose (approximately 40%) and the surfaces of the crystallites are accessible to moisture, whereas the crystalline part (approximately 60%) is not [63]. Therefore, considering the water storage capacity of the litter layer comprising needles and leaves, attention should be paid to the content of these components. It should be noted that the content of cellulose in needles decreased gradually, with the greatest reduction observed between 6 and 9 months of decay (Table 1). Only small changes in the percentage of leaf cellulose were noted during the decomposition process. In the case of lignin, a different phenomenon was observed, as the lignin content gradually increased in both needles and leaves. However, in the case of needles, this can indicate the time when the greatest growth occurred (from the third to the sixth month). In the case of leaves, the decline was rather constant. Finally, after 15 months of decomposition, the contents of cellulose and lignin in the needles and leaves were comparable. This may indicate their similar hygroscopic properties. Differences in the hygroscopic properties of needles and leaves after decomposition can only be caused by the difference in the shares of extractive substances.

5. Conclusions

This study found that the hydrological properties of the litter layer depend on the proportion of pine needles and oak leaves, their decomposition time, and their chemical properties. Results revealed that the water storage capacity of the litter layer increased with progress in its decomposition and with the proportion of oak leaves. The study suggests that promoting mixed-species stands, particularly with a predominance of deciduous species, can contribute to enhanced water storage capacity and improved ecological benefits.

