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Integrated Approach for Wastewater Treatment and Biofuel Production in Microalgae Biorefineries

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Abstract: The increasing world population generates huge amounts of wastewater as well as large energy demand. Additionally, fossil fuel's combustion for energy production causes the emission of greenhouse gases (GHG) and other pollutants. Therefore, there is a strong need to find alternative green approaches for wastewater treatment and energy production. Microalgae biorefineries could represent an effective strategy to mitigate the above problems. Microalgae biorefineries are a sustainable alternative to conventional wastewater treatment processes, as they potentially allow wastewater to be treated at lower costs and with lower energy consumption. Furthermore, they provide an effective means to recover valuable compounds for biofuel production or other applications. This review focuses on the current scenario and future prospects of microalgae biorefineries aimed at combining wastewater treatment with biofuel production. First, the different microalgal cultivation systems are examined, and their main characteristics and limitations are discussed. Then, the technologies available for converting the biomass produced during wastewater treatment into biofuel are critically analyzed. Finally, current challenges and research directions for biofuel production and wastewater treatment through this approach are outlined.

Keywords: microalgae; biorefinery; biofuels; wastewater treatment; greenhouse gases; sustainability



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1. Introduction

Microalgae are a well known group of photosynthetic organisms composed of more than 3000 aquatic species. Most of them are autotrophic, while the remaining are heterotrophic. Microalgae can grow in different wastewaters and convert sunlight and atmospheric CO₂ into biomass. Their cells are able to convert and store energy instead of using it for their growth and development. Therefore, microalgal biomass can be explored as new systems for biofuels production that are a potential substitute to fossil fuels due to renewability, sustainability, and short life cycle of algal growth. Recently, microalgal biomass is recognized as a carbon neutral fuel due to various phytochemical properties of biomass [1]. For these reasons, the development of microalgal biorefineries could be effective for reducing the demand of fossil fuel and lowering greenhouse gas (GHG) emissions, which mitigates the problems associated with global warming and climate changes. Microalgal biomass is considered essential feedstock for biofuel production, because microalgae can be cultivated through the year with higher productivity [2-5]. Moreover, they are highly potential candidates for resource recovery from different types of nutrient-rich wastewater. Nutrient-rich wastewaters are generated from various industrial sectors including aquaculture, dairy, food, pharmaceutical, swine, and textile industries as well as from municipalities. The wastewater derived from the above-mentioned sectors are rich in organic and inorganic nutrients that stimulates eutrophication, which is a major threat to ecosystems. Eutrophication affects mainly fishing industries and causes an annual loss

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of almost 2 BUS\$ (Billion US dollars) [2]. In addition, wastewater contains several toxic chemicals and pathogens, which affects the ecosystem. Moreover, in irrigation, untreated wastewater causes several issues including unnecessary vegetative growth, causing several plant diseases, leading to a decrease in the quantity and quality of crops [6]. Untreated wastewaters also cause chemical and biological contamination of ground water, leading to other negative consequences. Therefore, wastewater needs to be treated before being used in irrigation or discharged into water bodies. Traditionally, wastewater from various industrial sectors is treated using chemical (disinfection, flocculation, neutralization, and oxidation) and/or physical (floatation, grit chamber, and screening) methods [7]. Chemical/physical treatments remain expensive and generate significant quantities of slurry/sludge, which requires a secondary treatment [8]. Wastewater treatment processes consume a lot of energy (2–4% of total national electric power), and need skilled workers to operate the treatment plants which, in turn, have a high capital cost for infrastructures [9–11]. For these reasons, researchers are investigating microalgae-based technologies for resource recovery from wastewater and wastewater treatment. Ren et al. [12] reported that these technologies could represent a green and sustainable approach for wastewater treatment, allowing up to 95% recovery of nutrients from wastewater. During their growth in wastewater, microalgae produce biomass containing lipids, carbohydrates, and other compounds that can be used for biofuel production. Moreover, the treated water can be used in agriculture for irrigation [13,14]. As a result, these technologies can integrate wastewater treatment with biofuel production and water recycling for agriculture. Biofuel can be produced by two-stage wastewater treatment process. In the first step, microalgae are cultivated in wastewater under aerobic conditions, while in the second step they are used for biofuel production under anaerobic conditions [13,15–17]. Moreover, lipids can be extracted for biodiesel production from microalgal biomass [18], while the residual/leftover biomass can converted into different liquid and gases biofuel including bio-alcohols through fermentation [19], bio-H₂ through dark fermentation [3], and bio-CH₄ via anaerobic co-digestion [20,21]. Recent studies reported that algae technology has the potential to produce bioelectricity by a technology based on photosynthetic microbial fuel cells (PhotoMFC) [22,23].

Microalgae can be grown in different modes using various cultivation system including open (traditional), close (modern), turf scrubber, and hybrid (advanced) cultivation systems. However, microalgal technology coupled with wastewater treatment process require optimum nutrient load and composition of wastewater for efficient cultivation in different industrial wastewaters. Other parameters such as physical parameters of photobioreactors (PBRs) (design, volume, and volume to surface ratio) [24,25] and operating parameters (temperature, mixing, illumination, and CO₂ supply) play a significant role in nutrient recovery from wastewater [26]. Therefore, operating parameters need to be optimized to overcome key challenges of microalgal wastewater treatment technology. The algae-based resource recovery can allow the removal of pollutants from wastewater, while algal biomass can be explored for biofuel production, which can reduce the capital expenditures (CAPEX) for wastewater treatment process.

This review summarizes the potential of microalgal refineries for resource recovery from wastewater and different types of biofuel production. In particular, the characteristics of wastewater generated from various sources are examined for microalgae cultivation. Advancements in microalgae cultivation technologies from wastewater and the utilization of microalgal biomass to produce various biofuels are also discussed. Therefore, this review summarizes the prospective of existing microalgae cultivation systems for wastewater treatment and various types of biofuel production. It also discusses the promising advancement for development of integrated algal-based wastewater treatment and biofuel production. Furthermore, key limitations are discussed to overcome the major obstacles and for better development of algal-based technology.

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2. Wastewater Treatment Processes

Wastewater generated from various industrial sectors contains micropollutants [27], nutrients (nitrogen, sulfur, copper, phosphorus) [28], carbon-based pollutants (antibiotics, aromatic hydrocarbons, biocides, phenolic compounds, and surfactants etc.) and heavy metals (cadmium, chromium, copper, mercury nickel, lead, and zinc) [29]. These pollutants demand an appropriate method for removal from wastewater prior to release into lakes or other water bodies [30]. As mentioned above, different biological, chemical, and physical methods can be applied for wastewater treatment [28]. Among them, biological methods are considered as sustainable and cost-effective to remove these pollutants from wastewater [27]. However, still there is no individual method that can be applied to all/various types of wastewater from different industries due to the problems linked with individual methods and varied nutrient load in wastewater. Table 1 summarizes the advantage and disadvantage of conventional and modern wastewater treatment methods. However, traditional wastewater management systems exhibited several limitations, including larger area obligation, rigorous power requirement, as well as extensive maintenance, which increases the overall costs of the wastewater treatment process [31]. Therefore, alternative wastewater treatment process needs to be explored for efficient resource recovery from wastewater [32]. Therefore, the microalgae-based wastewater treatment process is considered more effective as well as having the potential for atmospheric CO₂. Microalgae uses CO₂ and available pollutants from wastewater for cell proliferation, while algal biomass can be used for biofuel production [33].

Microalgae are potential candidates for wastewater treatment and biofuel production due to faster growth rate, zero waste generation, and can be cultivated on non-agricultural land [4]. Therefore, integration of wastewater treatment process with microalgae cultivation system could be a forthcoming green technology for biofuel and energy generation. The environmental and economic benefits are occurring by the conjugation of wastewater treatment process with microalgae cultivation for biofuel production. Several studies reported that the different types of wastewater have been successfully utilized for cultivation of different species of microalgae, which can substantially decrease the operational costs of wastewater treatment and biofuel production [3,26]. While during microalgae growth water and nutrients, light and CO₂ are necessary, among them, wastewater provides water and nutrients, while sunlight and atmospheric CO2 can be used for cost effective cultivation [4]. Wastewater is the most suitable resource for microalgae cultivation due to various advantages such as (i) cheaper organic and inorganic carbon rich medium for growth; (ii) support large scale cultivation; (iii) able to provide ample trace elements; and (iv) able to cope up existing infrastructure wastewater treatment [34]. Several studies demonstrated the potential of microalgae-based wastewater treatment [28,30,31,35], and integrated biorefinery approaches have been proposed for biofuel production [36]. Recently, various species of microalgae have been explored for a variety of wastewater treatment, including brewery wastewater [37], domestic wastewater [38], textile wastewater [39], pharmaceutical waste streams [40], slaughter-house industry [41], heavy metal-containing wastewater [42], palm oil mill effluents [32], starch-containing textile wastewater [43], and agro-industrial wastewater [44]. Although a microalgal cultivation system for wastewater treatment offers several advantages, different challenges are also associated, which need to be remitted as summarized in Table 1.

 Table 1. Comparison of different wastewater treatment process for pollutant removal from wastewater.

Treatment Process	Principle	Source of Pollutants	Removal Efficiency	Advantage	Disadvantage	Reference
Adsorption	Adsorption of specific contaminate on the surface of absorbent	Agricultural, industrial, and municipal wastewater with organic pollutants	96%	Chemical less, eco-friendly, economical, better metal strap capability	Low selectivity, difficult maintenance, formation of waste products	[45,46]

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 Table 1. Cont.

Treatment Process	Principle	Source of Pollutants	Removal Efficiency	Advantage	Disadvantage	Reference
Adsorption, membrane-filtration, and photo-catalytic degradation (Hybrid)	Pollutant removal by serial treatment	Industrial wastewater with organic pollutants	88–92% of COD; 85–91% of detergents & 91–98% of TS	More efficient, proved highly treated water	Difficult up-scaling	[47]
Advance oxidation	Removal of contaminate through oxidation of reactive species	Industrial wastewater rich with organic and pesticidal contaminates	53–96% of COD & 21–85% of TOC	Broad application, removal of odor molecules, efficient for removal of organic contaminants	High cost, incomplete removal of pollutants	[48]
Biochar	Absorption contaminates for removal or degradation	Wastewater rich with dyes, heavy metals, organic inorganic pollutants, and phenolic compounds	65–99% of dyes & >90% of phenols	Economical, large surface area, more pores, highly efficient	Low removal efficiency of raw biochar, less sustainable	[49]
Biogenic Nanoparticles	Reduction or oxidation of metals by natural chemicals-based nanoparticles	Radio-active contamination, inorganic and organic pollutant	75–99% of dyes & 66–85% of heavy metals	Sustainable, less toxic, inexpensive, less energy requirement	Instability, tricky recovery of intracellularly synthesized nanoparticles	[50,51]
Biological method	Assimilation and dissimilation of pollutants	Nitrates, phosphates, dairy waste	>90% of COD & 38–90% of N ₂	Economical, high biodegradability, efficient elimination of pollutants	Slow, requires constant maintenance, lower applicability, performance limited by operational conditions	[52,53]
Microalgae	Uptake of pollutants as nutrients source for cell proliferation	Municipal and industrial wastewater rich with heavy metals, dyes, organic and inorganic pollutants	20–98% of TP	Higher removal efficiency, non-toxic, less energy requirement, self-sustainable, produced biomass can be used for biofuel production	Performance limited due to operational conditions as well as type of wastewater, challenging biomass recovery, demands larger land	[54]
Coagulation	Dissociation and hydrolysis	Heavy metals, textile, petroleum, cosmetics wastewater	>70% of COD & 90–100% of heavy metals	Ecofriendly, lower operational cost, efficient pollutant removal, energy efficient	Higher maintenance cost, difficult up-scaling, expensive	[55,56]
Filtration	Separation via porous membrane	Industrial, municipal, and textile wastewater	77% of COD; 99% of dyes & TC 74% of TN	Easy operation, cost-effective, and capable to remove suspended solid, alkalinity, inorganic and organic pollutants	Clogging of filter, limited removal of micro pollutant, higher cost of raw material	[31,57,58]
Filtration and coagulant- flocculation (Hybrid)	Integration of filtration, coagulating and flocculant	Phenolic compounds, organic pollutants, suspended solids	36% of COD; 81% fatty matter & 92% of TS	Highly efficient for removal of pollutants, lesser energy requirement	High maintenance cost	[59]

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Table	1.	Cont.
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Treatment Process	Principle	Source of Pollutants	Removal Efficiency	Advantage	Disadvantage	Reference
Microbial electro-chemical Technology	Oxidation or reduction of pollutants by respiring microbes	Recalcitrant matter, industrial, domestic and food-processing wastewater	>25–63% of COD	Wide applicability, production of electricity, and other valuable commodities	Challenging up-scaling, high cost	[60]
Nanomaterials	Play an action as absorbent for the photolytic degradation of contaminate	Inorganic and organic emerging pollutants, petrochemicals as well as heavy metals	90–100% of heavy metals	Highly efficient with higher adsorption efficiency, friendly with other techniques	Less ecofriendly, expensive, toxic	[51,61]

3. Microalgae Cultivation Systems

Nowadays, different types of cultivation systems are used for small scale and large scale cultivation of microalgae for different purposes. However, the choice of microalgae cultivation system usually depends on the selection of microalgal species, nutrient supply, as well as end application microalgal biomass. Based on the different importance of microalgae cultivation systems, these can be classified as open and closed cultivation systems.

3.1. Open Systems

In open systems, circular ponds, open ponds, raceway ponds, tanks, and unstirred ponds are mainly employed for large-scale cultivation of microalgae [62]. Open cultivation systems are reported as more economic than the closed cultivation systems due to low construction and operation cost. Therefore, these are mostly employed as well as reported to contribute up to 95% of total microalgal biomass generation. Despite these advantages, there are some disadvantages of an open system, like evaporation of water, CO₂ diffusion to the environment, poor light utilization by cells, land requirement, etc. [63]. Moreover, there are biotic factors that can limit the biomass production due to contamination of unknown species. Figure 1 showed the most commonly used open ponds cultivation system for large-scale microalgae cultivation.

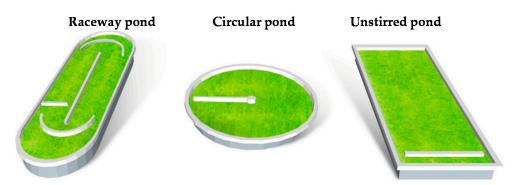


Figure 1. Schematic representation of close cultivation systems of microalgal cultivation.