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References

- 1. Wilhite, D. Managing Drought Risk in a Changing Climate. *Clim. Res.* 2016, 70, 99–102. [CrossRef]
- O'Gorman, P.A.; Schneider, T. The Physical Basis for Increases in Precipitation Extremes in Simulations of 21st-Century Climate Change. Proc. Natl. Acad. Sci. USA 2009, 106, 14773–14777. [CrossRef]
- 3. Dai, A. Drought under Global Warming: A Review. WIREs Clim. Change 2011, 2, 45–65. [CrossRef]
- Spinoni, J.; Vogt, J.V.; Naumann, G.; Barbosa, P.; Dosio, A. Will Drought Events Become More Frequent and Severe in Europe? *Int. J. Climatol.* 2018, 38, 1718–1736. [CrossRef]
- Gustafson, E.J.; Sturtevant, B.R. Modeling Forest Mortality Caused by Drought Stress: Implications for Climate Change. *Ecosystems* 2013, 16, 60–74. [CrossRef]
- 6. Grossiord, C.; Forner, A.; Gessler, A.; Granier, A.; Pollastrini, M.; Valladares, F.; Bonal, D. Influence of Species Interactions on Transpiration of Mediterranean Tree Species during a Summer Drought. *Eur. J. For. Res.* **2015**, *134*, 365–376. [CrossRef]
- Peters, M.P.; Iverson, L.R.; Matthews, S.N. Long-Term Droughtiness and Drought Tolerance of Eastern US Forests over Five Decades. *For. Ecol. Manag.* 2015, 345, 56–64. [CrossRef]
- Steckel, M.; Del Río, M.; Heym, M.; Aldea, J.; Bielak, K.; Brazaitis, G.; Černý, J.; Coll, L.; Collet, C.; Ehbrecht, M.; et al. Species Mixing Reduces Drought Susceptibility of Scots Pine (*Pinus sylvestris* L.) and Oak (*Quercus robur* L., *Quercus petraea* (Matt.) Liebl.)—Site Water Supply and Fertility Modify the Mixing Effect. For. Ecol. Manag. 2020, 461, 117908. [CrossRef]
- Gonzalez-Sosa, E.; Braud, I.; Dehotin, J.; Lassabatère, L.; Angulo-Jaramillo, R.; Lagouy, M.; Branger, F.; Jacqueminet, C.; Kermadi, S.; Michel, K. Impact of Land Use on the Hydraulic Properties of the Topsoil in a Small French Catchment. *Hydrol. Process.* 2010, 24, 2382–2399. [CrossRef]
- Kotlarz, J.; Nasiłowska, S.; Rotchimmel, K.; Kubiak, K.; Kacprzak, M. Species Diversity of Oak Stands and Its Significance for Drought Resistance. *Forests* 2018, 9, 126. [CrossRef]
- 11. Croise, L.; Lieutier, F.; Cochard, H.; Dreyer, E. Effects of Drought Stress and High Density Stem Inoculations with Leptographium Wingfieldii on Hydraulic Properties of Young Scots Pine Trees. *Tree Physiol.* **2001**, *21*, 427–436. [CrossRef] [PubMed]
- 12. Nardini, A.; Luglio, J. Leaf Hydraulic Capacity and Drought Vulnerability: Possible Trade-offs and Correlations with Climate across Three Major Biomes. *Funct. Ecol.* **2014**, *28*, 810–818. [CrossRef]
- 13. Nardini, A.; Lo Gullo, M.A.; Trifilò, P.; Salleo, S. The Challenge of the Mediterranean Climate to Plant Hydraulics: Responses and Adaptations. *Environ. Exp. Bot.* **2014**, *103*, 68–79. [CrossRef]
- 14. Doffo, G.N.; Monteoliva, S.E.; Rodríguez, M.E.; Luquez, V.M.C. Physiological Responses to Alternative Flooding and Drought Stress Episodes in Two Willow (*Salix* Spp.) Clones. *Can. J. For. Res.* **2017**, *47*, 174–182. [CrossRef]
- 15. Phillips, R.P.; Ibáñez, I.; D'Orangeville, L.; Hanson, P.J.; Ryan, M.G.; McDowell, N.G. A Belowground Perspective on the Drought Sensitivity of Forests: Towards Improved Understanding and Simulation. *For. Ecol. Manag.* **2016**, *380*, 309–320. [CrossRef]
- 16. Meier, I.C.; Leuschner, C. Belowground Drought Response of European Beech: Fine Root Biomass and Carbon Partitioning in 14 Mature Stands across a Precipitation Gradient. *Glob. Change Biol.* **2008**, *14*, 2081–2095. [CrossRef]
- Chitra-Tarak, R.; Ruiz, L.; Dattaraja, H.S.; Mohan Kumar, M.S.; Riotte, J.; Suresh, H.S.; McMahon, S.M.; Sukumar, R. The Roots of the Drought: Hydrology and Water Uptake Strategies Mediate Forest-Wide Demographic Response to Precipitation. *J. Ecol.* 2018, 106, 1495–1507. [CrossRef]
- 18. Osman, K.T. Forest Soils: Properties and Management; Springer International Publishing: Cham, Switzerland, 2013; ISBN 978-3-319-02540-7.
- 19. Schaap, M.G.; Bouten, W.; Verstraten, J.M. Forest Floor Water Content Dynamics in a Douglas Fir Stand. *J. Hydrol.* **1997**, 201, 367–383. [CrossRef]
- Bonan, G.B.; Shugart, H.H. Environmental Factors and Ecological Processes in Boreal Forests. Annu. Rev. Ecol. Syst. 1989, 20, 1–28. [CrossRef]
- 21. Mader, D.L.; Lull, H.W. *Depth, Weight, and Water Storage of the Forest Floor in White Pine Stands in Massachusetts*; USDA Forest Service Research Paper NE-109; Northeastern Forest Experiment Station: Upper Darby, PA, USA, 1968.
- 22. Walsh, R.P.D.; Voigt, P.J. Vegetation Litter: An Underestimated Variable in Hydrology and Geomorphology. *J. Biogeogr.* **1977**, 4, 253. [CrossRef]
- 23. Putuhena, W.M.; Cordery, I. Estimation of Interception Capacity of the Forest Floor. J. Hydrol. 1996, 180, 283–299. [CrossRef]
- Greiffenhagen, A.; Wessolek, G.; Facklam, M.; Renger, M.; Stoffregen, H. Hydraulic Functions and Water Repellency of Forest Floor Horizons on Sandy Soils. *Geoderma* 2006, 132, 182–195. [CrossRef]
- 25. Gerrits, A.M.J. The Role of Interception in the Hydrological Cycle; VSSD: Delft, the Netherlands, 2010.
- Ilek, A.; Kucza, J.; Szostek, M. The Effect of Stand Species Composition on Water Storage Capacity of the Organic Layers of Forest Soils. *Eur. J. For. Res.* 2015, 134, 187–197. [CrossRef]
- 27. IUSS Working Group WRB World Reference Base for Soil Resources 2014. International Soil Classification System for Naming Soils and Creating Legends for Soil Maps. Update 2015; World Soil Resour Rep. No 106 FAO: Rome, Italy, 2014.
- 28. Ilek, A.; Szostek, M.; Kucza, J.; Stanek-Tarkowska, J.; Witek, W. The Water Absorbability of Beech (*Fagus sylvatica* L.) and Fir (*Abies Alba* Mill.) Organic Matter in the Forest Floor. *Ann. For. Res.* **2019**, *62*, 21–32. [CrossRef]
- 29. Browning, B.L. Methods of Wood Chemistry; Wiley & Sons, Interscience Publishers: New York, NY, USA, 1967; Volume I.