3.1.1. Raceways Ponds

The raceway pond or stirred wheel open pond was firstly developed by Oswald in the 1960s [64]. The raceway ponds are the more common open cultivation system for microalgae biomass production in the last forty years [65]. The raceway pond is usually shallow, and depth ranged from 15 to 25 cm. Typically, they are formed either as a single-channel or as multi-channel system, which are designed by linking single raceways together. This open system is mainly well-avowed for the cultivation of four algal species: *Chlorella*, *Dunaliella*, *Spirulina*, and *Hematococcus* [66]. The reported biomass productivity of a raceway pond is 60 to 100 mg DW/L d [67].

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3.1.2. Circular Ponds

Circular ponds are oldest open system for microalgal biomass cultivation and are primarily used for largescale cultivation, particularly for *Chlorella* sp. in South East Asia [68]. Usually, this type of ponds is 40 to 50 m in diameter and 20 to 30 cm in depth [65]. In circular ponds, a long rotary arm is set in the middle of pond that performs the function of a paddlewheel. There is no doubt that the circular system is more efficient in contrast to the unstirred, but the cultivated microalgae are still prone to surrounding contamination [64].

3.1.3. Unstirred Open Ponds

As the term itself explains, an unstirred open pond is a system without mixing or devoid of stirrer unit. Unstirred ponds provide an inexpensive and clear way to produce some commercially microalgae species such as *Dunaliella salina* [69]. These ponds are simple in terms of construction without any special requirement. However, the lack of stirrer causes poor mixing that sometimes can be a limiting factor for large scale biomass cultivation. To ensure proper light penetration, unstirred open systems are consisted of <1 m in depth. Several reports revealed that transparent plastic covering can act as a promising solution to overcome the contamination related issues in unstirred open systems [64].

3.2. Closed Systems

In order to resolve the issues of an open cultivation system as well as to enhance the biomass productivity, the closed system came into existence. There is no doubt that the open systems are still preferred over a closed cultivation system to produce biomass at a low cost for production of low value products. On the other hand, to cultivate microalgal biomass for high-value products, there is a requirement of aseptic conditions and well-constructed systems with optimized operation conditions [63]. Despite these benefits of open systems over closed systems, there are a number of problems to make their use for pilot scale production [65]. Closed systems are mostly referred as photobioreactors (PBRs) as they are devoid of direct gases exchange and contaminants from the surrounding environment [70]. Recently, many designs of PBRs have been introduced to improve cultivation and reduce costs. The details of different type of PBRs with common configurations are discussed in the below subsection and Figure 2 showed the close cultivation system.

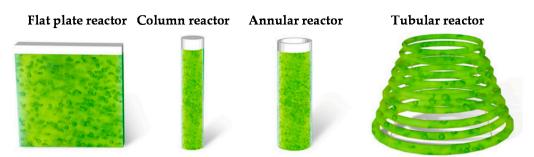


Figure 2. Schematic representation of close cultivation systems of microalgal cultivation.

3.2.1. Flat Plates

It is one of the common PBRs and appears as a rectangular-shaped box, commonly employed for pure cultures microalgae cultivation [64]. These type of PBRs are made up of transparent material such as glass, plexiglass, plastics, polycarbonates, etc. It can be situated indoor under an artificial light source or outdoor under sunlight [69,71]. Due to the short light path, light can easily penetrate to culture medium and mixing is mediated via air sparger in the form of air bubbles [64]. The common examples of flat plate PBRs are horizontal, vertical, v-shaped, and inclined flat plate PBR.

3.2.2. Tubular Reactors

Tubular PBRs are widely used on commercial level production of high-value products. They are formed of polyvinyl chloride or polypropylene acrylic that are usually transparent

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with small internal diameter to overcome light related problems [64]. There is an air pump for the sake of mixing and agitation of microalgal culture. There are two main types of tubular PBRs such as vertical (bubble and airlift column) and horizontal tubular (inclined, spiral, helicoidal, etc.). The other PBRs like helical tubular, polythene bags, and sleeves are also employed for microalgal biomass cultivation.

Vertical reactors: Bubble columns and airlift PBRs are mainly included in vertical tubular type PBRs. Both bubble and airlift are connected to an air sparger at the bottom of these PBRs [64]. Bubble columns are cylindrical in shape with a 20 cm internal diameter (ID) and the height is commonly greater than the twice of ID. Bubble column PBRs are highly considered due to lower fabrication cost, higher volume to surface ratio, proper energy and mass transfer, as well as efficient gaseous mixing and O_2 release [63]. In this type of PBR, the gas flow rate is the main parameter that needs to be regulated during operation, which affects the growth of microalgae. The main drawbacks of bubble column PBRs are the formation of biofilms on its wall, which are prone to photoinhibition [72].

The development of an airlift column is regarded as an improvement of bubble column and consists of two internal interconnecting zones. One zone is referred to as riser, while the other is called a downcomer zone [73]. In a riser zone, gas mixture flows upward from the bottom that is mediated via a sparger, and in a downcomer zone, there is no gas supply but medium flows toward the bottom; thus, culture medium circulates within both zones. Based on circulation mode, the airlift PBRs are designed as two forms, i.e., the internal loop and external loop. In an airlift column, the flow of culture medium circulates between both zones, which means it passes through light and dark phases that is referred to as the flashing light effect [74].

Horizontal reactors: Horizontal tubular PBRs are widely used at a commercialized scale for the generation of microalgal biomass in closed systems. These are made up of tubes and can be arranged in various possible orientations, viz., helicoidal, spiral, inclined, horizontal, etc., but all works on the same principal [64]. In these systems, there is maximal exposure of microalgae to light. In these PBRs, the reported ID can be ranged from 10 to 60 mm and the height can be of several hundred meters. For circulation, there is an air pump for mixing and agitation. Horizontal tubular photobioreactors (PBRs) are facilitated with better air-residence time that is reported to provide more CO_2 in dissolved form in the growth medium [64,75].

Helical reactors: These are made up of polyethylene with an ID of 2.4 to 5 cm and are spirally wound on cylinder-shaped, truncated pyramid, or truncated cone shaped structures for support purposes [74]. For circulation, a centrifugal pump is used in helical tubular (tube-shaped) PBRs. Helical tubular PBRs are considered as highly effective and generally employed for small volume of microalgal suspension. The main drawback of this type of closed system is the fouling phenomenon that can cause tubes blocking due to biofilm formation [64].

3.3. Advance Cultivation Systems

There are few other cultivation systems that can be used for growth of microalgae and the important ones are discussed below.

3.3.1. Algal Turf Scrubber (ATS) Cultivation Systems

An ATS is a modified cultivation system, which was firstly introduced in the early 1980s by Professor Walter Adey [76]. An ATS cultivation system provides a downward sloped surface for water or influent to flow across, in a pulsed or continuous manner as showed in Figure 3 [77], which promotes growth of macroalgae [30]. In the ATS system, microalgae grow efficacity due to appropriate uptake of inorganic compounds from wastewater and releasing dissolved oxygen (DO) through photosynthesis. ATS microalgae cultivation systems improved nutrients/pollutant removal efficiency as well as allows faster growth rate due to superior photosynthetic activities, which releases a higher amount of O₂. ATS systems are also exploited for wastewater treatment or resource recovery from

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different type of wastewater such as an animal farm (dairy farm, aquaculture) wastewater, sewage wastewater, and industrial streams wastewater. Furthermore, an ATS cultivation system offers 5–10 times higher microalgal biomass productivities as compared to other types of land-based cultivation, but are limited due to lower handling capacity [30]. Moreover, in the ATS cultivation system, microalgae cultivation is able to grow in a continuous mode without centrifuging the biomass during production [77–79]. Therefore, higher biomass productivity and nutrient removal efficiency and easy biomass harvesting potentially decreases the overall cost involved in resource recovery from wastewater [80]. Moreover, enhanced growth rate and subsequently harvested biomass from an optimized ATS cultivation system could be utilized for various applications [81]. Siville and Boeing [82] reported that an ATS is more efficient for resource recovery from wastewater applying the microalgal technology. Several results indicated that ATS microalgae cultivation systems can permit the higher rate removal of nitrate from various wastewaters and produced biomass can be exploited for commercial purposes. The key advantages and disadvantages are discussed in Table 2 and compared with other cultivation approaches.

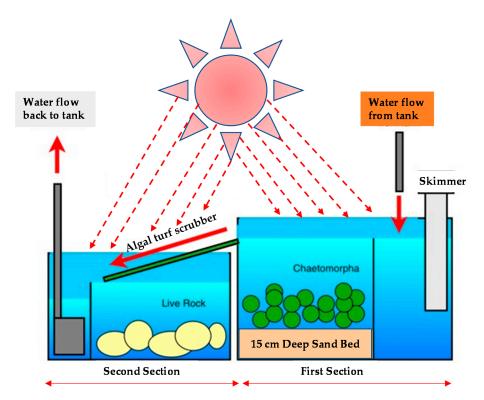


Figure 3. Algal Turf Scrubber (ATS) cultivation system for resource recovery from wastewater.

3.3.2. Hybrid Cultivation Systems (HCS)

As the name suggests, it comprises both a closed and open system. It involves integration of two or more types of a cultivation system with an integration of biomass harvesting system. The first part of cultivation occurs in a closed system or open system and then the subsequent phase is carried out for biomass harvesting [74,83]. A hybrid cultivation system provides better growth conditions that lead to efficient resource recovery from wastewater. Therefore, hybrid cultivation systems are more favorable for higher biomass production, which can reduce the overall CAPEX of the microalgal technology [30]. Several researchers carried out the cultivation of microalgae using different cultivation approaches and found that a two stage hybrid cultivation system allows to attain increased biomass [62]. Adesanya et al. [84] conducted similar studies and found that a two-stage hybrid cultivation system (integrated airlift tubular PBRs with raceway ponds) allow to attain higher biomass productivity for lipid production. Yun et al. [85] cultivated *Parachlorella* sp. JD076 for 47 days using a hybrid cultivation system and found that a

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hybrid cultivation system has a high efficiency. The result showed that hybrid cultivation systems showed higher 40% of biomass and 62% lipid yields as compared to small-scale open raceway ponds. Moreover, Swain et al. [86] investigated the effect of two-stage cultivation process for biodiesel production from *Tetraselmis* sp. using the simulated dairy wastewater. Two stage cultivation showed efficient nutrient removal efficiency from simulated dairy wastewater and *Tetraselmis* sp. biomass was used for biodiesel production. Therefore, HCSs are highly recommended for cost-effective wastewater treatment and resource recovery, while the produced algal biomass can be explored for various types of biofuel production. Moreover, the main advantage and disadvantage of hybrid cultivation is summarized in Table 2 and compared with other cultivation systems.

Table 2. The main benefits and drawbacks of different microalgae cultivation systems.

Cultivation System	Advantage	Disadvantage
Open system	Easy construction and operation Lower cost due to less energy consumption Large open outdoor is widely used to cultivate microalgae for biofuel production using wastewater	Poor light utilization Evaporative losses of water Require Larger areas Chance of contamination Climatical dependency Biomass harvesting challenging and expensive due to lower cell density
Closed system	Optimal growth parameters can be controlled Minimize the CO ₂ and water losses Higher biomass productivity Lesser contamination	Capital expenditures (CAPEX) and operating expense (OPEX) are high Over-heating and biofouling Up-scaling very expensive
Algal turf scrubber (ATS)	Very low and fuel production cost Higher biomass productivities and easy biomass harvesting over open cultivation system Low maintenance and monitoring Carbon sequestration from environment	Need sufficient space but lesser as compare to open cultivation system Lower wastewater handling capacity Need considerable infrastructure
Hybrid cultivation system (HCS)	Higher biomass productivity Lesser contamination risk Low maintenance and monitoring Carbon sequestration from environment	Higher CAPEX and OPEX of process Require large infrastructure Need high maintenance and continuous monitoring

Therefore, interest among the different microalgae cultivation systems and efficient approaches are developed to accomplish optimal and economical production of microalgal biomass. All types of cultivation approach showed a varied biomass production rate, which remains changed due to different operational conditions. However, to highlight the more appropriate microalgae cultivation system, critical comparative studies are recommended to determine an appropriate cultivation system for wastewater treatment and biofuel production. Table 2 highlights the main pros and cons of the above discussed cultivation systems. In summary, it can be concluded that open cultivation systems generally employed for wastewater treatment due to capacity for lager volume treatment, which can reduce the overall cost of the wastewater treatment process. The close cultivation system is mostly used to achieve higher biomass productivity for production of high value-added compounds from algal biomass. However, the development of an advanced hybrid cultivation system can allow to improve the biomass yield, which can allow to offset the overall cost.

4. Mode of Microalgae Cultivation

Microalgae are prokaryotic organisms that use atmospheric CO_2 and sun light during photosynthesis as well as being able to use organic carbon from wastewater. The design (mostly closed system) and strategy used in cultivation of microalgae limit sunlight penetration for commercialized production of microalgae, but fluorescent light can be used to overcome these limitations. However, an artificial light source for cultivation increases CAPEX and operating expense (OPEX) of the process [87]. Microalgae are able

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to grow under the presence and absence of light using organic and inorganic carbon sources under three different cultivation modes, i.e., (i) phototrophic; (ii) heterotrophic; and (iii) mixotrophic. The key advantage and disadvantage of each mode is described in Table 3. However, among the different cultivation modes, phototrophic cultivation offers slow growth rate and lower biomass productivity due to lesser light availability when cell number and diameter increased due to self-shading [88].

Table 3. The key variation in different microalgae cultivation mode as well as the main advantage and disadvantage.

Type of Cultivation Mode	Energy Supply	Type of Carbon Source	Advantages	Disadvantages
Autotrophic	Light	Inorganic carbon source (CO ₂)	Lower CAPEX and OPEX Higher productivity of pigments Ability to remove nutrients from wastewater	Lower biomass production rate
Heterotrophic	Organic carbon source	Organic carbon source (glucose etc.)	Higher biomass production rate	Higher cost of carbon sources Chance of microbial contamination Higher CAPEX and OPEX
Mixotrophic	Light and organic carbon source	Inorganic and organic carbon source (CO ₂ & glucose)	Higher growth rate Ability to remove nutrients from wastewater	Higher cost of carbon sources Chance of microbial contamination Higher CAPEX and OPEX

There are more cultivation modes with external organic carbon sources categorized as mixotrophic and heterotrophic cultivation modes. In this cultivation mode, growth rate increased by 2 to 5 folds [89]. In the heterotrophic mode, microalgal cells use an organic carbon source for their cell proliferation. Under the heterotrophic mode, microalgae cells produce CO_2 and are unable to mitigate CO_2 , which is a major drawback. Under the mixotrophic mode, microalgae cells utilize inorganic carbon (CO_2) and organic carbon for cell proliferation. Therefore, these trophic modes have a significate role in cultivation of microalgae; however, phototrophic cultivation is more sustainable and offers several environmental benefits [4,26]. Therefore, appropriate cultivation mode needs to adopt for high-rate biomass productivity and to attain environmental sustainability.