- 30. Technical Association of the Pulp and Paper Industry. *Acid Insoluble Lignin in Wood and Pulp, T 222 Cm-06;* Technical Association of the Pulp and Paper Industry: New York, NY, USA, 2006.
- Technical Association of the Pulp and Paper Industry. Solvent Extractives of Wood and Pulp, T 204 Cm-97; Technical Association of the Pulp and Paper Industry: New York, NY, USA, 2007.
- 32. Technical Association of the Pulp and Paper Industry. *Ash In Wood, Pulp, Paper, And Paperboard: Combustion At* 525°*C, T* 211 *Om*-02; Technical Association of the Pulp and Paper Industry: New York, NY, USA, 2002.
- 33. Zhang, Z.; Chen, Y.; Zhang, Z.; Cui, H.; Lei, Y.; Wang, D.; Sui, J. Water-Holding Characteristics of Litter in Different Forests at the Lianxiahe Watershed. *Front. For. China* 2006, *1*, 413–418. [CrossRef]
- Sato, Y.; Kumagai, T.; Kume, A.; Otsuki, K.; Ogawa, S. Experimental Analysis of Moisture Dynamics of Litter Layers—The Effects of Rainfall Conditions and Leaf Shapes. *Hydrol. Process.* 2004, 18, 3007–3018. [CrossRef]
- 35. Xing, Z.; Yan, D.; Wang, D.; Liu, S.; Dong, G. Experimental Analysis of the Effect of Forest Litter Cover on Surface Soil Water Dynamics under Continuous Rainless Condition in North China. *Kuwait J. Sci.* **2018**, *45*, 75–83.
- 36. Fernald, A.; Gallegos, J.; VanLeeuwen, D.; Baker, T. Evaluation of Litter Hydrology in Ponderosa Pine and Mixed Conifer Stands in Northen New Mexico, USA. *N. M. Acad. Sci.* **2012**, *4*, 121–136.
- 37. Ilek, A.; Szostek, M.; Mikołajczyk, A.; Rajtar, M. Does Mixing Tree Species Affect Water Storage Capacity of the Forest Floor? Laboratory Test of Pine-Oak and Fir-Beech Litter Layers. *Forests* **2021**, *12*, 1674. [CrossRef]
- Hashimi, R.; Huang, Q.; Dewi, R.K.; Nishiwaki, J.; Komatsuzaki, M. No-Tillage and Rye Cover Crop Systems Improve Soil Water Retention by Increasing Soil Organic Carbon in Andosols under Humid Subtropical Climate. *Soil Tillage Res.* 2023, 234, 105861. [CrossRef]
- Raaflaub, L.D.; Valeo, C. Assessing Factors That Influence Spatial Variations in Duff Moisture. *Hydrol. Process.* 2008, 22, 2874–2883. [CrossRef]
- 40. Lehmann, J.; Kleber, M. The Contentious Nature of Soil Organic Matter. Nature 2015, 528, 60–68. [CrossRef] [PubMed]
- Canessa, R.; Van Den Brink, L.; Saldaña, A.; Rios, R.S.; Hättenschwiler, S.; Mueller, C.W.; Prater, I.; Tielbörger, K.; Bader, M.Y. Relative Effects of Climate and Litter Traits on Decomposition Change with Time, Climate and Trait Variability. J. Ecol. 2021, 109, 447–458. [CrossRef]
- Jílková, V.; Straková, P.; Frouz, J. Foliage C:N Ratio, Stage of Organic Matter Decomposition and Interaction with Soil Affect Microbial Respiration and Its Response to C and N Addition More than C:N Changes during Decomposition. *Appl. Soil Ecol.* 2020, 152, 103568. [CrossRef]
- 43. Lin, L.; Webster, J.R. Detritus Decomposition and Nutrient Dynamics in a Forested Headwater Stream. *Ecol. Model.* **2014**, 293, 58–68. [CrossRef]
- Błońska, E.; Piaszczyk, W.; Staszel, K.; Lasota, J. Enzymatic Activity of Soils and Soil Organic Matter Stabilization as an Effect of Components Released from the Decomposition of Litter. *Appl. Soil Ecol.* 2021, 157, 103723. [CrossRef]
- 45. Gulis, V.; Suberkropp, K. Leaf Litter Decomposition and Microbial Activity in Nutrient-enriched and Unaltered Reaches of a Headwater Stream. *Freshw. Biol.* 2003, *48*, 123–134. [CrossRef]