5. Microalgal Biorefinery for Biofuels Production

Microalgae biomass are an abundant resource of several molecules and molecules transformed into various types of biofuels. Biofuel can allow to meet the growing demands of fossil fuels, which can be generated by the microbial conversion of biomass derived from different sources. Among the various types of biomass, algal biomass is highly considered for biofuel production due to higher lipid and sugar content in algal biomass [3,4,26]. Microalgal biomass has the potential for production of a broad range of biofuel through different routes. Microalgal biomass can be transformed into biodiesel through the transesterification of lipid [90], bioethanol, bioH₂ through the fermentation of carbohydrates [3,91], while bioCH₄ via co-digestion [92]. Microalgae based biofuels production offers several advantages such as (i) arable land not required for mass cultivation of microalgae; (ii) it offers high photosynthesis activity as compared to terrestrial plants, which results in high biomass productivity as well as offer higher CO₂ mitigation; (iii) does not require agriculture land, which is not allowed to complete with food crops; (iii) fertilizer or addition of extra nutrient not required; (iv) and being able to remove the

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pollutants from wastewater [2,3,27]. Several microalgal species lives in the environment; among them, only a few are explored for various commercial applications in different sectors. Figure 4 summarized the key biofuels produced from microalgae and highlight the key benefit of microalgal-based biofuels. Furthermore, microalgae biomass based different biofuel are briefly discussed in the below subsection.

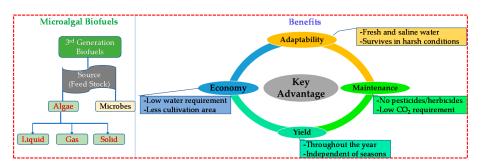


Figure 4. Key advantages of third generation biofuel.

5.1. Biodiesel

In the last 30 years, the key research focus of the U.S. Department of Energy Aquatic Species program was to produce biodiesel from microalgal biomass, which was cultivated in high-rate algal ponds (HRAPs) [93]. During the research investigation in this program, it was concluded that optimum climatical conditions allow the enhanced biomass productivity with high lipid content as compared to terrestrial crops. Microalgal growth rates are very high under optimum conditions, which allows productivity ~50–100-tons algae dry biomass/hector/year with a 25–50% lipid content in the dry biomass [94]. These productivities can increase by applying the modern approaches such as alteration in metabolic pathways of microalgal species. However, in the microalgal biomass, lipids content and quality differs among species to species and even strains within a species as well as many other factors such as operational parameters, while nitrogen limitation in growth medium improves lipid productivity [94]. Microalgae is able to produce a considerable quantity of lipids in the range of ~5000 to 100,000 L/hector/day, while microalgal derived biodiesel have a high energy content (39 to 41 KJ/g). Moreover, biodiesel can be produced by direct transesterification with heterogenous/homogenous catalyst or via in-situ(trans)esterification of the microalgal lipids [95]. Several studies have discussed the recent advancement in various technologies for biodiesel production from microalgal lipids and highlighted the drawbacks of traditional transesterification method for biodiesel production [96,97]. Furthermore, different microalgae have various physio-chemical properties, which make a significant variation on microalgae-derived biodiesel that cause a different impact on engine performance. However, it is essential to analyze the physical properties of microalgal-derived biodiesel to improve the engine performance, combustion, and emission. The physical properties of biodiesel derived from different microalgal species are summarized and compared with the American Society for Testing and Materials (ASTM) D6751 standards as showed in Table 4 [98].

5.2. Biogas

Microalgal biomass contains high various molecules, which can transform into biogas via anaerobic digestion (AD) process. During the AD process, generated biogas can be utilized for different purposes, such as electricity, heat, and/or transport. Moreover, AD process offers lower CAPEX and OPEX in comparison to available biofuel production technologies such as pyrolysis and gasification. In addition, the AD process can be used for nutrient recycling from wastewater, and after, AD process treated water could be recycled for microalgae cultivation [99,100]. In most of the investigations for biogas production using microalgal biomass, it has been carried out via methanogenesis in co-digestion system with sewage sludge (SS) or with other co-substrates [101]. Solé-Bundó et al. [102] investigated the effect of co-digestion of microalgae biomass and primary sludge (PS)

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on biogas productivity as well as on microcontaminants removal efficiency. The results showed that co-digestion enhances 65% CH₄ productivity as well as microcontaminants removal efficiency achieved up to 90%. Table 5 summarizes the microalgae anaerobic co-digestion (AcoD) with different type of co-substrates under different operating conditions. However, operational conditions have a significant role in methane yield, therefore optimum conditions need to be considered for enhanced methane production [103–107].

Table 4. Physicochemical properties of microalgal biodiesel. Reproduced from [98], Elsevier B.V.: 2021.

Algae Species	CN	CP (°C)	CV (MJ/kg)	FP (°C)	PP (°C)	SG	Viscosity (Centipoise) at 40 °C
ASTM) D6751 standards	40	-	-	52	-	0.86-0.90	1.9-6.0
Botryococcus braunii	-	_	50.00	138	-	0.85	5.34
Chlorella protothecoides	52	-27	40.04	124	-11	0. 87	2.80
Chlorella vulgaris	60	0	17.44	140	-11	0.87	4.80
Cladophora	-	_	33.60	110	-12	0.89	3.80
Crypthecodinium cohnii	46.5	16.1	39.86	95.0		0.91	5.06
Dunaliella salina	50.5	0	34.00	129	-6	0.85	2.40
Entromorpha	50	-1	39.94	194	-6	0.86	3.12
Nannochloropsis oculata	46	3.39–12.14	16.80	180	-4	0.85	5.76
Neochloris oleoabundans	55	-10	39.76	126	-12	0.89	5.54
Spirogyra	-	3	13.62	78	-7	0.88	4.40
Spirulina	70	-3	34.50	130	-9	0.99	0.27 at 300 °C
Stoechospermum marginatum	63	_	42.052	128	-	0.89	4.84

Note: CN: cetane Number; CP: cloud point; CV: calorific value; FP: flash point; PP: pour point; SG: specific gravity; V: viscosity.

Table 5. Anaerobic co-digestion of microalgal biomass of different species for biogas production with various other co-substrate.

Microalgae	Co-Substrate	Mixture Ratio	T (°C)	MYC (mLCH ₄ /gVS)	YI (%)	Reference
Chlorella sorokiniana	WAS	25:75 (VS)	37	442	26	[108]
Chlorella sorokiniana & Scenedesmus sp.	WAS	9:91 (VS)	37	400	11	[109]
Chlorella sorokiniana & Scenedesmus sp.	WAS	9:91 (VS)	37	560	28	[109]
Chlorella vulgaris	WAS	75:25 (COD)	35	107	6	[110]
Chlorella vulgaris	PS	50:50 (COD)	35	283	16	[110]
Chlorella vulgaris	PS	75:25 (COD)	35	293	13	[110]
Chlorella vulgaris	Chicken litter & glycerol	30:67:3 (TS)	37	131	16	[111]
Chlorella sp.	WAS	41:59 (TS)	37	468	23	[112]
Chlorella sp.	WAS	21:79 (VS)	37	253	5	[113]
Chlorella sp.	Septic sludge	50:50 (VS)	35	547	84	[114]
Dunaliella salina	Olive mill solid waste	50:50 (VS)	35	285	48	[115]
Lipid-spent Botryococcus braunii	WAS	75:25 (VS)	35	393	7	[116]

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Table 5. Cont.

Microalgae	Co-Substrate	Mixture Ratio	T (°C)	MYC (mLCH ₄ /gVS)	YI (%)	Reference
Micractinium sp.	WAS	21:79 (VS)	37	236	0	[113]
Nannochloropsis salina	Corn cob	25:75 (v/v)		610	18	[117]
Nannochloropsis salina	Corn silage	14:86 (v/v)	37	660	15	[117]
Scenedesmus quadricauda	WAS	49:51 (VS)	35	222	12	[92]
Scenedesmus sp.	Opuntia maxima	25:75 (VS)	37	234	65	[118]
Scenedesmus sp.	Opuntia maxima	10:90 (VS)	37	166	7	[118]
Scenedesmus sp.	Paper sludge	74:26 (VS)	37	173	35	[119]
Scenedesmus sp.	Pig manure	15:85 (VS)	37	319	5	[120]
Scenedesmus sp. and Chlorella sp.	SS	40:60 (VS)	35	237	2	[121]
Selenastrum capricornutum	SS	50:50 (VS)	33	392	9	[122]
Spirulina platensis	SS	67:37 (VS)	35	343	15	[123]
Spirulina platensis	Pretreated switchgrass	50:50 (TS)	50	354	10	[124]
Taihu blue algae	Corn straw	20:80 (VS)	35	325	114	[125]

Note: T: temperature; YI: yield increased; MYC: methane yield with co-digestion; WAS: waste activated sludge; SS: sewage sludge; PS: primary sludge.

During the AcoD process, the produced biogas contains CH₄ (50–65%) and CO₂ (40– 50%) and traces amount of H₂SO₄, N₂O, which need separated from the methane [103]. Therefore, innovative and sustainable biogas upgrading technologies are immediately required. However, there are several traditional upgrading technologies (chemical adsorption, membrane separation, and pressure swing adsorption) available, which demand a high cost [126]. Recently, microalgae-based biogas upgrading are under investigation to avoid the major disadvantages of conventional biogas upgrading [127]. Table 6 summarizes the different microalgal based biogas upgrading, and biogas upgrading can be achieved with approximately 97%, which differs based due on variation in the reactors and operating parameters as well as the selected microalgal strains for biogas upgrading [126]. Toledo-Cervantes et al. [128] reported the microalgal based biogas upgrading results maximum elimination of CO₂ and H₂S, while CH₄ achieved with a purity of 95%. Meier et al. [5] studied the impact of a light/dark cycle on CO₂ removal from biogas employing the photosynthetic microalgal biogas upgrading. The results showed that microalgal based CO₂ sequestration resulted around 89–93% CO₂ removal. Similarly, Franco-Morgado et al. [129] investigated the effect of the light illumination system on the upgrading of synthetic biogas and observed the 99.5% of H₂S with 91.5% removal of CO₂.

Prandini et al. [130] reported that microalgae-based biogas upgrading with wastewater treatment process can accelerate nutrient removal efficiency, which results in pure bioCH₄. Furthermore, cultivated biomass can be used for AcoD process as feedstock to produce biogas, or it can be explored for microalgal biodiesel production. Therefore, a microalgal based refinery needs to adopt in AD plants for sustainable biogas upgrading and simultaneous low-cost biogas production via methanogenesis.

Table 6. Potential of different microalgae species for biogas upgrading technologies.

M:1 Ci	Cultivation	n System	H ₂ S	CO P 10/	0.0/	CH ₄ Recovery %	n (
Microalgae Species	In-/Outdoor	System	Removal	CO ₂ Removal %	Removal % O ₂ % C		Reference
Chlorella sp.	Outdoor	HRAP	Yes	95	0.1-2.0	94	[131]
Chlorella vulgaris				55.39		80.4	
Scenedesmus obliquus	O1-1	EPB		62.31	0.62	82.64	[132]
Neochloris oleoabundans	- Outdoor	ELD		54.39		80.06	[132]

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Table 6. Cont.

M:	Cultivation System		H ₂ S	CO P 10/	0.0/	CH D	.
Microalgae Species	In-/Outdoor	System	Removal	CO ₂ Removal %	$O_2\%$	CH ₄ Recovery %	Reference
Geitlerinema sp. (61.5%), Staurosira sp. (1.5%) & Stigeoclonium tenue (37%)	Indoor	HRAP	Yes	98.8	0.03	97.2	[128]
Mychonastes homosphaera	Indoor	HRAP	Yes	98.8	0.7	96.2	[133]
Planktolynga brevicellularis (81%), Stigeoclonium tenue (14%) and Limnothrix planktonica (5%)	Indoor	HRAP	Yes	72–79	1.2-0.7	81	[134]
Scenedesmus sp.	Indoor	EPB	Yes	66.7	17.8	64.7	[129]

Note: HRAP: high-rate algal ponds; EPB: enclosed photobioreactor bag.

5.3. Biohydrogen

Biohydrogen (BioH₂) is the next-generation fuel source having remarkable benefits compared to other liquid or gaseous fuel [135-139]. Microalgae can naturally produce bioH₂ through photolysis, or its biomass can be used as feedstock for fermentative bioH₂. The photolysis is classified into indirect and direct photolysis. During the direct photolysis process, water is split into electron (e⁻) and proton (H⁺), then e⁻ moves to Fe-Fehydrogenase, which forms bio-H₂. But the whole process is dependent on the hydrogenase enzyme. However, a small amount of oxygen can inhibit the activity of hydrogenase resultant low yield of bioH₂ [3,140,141]. The utilization of sulphur deprived medium or addition of oxysorb in medium might be a fundamental solution for reducing the amount of oxygen in the surrounding atmosphere. Whereas in indirect photolysis, the produced oxygen utilizes for the metabolic process to produce other value-added biomolecules; the resultant amount of oxygen is reduced. In the breakdown of biomolecules, e is generated, which can undergo for plastoquinone and then transferred into the electron transport chain where bio-H₂ is produced by the help of Fe-Fe hydrogenase. Furthermore, genetic modification in the bio-H₂ase gene can increase the resistance ability of the hydrogenase enzyme. Goswami et al. [3] reviewed the different genetic approach used for enhancement of bio-H2 production in microalgae. A different genetic approach such as (i) overexpression of PSII gene, cytochrome b6f, hemA, and lba gene, translational repressor NAB1 protein; (ii) knockout of light-harvesting gene, IFR1 protein, OEE2 gene; (iii) cloning of pyruvate oxidase gene, DT hydA gene; (iv) antisense transformation of sulp/sulp2 gene or amino acid substitution; (v) insertion of GAL4 gene, CRY1, and CRY2 gene, VP 16 and other light inducible system can enhance the bioH₂ production in microalgae [3]. Figure 5 shows the different genetic approach, which can enhance the bioH₂ production inside the microalgal cells.