- He, J.; Chen, B.; Xu, W.; Xiang, C.; Kuang, W.; Zhao, X. Driving Factors for Soil C:N Ratio in Woody Plant Communities across Northeastern Qinghai-Tibetan Plateau. *CATENA* 2023, 233, 107504. [CrossRef]
- 47. Ottmar, R.; Andreu, A. Litter and Duff Bulk Densities in the Southern United States. In *Joint Fire Science Program Project* #04-2-1-49 *Final Report;* Fire and Environmental Applications Team: Seattle, WA, USA, 2007.
- 48. Ilek, A.; Kucza, J.; Szostek, M. The Effect of the Bulk Density and the Decomposition Index of Organic Matter on the Water Storage Capacity of the Surface Layers of Forest Soils. *Geoderma* **2017**, *285*, 27–34. [CrossRef]
- 49. Percy, K.E.; Jagels, R.; Marden, S.; McLaughlin, C.K.; Carlisle, J. Quantity, Chemistry, and Wettability of Epicuticular Waxes on Needles of Red Spruce along a Fog-Acidity Gradient. *Can. J. For. Res.* **1993**, 23, 1472–1479. [CrossRef]
- Wang, H.; Shi, H.; Wang, Y. The Wetting of Leaf Surfaces and Its Ecological Significances. In Wetting and Wettability; Aliofkhazraei, M., Ed.; InTech: Atyrau, Republic of Kazakhstan, 2015; ISBN 978-953-51-2215-9.
- 51. Beluns, S.; Platnieks, O.; Sevcenko, J.; Jure, M.; Gaidukova, G.; Grase, L.; Gaidukovs, S. Sustainable Wax Coatings Made from Pine Needle Extraction Waste for Nanopaper Hydrophobization. *Membranes* **2022**, *12*, 537. [CrossRef]
- 52. Shi, H.; Wang, H.; Li, Y. Wettability on Plant Leaf Surfaces and Its Ecological Significance. *Shengtai XuebaoActa Ecol. Sin.* **2011**, *31*, 4287–4298.
- 53. Hanover, J.W.; Reicosky, D.A. Surface Wax Deposits on Foliage of Picea Pungens and Other Conifers. *Am. J. Bot.* **1971**, *58*, 681–687. [CrossRef]
- 54. Jeffree, C.E. The Fine Structure of the Plant Cuticle. *Annu. Plant Rev.* **2006**, *23*, 11–125.
- 55. Muhammad, S.; Wuyts, K.; Nuyts, G.; De Wael, K.; Samson, R. Characterization of Epicuticular Wax Structures on Leaves of Urban Plant Species and Its Association with Leaf Wettability. *Urban For. Urban Green.* **2020**, *47*, 126557. [CrossRef]
- Grünhofer, P.; Herzig, L.; Schreiber, L. Leaf Morphology, Wax Composition, and Residual (Cuticular) Transpiration of Four Poplar Clones. *Trees* 2022, 36, 645–658. [CrossRef]
- 57. Samuels, L.; Kunst, L.; Jetter, R. Sealing Plant Surfaces: Cuticular Wax Formation by Epidermal Cells. *Annu. Rev. Plant Biol.* 2008, 59, 683–707. [CrossRef] [PubMed]
- 58. Steinbauer, M.J.; Davies, N.W.; Gaertner, C.; Derridj, S. Epicuticular Waxes and Plant Primary Metabolites on the Surfaces of Juvenile Eucalyptus Globulus and E. Nitens (Myrtaceae) Leaves. *Aust. J. Bot.* **2009**, *57*, 474. [CrossRef]

- 59. Neinhuis, C.; Barthlott, W. Seasonal Changes of Leaf Surface Contamination in Beech, Oak, and Ginkgo in Relation to Leaf Micromorphology and Wettability. *New Phytol.* **1998**, *138*, 91–98. [CrossRef]
- 60. Gou, X.; Guo, Z. Superhydrophobic Plant Leaves: The Variation in Surface Morphologies and Wettability during the Vegetation Period. *Langmuir* **2019**, *35*, 1047–1053. [CrossRef]
- 61. Brewer, C.A.; Nuñez, C.I. Patterns of Leaf Wettability along an Extreme Moisture Gradient in Western Patagonia, Argentina. *Int. J. Plant Sci.* **2007**, *168*, 555–562. [CrossRef]
- 62. Skaar, C. Wood-Water Relationships; Springer: Berlin/Heidelberg, Germany, 1984.
- 63. Sumi, Y.; Hale, R.; Meyer, J.A.; Leopold, A.; Ranby, G. Accessibility of Wood and Wood Carbohydrates Measured with Tritiated Water. *Tappi J.* **1964**, *47*, 621–624.

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