The dark fermentation and photo fermentation of using microalgal biomass is quite beneficial in terms of bioH₂ yield. The microalgal biomass contains a high amount of carbohydrates, protein, and lipids; therefore, it is recognized as potential feedstock for fermentative bioH₂ production [142]. The fermentation for bioH₂ production depends on the fermentative microorganisms and biomass pretreatment [3]. The fermentative microorganisms consume the biomass as substrates and convert these substrates to organic acids, CO₂, and acetate [143]. The breakdown of polymeric substances generates e⁻, which reduces with the help of enzymes hydrogenase (present in bacteria) and form bioH₂. These processes required additional pre-treatment steps. For example, Liu et al. [144] produced the fermentative bioH₂ using the pretreated microalgal biomass with 1.5% HCL and uses *C. butyricum* CGS5 for fermentation. Similarly, Chen et al. [145] pretreated the microalgal biomass with 1% H₂SO₄; during this experimentation, 2.87 mmol/g bioH₂ was produced via a dark fermentative pathway. Furthermore, Liu et al. [146] pretreated the microalgal biomass through the mixture of the biological enzyme (cellulase and protease). The result

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showed that 48% of bioH₂ was produced from the total microalgal biomass. However, the pretreatment steps are a costly process, whereas in photo fermentation, process light is required, which increases the cost of production [3]. Furthermore, integration of different microalgal bio-H₂ production might be a possible solution for obtaining the high bioH₂ yield at low cost.

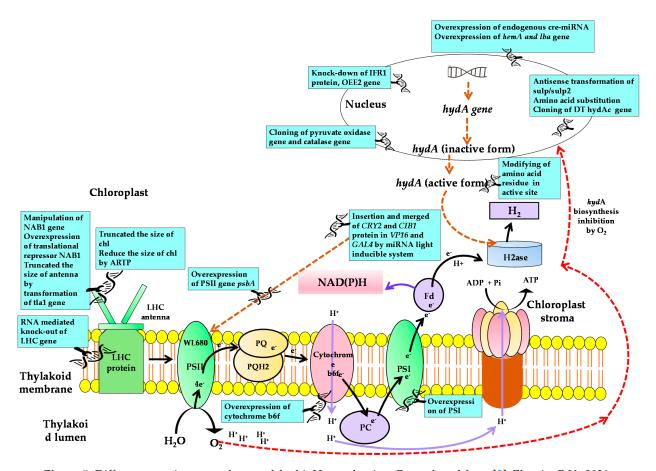


Figure 5. Different genetic approaches used for bioH₂ production. Reproduced from [3], Elsevier B.V.: 2021.

5.4. Bioelectricity

Microalgal based microbial fuel cell (MB-MFC) gains recent research attention due to its significant application in wastewater treatment and bioelectricity generation. Besides that, this technique also helps in carbon sequestration, bioH₂ production, and desalination [147]. MB-MFC is a biochemical system, which generates bioelectricity in the sequential biocatalytic reactions. MB-MFC systems are divided into two compartments: cathode and anode, and these are divided by ion exchange membrane separator. The catalytic oxidation of the substrates produces H⁺ and e⁻ at the anode side. The e⁻ undergo to the anodic surface via electrical mediators or direct e⁻ transfer through H⁺ membrane, which is existing on the surface of the microalgae. Then, the e is subsequently moved to the cathodic compartment via the outer circuit and it generated current flow in the direction of cathodic to anodic. At the same time, H⁺ also passes to the ion-exchange membrane to cathode chamber, where e⁻ reduces O₂ to H₂O. The O₂ is pumped out to the atmosphere from the cathode [148,149]. Jaiswal et al. [147] provided significant knowledge about MB-MFCs system. Different microalgae strains utilized for bioelectricity generation are presented in Table 7. Some recent investigation suggested that the integration of MB-MFCs system with microalgal based bioH₂ production are a cost-effective approach for wastewater treatment and bioH₂ production. In this system, wastewater is used as nutrient rich substrates for bacterial growth, the bacteria oxidize the substrates and generates H⁺ and e⁻, the e⁻ moves towards the anode, and is then transferred to the cathode where electron flow generates

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a bioelectric current. Then, H^+ moves towards cathode through the proton membrane exchanger and reacts with O_2 (which are produced during microalgal respiration) and forms H_2O . In algal cells, direct photolysis occurs, which produces $bioH_2$, furthermore, produced algal biomass during this treatment process can be utilized for fermentative $bioH_2$ production [147]. The overall process of MB-MFCs based $bioH_2$ is illustrated in Figure 6. Moreover, advance approaches are integrated for advancement of MB-MFC, while its large scale realization need to be demonstrated for real-world application and commercialization [150]. Although, single MB-MFC process is ineffective to generate the power of different commercial implementation; therefore, it can be integrated with AD technology.

Table 7. Different microalgae used for bioelectricity generation.

Microalgae	Cathode	Anode	Power Density	Reference
Chlorella vulfaris; Ulva lactuca	Graphite fiber brush	Air cathode with platinum (Pt) catalyst	980.0 mW/m ²	[151]
	Carbon fiber brushes	Carbon felt containing Pt catalyst	187.0 mW/m ²	[152]
	Toray carbon cloth	Toray carbon cloth	13.5 mW/m ²	[153]
	Graphite felt	Carbon fiber cloth	2572.8 mW/m ²	[154]
Chlorella vulgaris	Carbon felt	Carbon felt	24.4 mW/m ²	[155]
Cmorena vulgaris	Carbon felt	Carbon fiber cloth	2485.3 mW/m^3	[156]
	Carbon fiber brushes	Carbon cloth	5600.0 mW/m ³	[157]
	Graphite rod	Graphite rod	0.95 mW/m^2	[158]
	Carbon felt	Carbon fiber cloth	3720.0mW/m^3	[159]
Chlorella sp.	Graphite carbon	Graphite carbon	3.35 mW/m^2	[160]
Lagoon (algae culture)	Plain carbon cloths	Plain carbon cloths	11.5 mW/m^2	[161]
Laminaria saccharina	Graphite felt	Graphite felt	250.0Mw/m^2	[162]
Mixed algae culture	Carbon fiber brush	Carbon cloth coated with platinum	30.0 mW/m^2	[163]
Scenedesmus obliquus	Toraycarbon paper	Toray carbon paper	102.0 mW/m ²	[164]
Synechococcus leopoliensis	Black acrylic	Carbon fiber veil	42.5 mW/m^3	[165]

5.5. Biochar

Biochar is a carbon-rich charcoal made up by thermal decomposition (pyrolysis, hydrothermal liquefaction, and torrefaction) of different organic biomass under low oxygen and high temperature [166,167]. Biochar generally used as biofertilizer or absorbent for wastewater treatment, carbon sequester, etc. [167]. However, recent studies suggested that it can be used as a source of coal or coal fuel for the electricity generation [26,167]. Several studies reported that microalgal biomass have the potential for conversion into biochar [26]. Among the thermal conversion process, pyrolysis and torrefaction are promising technologies for biochar production from microalgal biomass [168]. Biochar produced by pyrolysis of microalgal biomass showed higher surface area, which has the properties of bio-adsorbent and can be used for removal of pollutant from liquid solution. The biochar produced by the pyrolysis process showed lower calorific value as compared to biochar produced from torrefaction process. Therefore, torrefaction process can be considered for biochar production from microalgal biomass. Due to high calorific value of biochar produced from torrefaction process, recently the research increased significantly to convert various types of biomass as solid coal fuel that can allow to partially replace the demand of

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fossil fuel [169]. Moreover, biochar derived from the torrefaction process offers a higherheating value (HHV) and a higher carbon number, improved grindability and aquaphobic in nature, lower moisture content, and ashes, as well as atomic ratio in comparison to unprocessed biomass/feedstock [170]. In the torrefaction process, moderate operational parameters allow to attain better productivity and yield of solid mass in comparison to traditional pyrolysis process at a high temperature [171]. Gan et al. [167] reviewed the biochar production from microalgal biomass through the torrefaction process. The review summarized that operational parameters such as temperature and residence time are the key parameters that influence the quality of biochar and solid yield. Biochar obtained from torrefaction of C. vulgaris ESP-31 biomass showed high higher heating value (HHV), while biochar obtained from Chalamydomonas sp. JSC4 showed lower atomic ratio and high carbon content. Biochar produced from microalgal biomass using pyrolysis process showed high bio-adsorbent property, which can used to absorb the pollutants from wastewater. Therefore, a microalgal biorefinery can be employed for sustainable development of circular bioeconomy. Moreover, Table 8 summarizes the fuel properties and comparison of different biochar derived from microalgal biomass and other biomass in dry and wet torrefaction.

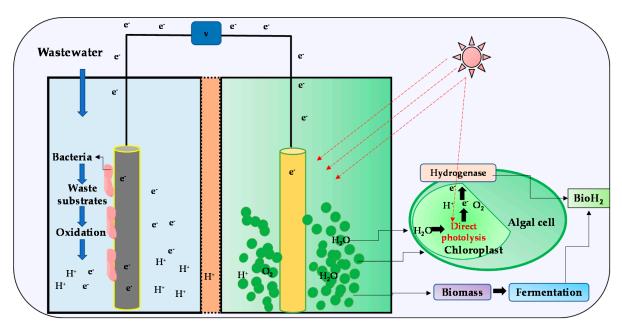


Figure 6. Microalgal based microbial fuel cells (MB-MFCs)-based wastewater treatment and bioH₂ production. Reproduced from [2,147], Elsevier B.V.: 2020.

Table 8. Fuel properties and comparison of different biochar derived from microalgal bio-mass and other biomass.

Type of Biomass		t & T	Ultimate Analysis (wt%)				HHV	T3 ((0 ()	
			С	Н	N	0	(MJ/kg)	EY (%)	Reference
Microalgae	Chalamydomonas sp. JSC4 residue	15–60 min; 200–300 °C	51.6–72.6	7.2–4.4	4.0-6.4	37.2–16.5	17.6–24.8	74.3–99.8	[172]
	Chlamydomonas sp. JSC4	30 min; 300 °C	63.6	5.01	6.0	25.4	-	_	[173]
	Chlorella vulgaris ESP.31	15–60 min; 200–300 °C	49.1–65.3	7.9–5.1	5.0–6.7	38.0–22.9	17.9–25.2	_	[174]
	Chlorella vulgaris ESP.31 by wet torrefaction	30 min; 170 °C	59.0	7.8	8.6	24.5	26.02	62.95	[175]
	Scenedesmus obliquus CNW-N	60 min; 200–300 °C	36.9–39.3	5.5–3.6	6.5–7.3	28.2–23.2	-	-	[176]

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Table 8. Cont.

	(P)	t & T Ultimate Analysis (wt%)					HHV		
Type of Biomass			С	Н	N	О	(MJ/kg)	EY (%)	Reference
Other biomass	Calophyllum inophyllum L	10 min; 260 °C	59.1	4.9	0.3	35.7	23.6	65.2	[177]
	Energy sorghum	30 min; 275 °C	55.2	4.9	1.7	38.1	23.80	73	[178]
	Humulud lupulud	10 min; 260 °C	60.5	6.0	2.7	30.8	25.3	38.5	[177]
	Jatropha-seed residue	30 min; 300 °C	61.1	5.2	4.2	20.7	27.01	-	[179]
	Waste bamboo chopsticks	40 min; 290 °C	55.5	5.4	0.2	38.3	23.04	-	[180]
	Landfill food waste	40 min; 275 °C	61.2	5.8	3.4	29.6	26.15	77.2	[181]
	Leucaena by microwave torrefaction	250 W	76.3	2.6	1.0	15.1	28.25	34.04	[182]
	Leucaena by microwave torrefaction	250 W	80.3	2.8	1.1	15.9	29.72	36	[183]
	Plumeria alba	10 min; 260 °C	60.7	6.8	0.6	31.9	25.7	45.7	[177]
	Sewage sludge	400 W	66.8	2.3	2.7	28.3	13.21	19	[183]
	Sunflower seed shell	60 min; 300 °C	69.5	5.3	0.5	24.6	27.6	-	[184]
	Sweet sorghum bagasse	30 min; 300 °C	59.3	4.6	0.9	35.2	26.88	70	[178]

Note: t & T: temperature and time duration; HHV: higher heating value; EY: energy yield.

6. Challenges of Microalgal Biorefineries and Future Perspectives

Microalgal biorefineries have gained tremendous attention for treatment of different types of wastewater generated from various industries and to produce various types of biofuel as well as other biotechnological, industrial, and environmental applications. However, there are still several challenges that need to be addressed, which include lower biomass productivity, harvesting of algal biomass, and energy consumption, as well as higher production cost for algal biofuel than the fossil-based fuels. To overcome these key challenges, flue gases and wastewater have been used for microalgae cultivation, which decrease the cost of nutrients and carbon source, but operational costs are high. However, the commercial production of microalgae biofuel remains a major constraint due to higher cost of microalgae cultivation and biomass harvesting. Therefore, algal biomass needs to be explored for potential application in different sectors mainly in cosmeceuticals and other sectors. Algal biomass cultivated using wastewater contains numerous toxic compounds including heavy metals, which demands efficient extraction technology for selective extraction of the desired compound with high purity from algal biomass. In this case, supercritical fluid extraction technology can be employed for selective extraction of desired compounds from microalgal biomass [185–187]. However, there are several knowledge gaps and problems associated with biomass production and lower yield, high expenses, and lack in commercialization of algal bioprocess. To enhance the biomass productivity, high inputs of nitrogen source can be used to enhance the biomass productivity, while modern genetic engineering tools such as CRISPR-Cas9, TALEN, and ZFN-17 can be applied to alter the genome and metabolic pathways of microalgae to enhance the biomass productivity for biofuel production as well as synthesis of various bioactive compounds for various commercial applications. For the reduction of energy consumption during microalgae-based bio-fuel production, further strategies need to be

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explored. Therefore, several steps need to be integrated to achieve a sustainable low-cost bio-fuel production process.

7. Conclusions

The present review focused on the microalgal biorefinery for resource recovery from wastewater in a sustainable manner. It also discussed the available technologies for treatment of various wastewater and highlighted that wastewater treatment from microalgaebased systems can provide a higher pollutant removal compared to most traditional systems, in a sustainable way. In addition, different microalgae cultivation systems for costeffective removal of various types of pollutant/nutrient from wastewater were examined. Among the various types of existing microalgae cultivation systems, ATS and HCS are highly recommended for fast and efficient removal of pollutants form wastewater. They also allow easier, faster, and cheaper biomass harvesting. Furthermore, microalgal biomass can be exploited for various types of biofuel production. Biodiesel produced from various microalgal strains, among them S. marginatum, derived biodiesel with the highest calorific value of 42.052 MJ/kg. However, in a biofuel scenario, microalgal biomass can be used for bioCH₄ production during co-digestion and microalgal biorefineries can be integrated for biogas upgrading, while the effluent from after AD process can be used as cultivation of microalgae. Microalgae based biogas upgrading showed up to ~98% CO₂ removal efficiency. In conclusion, development of integrated sustainable microalgal bio-refinery with multiple product recovery will provide an environmentally and economically sustainable solution to wastewater treatment and biofuel production.

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